

# Mercury exposure Assessment in fish and humans from Sundarban Mangrove Wetland of India

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Present study had documented total mercury levels in six commonly consumed fish species, and performed a cross-sectional study on local residents to gauge their intake of fish (via dietary survey) and mercury exposure (via hair biomarker analyses). Mean total mercury content in edible composites of locally-caught fishes (topse, hilsa, mackerel, topsa, sardinella, khoira) was low and ranged from 0.01 to 0.11  $\mu\text{g g}^{-1}$  mercury, dry weight. In a cross-sectional study of 58 area residents, the mercury content in hair ranged from 0.25 to 1.23  $\mu\text{g g}^{-1}$ , with a mean of  $0.65 \pm 0.23 \mu\text{g g}^{-1}$ . Hair mercury level was not influenced by gender, age, or occupation. Mean number of meals consumed per week was  $3.1 \pm 1.1$ , and all participants consumed at least one fish meal per week. When related to fish consumption, a significant positive association was found between number of fish meals consumed per week and hair mercury levels.

**[Keywords:** Exposure assessment, Methylmercury, Fish consumption, Biomarkers, Sundarban]

## Introduction

Mercury is a heavy metal of global public health concern<sup>1</sup>. Approximately 6,000 tons of mercury is released into the environment annually and concentrations continue to rise in many regions of the world<sup>2</sup>. A majority of this mercury is released from coal-fired power plants, and largely from point sources in India and China<sup>3</sup>. Epidemiological studies in China have started to document elevated exposures to mercury<sup>4</sup>, though in India much less is known about human exposures and associated health risks.

The public is primarily exposed to mercury (as methylmercury) through fish consumption<sup>1</sup>. Though released from most industries as an inorganic compound, upon deposition into aquatic ecosystems microorganisms can methylate this form of mercury

into methylmercury. As a methylated chemical, mercury can effectively cross biological membranes and accumulate in organisms, biomagnify through aquatic food chains, and build up in the tissues of fish-consumers<sup>5</sup>. The concentrations of mercury in tissues of fish-consumers may be 10-million times greater than ambient levels in the environment<sup>6</sup>. In India, there exist several studies showing the presence of mercury in fish<sup>7,8,9</sup>, and in many cases the measured values exceed consumption guidelines set by the U.S. EPA<sup>10</sup> ( $0.3 \mu\text{g g}^{-1}$ ) or WHO<sup>11</sup> ( $1.0 \mu\text{g g}^{-1}$ ).

Despite the ubiquity of mercury in Indian fish, little is known about human exposures via fish consumption. More than half India's population is estimated to eat fish and seafood on a regular basis and over 30% of its population relies upon it from

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livelihood<sup>7</sup>). Fish and seafood are a major source of dietary proteins, essential elements, and omega-3 fatty acids<sup>12</sup>. In addition, harvesting these items is of immense recreational, economical, and cultural importance to several groups. Clearly, mercury contamination of fish may have a range of deleterious societal impacts in India.

Within India, the Sundarban region of the state of West Bengal is an area worth studying in terms of mercury exposure<sup>13</sup>. The Sundarban wetland is a vast mega-delta in West Bengal, India, comprising over 100 islands and approximately 6.5 million people. Sagar Island is the most significant fishing center of Sundarban, particularly during the winter months. The potential source of mercury in this region are industrial sources (Paper factories, electronic industries, etc), agricultural run-offs (mercury-containing fungicides) and sewage sludge from the upper stretch of Ganga river<sup>14</sup>. With rapid development of electronic industries in West Bengal, a large number of outdated electronic products in the form of 'e-waste' contribute to the mercury sources in the study area<sup>13</sup>. The estuary receives raw sewage from the megacity of Calcutta located 85 km upstream. Nearby and bordering states (Bihar, Orissa, West Bengal) contain more than 50% of India's coal resources<sup>3</sup>. Studies from the Sundarban coastal regions have documented mercury in sediment<sup>13</sup>. In addition to potential health impacts, the presence of mercury may degrade fish and seafood which are critical to the sustenance, livelihood, and economy of local residents. The goal of this project was to increase understanding of mercury exposure in the region (and by extension, India given the dearth of information on this matter) by addressing the following objectives: A) to document total mercury levels in six commonly consumed fish species; and B) to perform a cross-sectional study on local residents to gauge their intake of fish (via dietary survey) and mercury exposure (via hair biomarker analyses).

## Materials and Methods

The study was performed within the Sundarban coastal region of the state of West Bengal, India. The sampling locations occur at the southernmost front

of Sagar Island (Figure 1), which is the largest island formed at the mouth of the Hugli estuary. The island



Fig. 1—Map showing location of the sampling sites in the Sundarban region of West Bengal, India.  $S_1$ - $S_6$  denote fish sampling sites and  $S_3$ - $S_4$  denote locations of the cross-section survey.

is approximately 300 km<sup>2</sup> in area with a population of over 160,000 individuals. Six species common to the region were sampled: pomphret (*Pampus pampus*), hilsa (*Tenualosa ilisha*), mackerel (*Rastrelliger kanagartha*), sardin (*Sardinella* sp), topse (*Polynemus paradises*) and khoira (*Setipinna phasa*). After returning to the laboratory the weights and lengths of the fish species were noted. The fishes were cut open and the edible muscles were sliced and washed with Milli-Q water. The samples were then dried at 40°C until dryness (3-4 days) and pulverized with a mortar and pestle. A composite of 5-6 individuals from each species was analyzed for total mercury.

Hair is used to gauge exposure to methylmercury<sup>1</sup>). Hair samples of 58 area residents were collected using methods outlined previously<sup>15</sup> after obtaining consent. Approximately 20-30 strands of hair were cut from the occipital region of the scalp. In addition to sampling hair, a brief survey was administered to gather information on gender, age, occupation, and fish consumption habits.

Concentration of total mercury in each fish and hair sample was measured in a Direct Mercury Analyzer 80 (DMA-80, Milestone Inc, CT) according to U.S. EPA Method 7473 as previously described by

us<sup>16,17</sup>. About 10–30mg of dried sample was weighted in a nickel sampling boat and placed into the DMA-80. Following decomposition of sample at 800°C, liberated mercury was next trapped using gold and then subsequently desorbed, carried to an absorbance cell and quantified spectrophotometrically. Analytical accuracy and precision were determined through the use of certified Standard Reference Materials (SRM) and intermittent analysis of duplicate samples. Recoveries of mercury in the DOLT-4 SRM (National Research Council of Canada) and the Japanese NIES hair SRM#13 were 105.3% (range: 104.7 – 106.0%) and 80.2% (range: 77.1 – 83.2%), respectively. Variability (measured by %RSD) of the two SRMs was < 6.0%. None of the results were adjusted based upon the reported SRM recoveries. The detection limit (0.037 ng mercury) was calculated as 3 times the standard deviation of the mean blank value, and none of the samples fell below this value. All mercury values are reported on a dry weight basis. The research protocol was approved by the Ethics Committee of India<sup>18</sup>. All the subjects were taken prior consent for analysis of the scalp hair for mercury with the help of short demonstration of the procedure before them.

#### Data Analysis

All statistical operations were performed using SPSS (v11.5, Chicago IL). Preliminary data analysis included tabulation of descriptive statistics for all measurements. The primary relationships of interest were associations between hair mercury levels and fish consumption, gender, age and occupation, and

were evaluated using parametric statistical procedures. All data are indicated as mean  $\pm$  standard deviation.

#### Results and Discussion

The length and weight of the sampled fish and their feeding habits are summarized in Table 1. The mean total mercury content in the fish composites ranged from 0.01 to 0.11  $\mu\text{g g}^{-1}$  (Table 1). The largest mercury concentration was found in topse followed by hilsa, mackerel, sardinella, and khaira. Topse (*Polynemus paradiseus*) being the predator fish mainly feeding on crustaceans (mainly shrimps), small fishes and benthic organisms whereas khaira (*Setipinna phasa*) mainly feeds on mysids and copepods. None of the total mercury values in these fishes exceed fish consumption guideline values used by the USEPA<sup>10</sup> (0.3  $\mu\text{g g}^{-1}$ ) or WHO<sup>11</sup> (1.0  $\mu\text{g g}^{-1}$ )

The concentrations reported in this study are generally lower than a recent NGO study focused on mercury in fish from several markets in Calcutta<sup>7</sup>. For example, values for total mercury in topse (0.41  $\mu\text{g g}^{-1}$ ), khoira (0.21  $\mu\text{g g}^{-1}$ ) and hilsa (0.69  $\mu\text{g g}^{-1}$ ) in the report by Chacraverti and Kumar<sup>7</sup> were about 10-times greater than what we report here. In addition, the values we report in this study (generally less than 0.1  $\mu\text{g g}^{-1}$ ) are also lower than values measured in a range of other fish species sampled from the Ganges river in West Bengal<sup>9</sup> and from the East Calcutta Wetlands<sup>8</sup>. This comparatively low values of mercury may be related with different sample sizes, ages and

Table 1— List of fish sampled from the Sundarban Wetlands of West Bengal, India. Length and weight data are expressed as mean  $\pm$  standard deviation, and the total mercury is a composite of 5–6 individual fish and reported on a dry weight basis.

Common Name	Species Name	Length (mm)	Weight (gm)	Total Mercury ( $\mu\text{g/g}$ )
Topse	<i>Polynemus paradiseus</i>	17.8 $\pm$ 1.8	19.6 $\pm$ 10.9	0.033
Khoira	<i>Setipinna phasa</i>	16.1 $\pm$ 1.3	55.5 $\pm$ 18.9	0.010
Pomphret	<i>Pampus pampus</i>	22.7 $\pm$ 1.6	145.2 $\pm$ 46.6	0.105
Mackerel	<i>Rastrelliger kanagurta</i>	17.9 $\pm$ 1.3	60.9 $\pm$ 13.4	0.048
Hilsa	<i>Tenualosa ilisha</i>	25.0 $\pm$ 0.2	166.8 $\pm$ 10.1	0.058
Sardine	<i>Sardinella sp</i>	28.5 $\pm$ 1.3	165.8 $\pm$ 22.7	0.027

characteristics of the captured environment<sup>19,20</sup>. In contrast similar low values of mercury (0.002 – 0.198  $\mu\text{g g}^{-1}$ ) in fishes were obtained in commonly consumed fish species from Taiwan<sup>21</sup> According to Wang<sup>22</sup> the assimilation of metals mainly depend the food conditions such as the food density and food type<sup>23,24,25</sup>. These external conditions may significantly affect the ingestion, digestion, solubilization<sup>26,27</sup>, membrane transport<sup>28</sup>, and gut passage time<sup>29</sup> and subsequently affect the dietary Assimilation Efficiency.

The present work represents a case study with limited sample size from Sundarban wetland. In order to evaluate the fish advisory level further studies are required with adequate number of collected fish species from this region. For the epidemiological portion of this study, 58 participants were recruited and equally distributed between genders. The mean age was  $27.3 \pm 14.0$ , and ranged from 4 to 70 years. In terms of occupation, 31% were involved in fishing, 24.1% were students, 32.8% were housewives, 6.9% were children, and the rest were involved in other activities. Nearly all (84.5%) participants were

literate. None of the women (n=29) smoke or drank alcohol, whereas 23/29 men smoked and 8/29 men drank.

The mercury content in hair ranged from 0.25 to 1.23  $\mu\text{g g}^{-1}$ , with a mean of  $0.65 \pm 0.23 \mu\text{g g}^{-1}$  (Table 2). There were no statistically significant differences in hair mercury values when stratified according to gender, age, or occupation (Table 2). Despite concerns of mercury pollution in India there exist few mercury human biomonitoring studies from the country for us to compare our work to. In a study of individuals associated with the Bhabha Atomic Research Center (BARC) in Bombay, mean values in blood (5.2  $\mu\text{g/L}$ ), urine (6.2  $\mu\text{g/L}$ ) and hair (1.2  $\mu\text{g g}^{-1}$ ) were reported though sample size and quality control values were not reported<sup>30</sup>. As part of an international study comparing trace element exposures across five countries, in a sample of 255 from Bombay and New Delhi the mean hair mercury value was 1.3 ppm<sup>31</sup>. In a study of 354 residents of Agra, the mean hair mercury values in males (0.73  $\mu\text{g g}^{-1}$ , range: 0-21) and females (0.77  $\mu\text{g g}^{-1}$ , range: 0-19.5) were similar<sup>32</sup>.

Table 2–Hair total mercury values ( $\mu\text{g/g}$ ) in residents from the Sundarban Wetlands of West Bengal, India. Data are stratified according to gender, age, and occupation with ANOVA p-values reported in column 1.

		N	Mean ( $\pm$ SD)	Median	Range
	All participants	58	0.65 (0.23)	0.60	0.25-1.23
Gender (p=0.92)	Male	29	0.66 (0.26)	0.60	0.25-1.23
	Female	29	0.65 (0.20)	0.59	0.35-1.09
Age (p = 0.15)	<18 years	20	0.73 (0.22)	0.71	0.36-1.12
	19-35 years	18	0.62 (0.26)	0.59	0.25-1.23
	>35 years	20	0.60 (0.19)	0.60	0.31-1.04
Occupation (p=0.35)	Fisherman	18	0.64 (0.28)	0.61	0.26-1.23
	Homemaker	19	0.60 (0.15)	0.58	0.35-0.87
	Student	14	0.67 (0.21)	0.60	0.36-1.09
	Child	4	0.86 (0.18)	0.94	0.59-0.96
	Other	3	0.72 (0.44)	0.78	0.25-1.12

In a study of autistic children from Chennai<sup>33</sup>, the mean hair mercury value in controls ( $0.37 \mu\text{g g}^{-1}$ ,  $n=50$ ) was significantly lower than values measured in cases ( $0.65 - 3.1 \mu\text{g g}^{-1}$ ,  $n=15$  per group, 3 groups according to Childhood Autism Rating Scale values). The mercury biomarker values reported in these aforementioned studies are similar to our findings. In addition, the mean hair mercury values are similar to values reported from other countries such as Pakistan<sup>34</sup> Iran<sup>35</sup>, USA<sup>36</sup>, and Germany<sup>37</sup>. The mercury levels are also lower than values observed among populations who depend on fish as a principal component of their diet, such as mothers from the Faroe Islands (median:  $4.5 \mu\text{g g}^{-1}$ ) and Seychelle Islands ( $5.8 \mu\text{g g}^{-1}$ )<sup>1</sup>.

Moving beyond biomarker measures, our study also asked about fish consumption and reported a significant association ( $p<0.001$ ) between fish consumption (queried as number of meals consumed per week) and hair mercury levels (Figure 2). Individuals with the highest mean hair mercury values also indicating to consuming 5 or more fish meals per week. To our knowledge, this is the first study in an Indian population to make such an association, and thus can be added to a large database of studies from across the world<sup>1</sup>. We do acknowledge key limitations of our survey methods (i.e., specific fish not identified, portion sizes lacking), and these should be addressed in the future to help increase understanding of what

specific fish (and possibly other food items) that may contribute to mercury burdens in India. For example, carefully linked fish consumption surveys and biomarker studies in the U.S. have enabled researchers to show that consumption of tuna contribute a majority of the mercury to the average citizen<sup>36</sup> (while in certain areas of China this mercury is largely derived from rice consumption<sup>38</sup>).

In terms of fish consumption, the mean number of meals consumed per week was 3.1 (1.1), ranged from 1 to  $>5$ , and was normally distributed. All participants consumed at least one fish meal per week. As indicated earlier, those that consumed five or more fish servings per week (1.1 ppm,  $n=9$ ) had significantly more hair mercury that those consuming one meal per week (0.3 ppm,  $n=2$ ). To perform a robust analysis, fish consumption was stratified into individuals consuming two or fewer meals per week, three meals, and four or more meals. In doing so, a significant difference was observed when these groupings were compared against mean hair mercury values ( $F=72.9$ ,  $p<0.001$ , Figure 2). Such an association has been documented in several other populations worldwide<sup>1</sup> and we believe our data is the first from an Indian population.

Fish consumption did not vary according to gender or occupation. Males consumed an average of  $3.1 \pm 1.0$  fish meals per week while females consumed an average of  $3.1 \pm 1.1$  fish meals per week. Fisherman ( $n=18$ ) consumed an average of  $3.0 \pm 1.1$  fish meals per week and this was not significantly different from the other occupational groupings. There was a significant age-related difference in consumption. The mean age of individuals that consumed more than 3 fish meals per week was  $18.1 \pm 11.1$  years and this was significantly greater than mean age of individuals that consumed 3 fish meals per week ( $29.0 \pm 15.5$  years) and those that ate less than 3 fish meals per week ( $34.7 \pm 8.9$  years). In the literature, there is conflicting information concerning age-related differences in mercury biomarker levels.

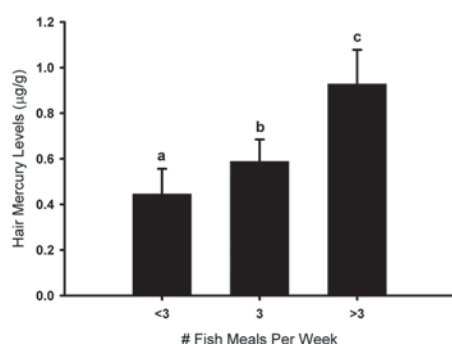


Fig. 2—Hair total mercury values ( $\mu\text{g/g}$ ) in residents from the Sundarban Wetlands of West Bengal (India) in relation to number of self-reported meals of fish consumed per week. Letters denote significant ( $p<0.001$ ) differences between the bars based on a one-way ANOVA.

## Conclusions

Present study provide information into the  $T_{Hg}$  concentration in fish samples from Sundarban and the probable ecotoxicological health hazards from fish consumption. Data pertaining to the THg levels in scalp hair samples of the residents of Sundarban demand further research as the concentration supersedes the WHO prescribed level.

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## References

- Mergler D, Anderson HA, Chan HM, Mahaffey KR, Murray M, Sakamoto M, Stern AH. 2007 Methylmercury exposure and health effects in humans: A worldwide concern. *Ambio*: 36, 3- 11.
- Swain EB, Jakus PM, Rice G, Lupi F, Maxson PA, Pacyna J M, Penn A, Spiegel S J, Veiga, M.M. 2007. Socioeconomic consequences of mercury use and pollution. *Ambio*: 36, 45-61.
- Mukherjee AB, Zevenhoven R. 2006. Mercury in coal ash and its fate on the Indian subcontinent: A synoptic review. *Sci Tot Environ*:368, 384-392.
- Zhang L, Wong MH. 2007 Environmental mercury contamination in China: Sources and impacts. *Environ Int*:33, 108-121.
- Clarkson TW, Magos L 2006. The toxicology of mercury and its chemical compounds. *Crit Rev Toxicol*. 36, 609-62.
- Evers D C, Wiener JG, Basu N, Bodaly RA, Morrison HA, Williams KA. 2011. Mercury in the Great Lakes region: bioaccumulation, spatiotemporal patterns, ecological risks, and policy. *Ecotoxicol*. 20, 1487-149.
- Chacraverti S, Kumar A. Fishing Toxics 2010– Mercury Contamination of Fish in West Bengal. Published by Toxics Link, Calcutta.
- Bhattacharyya S, Chaudhuri P, Dutta S, Santra SC. 2010. Assessment of total mercury level in fish collected from East Calcutta Wetlands and Titagarh sewage fed aquaculture in West Bengal, India. *Bull Environ Contam Toxicol*. 84: 618-22.
- Pal, M., Ghosh, S., Mukhopadhyay, M., Ghosh, M. 2012 Methyl mercury in fish-a case study on various samples collected from Ganges river at West Bengal. *Environ Monit Assess*. 184: 3407-14.
- USEPA. 2001. Ecological risk assessment bulletins-supplement to RAGS, region 4. Washington, DC. USEPA.
- FAO. 1991. Case Studies in forest-based small scale enterprises in Asia: Rattan, matchmaking and handicrafts, Rome.
- Chapman L, Chan HM. 2000. The influence of nutrition on methylmercury intoxication. *Environ Health Pers*. 108: 29-56.
- Chatterjee M, Canário J, Sarkar SK, Branco V, Godhantaraman N, Bhattacharya BD, Bhattacharya A. 2012. Biogeochemistry of mercury and methylmercury in sediment cores from Sundarban mangrove wetland, India-a UNESCO World Heritage Site. *Environ Monit Assess*. 184 (9), 5239-54.
- Kwokal •, Sarkar SK, Chatterjee M, Franciskovis-Billinski S, Billinski H, Bhattacharya A, Bhattacharya BD, Alam MA. 2008. An Assessment of Mercury Loading in Core Sediments of Sunderban Mangrove Wetland, India (A Priliminary Report). *Bull Environ Contam Toxicol* 81:105-112.
- Goodrich J, Wang Y, Gillespie B, Werner R, Franzblau A, Basu N. 2011. Glutathione enzyme and selenoprotein polymorphisms affect mercury biomarker levels in Michigan dental professionals. *Toxicol Appl Pharmacol*. 257, 301-308.
- Paruchuri Y, Siuniak A, Johnson N, Levin E, Mitchell K, Goodrich J, Renne E, Basu N. 2010. Occupational and environmental mercury exposure among small-scale gold miners in the Talensi-Nabdam District of Ghana's Upper East region. *Sci Tot Environ*.; 408, 6079-6085.
- Nam D-H., Basu N. 2011. Rapid methods to detect organic mercury and total selenium in biological samples. 5:3.
- Dent N J, Krishnan A. 2008. Ethics Committes in India. *The Quality Assur J*. 11: 2, 143-150.
- Renzoni A, Zino F, Franchi E. 1998. Mercury levels along the food chain and risk for exposed populations. *Environ Res* 77: 68-72.
- Storelli M M, Giacominielli-Stuffler R. Storelli A, D'Addabbo, R., Palermo, C. and Marcotrigiano, G. O. 2003. Survey of total mercury and methylmercury levels in edible fish from the Adriatic Sea. *Food Addit Contam* 20: 1114-1119.
- Chen YC, Chen MH. 2006. Mercury levels of Seafood Commonly Consumed in Taiwan. *J Food Drug Anal*. 14 (4):373-378.
- Wang WX. 2012. Biodynamic understanding of mercury accumulation in marine and freshwater fish. *Adv Environ Res* 1 (1): 15-35.

- 23 Ni IH, Wang WX, Tam YK. 2000. The transfer of Cd, Cr, and Zn from zooplankton to mudskipper and glassy fishes. *Mar Ecol Prog Ser* 194: 203-210.
- 24 Wang WX, Wong RSK. 2003. Bioaccumulation kinetics and exposure pathways of inorganic mercury and methylmercury in a marine fish, the sweetlips *Plectorhinchus gibbosus*. *Mar Ecol Prog Ser* 261: 257- 268.
- 25 Zhang L, Wang WX. 2006. Significance of subcellular metal distribution in prey in influencing the trophic transfer of metals in a marine fish. *Limnol Oceanogr* 51 (5): 2008-2017.
- 26 Leaner JJ, Mason RP. 2002. Methylmercury accumulation and fluxes across the intestine of channel catfish, *Ictalurus punctatus*. *Comp Biochem. Phys C* 132 (2),: 247-259.
- 27 Goto D, Wallace WG. 2009. Influences of prey- and predator-dependent processes on cadmium and methylmercury trophic transfer to mummichogs (*Fundulus heteroclitus*). *Can J Fish Aquat Sci* 66 (5): 836- 846.
- 28 Mason RP, Reinfelder JR, Morel FM. 1996. Uptake, toxicity, and trophic transfer of mercury in a coastal diatom. *Environ Sci Technol* 30 (6): 1835-1845.
- 29 Xu Y, Wang WX. 2002. Exposure and food chain transfer factor of Cd, Se, and Zn in a marine fish, *Lutjanus argentimaculatus*. *Mar Ecol Prog Ser* 238: 173-186.
- 30 Panday V.K, Parameswaran M, Soman SD.1986. The distribution of mercury in the Indian population. *Sci Tot Environ*. 48 (3), 223-30.
- 31 Takagi Y, Matsuda S, Imai S, Ohmori Y, Masuda T, Vinson JA, Mehra MC, Puri BK, Kaniewski A. 1986. Trace elements in human hair: an international comparison. *Bull Environ Contam Toxicol*. 36, 793-800.
- 32 Sharma R, Chandreshwor Singh L, Tanveer S, Verghese P S, Kumar A. 2004.Trace element contents in human head hair of residents from Agra City, India. *Bull Environ Contam Toxicol*. 72: 530-534.
- 33 Lakshmi MD, Geetha A. 2011. Level of trace elements (copper, zinc, magnesium and selenium) and toxic elements (lead and mercury) in the hair and nail of children with autism. *Biol Trace Element Res*. 142: 148-58.
- 34 Anwar M, Ando T, Maaz A, Ghani S, Munir M, Qureshi IU, Naeem S, Tsuji, M, Wakamiya J, Nakano A, Akiba S. 2007. Scalp hair mercury concentrations in Pakistan. *Environ Sci*. 14: 167-75.
- 35 Fakour H, Esmaili-Sari A, Zayeri F. 2010. Scalp hair and saliva as biomarkers in determination of mercury levels in Iranian women: Amalgam as a determinant of exposure. *J Haz Mater*. 177: 109-113.
- 36 Pesch A, Wilhelm M, Rostek U, Schmitz N, Weishoff-Houben M, Ranft U, Idel H. 2002. Mercury concentrations in urine, scalp hair, and saliva in children from Germany. *J Expo Anal Environ Epidemiol*; 12, 252-8.
- 37 Wang Y, Goodrich JM, Gillespie B, Werner R, Basu N, Franzblau A. 2012. An investigation of modifying effects of metallothionein single-nucleotide polymorphisms on the association between mercury exposure and biomarker levels. *Environ Health Pers*. 120: 530-4.
- 38 Zhang H, Feng X, Larssen T, Qiu G, Vogt RD. 2010. In inland China, rice, rather than fish, is the major pathway for methylmercury exposure. *Environ Health Pers*. 118, 1183-8.