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2 **Multilevel environmental assessment of the anaerobic treatment of dairy processing**
3 **effluents in the context of circular economy**

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11 **Abstract**

12 The capability of Anaerobic Digestion (AD) in minimising waste and retaining the value of materials and
13 energy within the biological and technical cycles in the Dairy Industry (DI) makes AD a critical
14 instrument of transition to circular economy. The aim of this paper is to propose a framework and an
15 approach for measuring the environmental performance of the anaerobic treatment of dairy processing
16 effluents based on circular economy principles. The potential of AD to close the water, energy and
17 nutrient circular loops is investigated together with the relevant environmental costs and benefits at
18 different levels of the dairy supply chain. The developed methodology was based on Material Flow
19 Analysis (MFA) and Life Cycle Assessment (LCA) applied at three different system levels: the anaerobic
20 digestion plant, the dairy processing facility, the entire dairy supply chain. The approach is demonstrated
21 in a dairy facility coupled with a full-scale AD unit. The results show that the excess electricity (426
22 MWh/annum) and heat (1236 MWh/annum) produced from the anaerobic digestion plant cause
23 significant reduction of the overall environmental impact of the processing facility. The recovered energy
24 from anaerobic digestion provides 20% of the energy requirements of the factory reducing the total
25 carbon footprint emissions by 13% compared to the baseline scenario.

27 **Keywords:** Circular economy indicators, wastewater, anaerobic digestion, dairy processing, LCA

1. Introduction

Dairy products form an essential source of daily nutrients in human diets (Weaver, 2010), whereas the dairy industry, in 2012, accounted for 13.6% of the food and drink industry turnover (Wijnands and Verhoog, 2016). Several studies discuss the need to reduce the amount of dairy products in European diet patterns (Freibauer et al., 2011), which is estimated to be beneficial both to the environment (Godfray et al., 2010) and the human health (Hawkesworth et al., 2010). However, adaptation and mitigation strategies for reduction of greenhouse gas emissions (GHG) and enhancement of environmental resilience remain main challenge for the dairy industries (Prasad et al., 2004). The production of dairy products is combination of processes, including agriculture, livestock farming, manufacturing, packaging, distribution, retail and consumption (Kirilova and Vacklieva-Bancheva, 2017). Therefore, the development of sustainable dairy value chains should take into account the reduction of the environmental impacts and cause-effect relationships within all stages of the supply chain. Dairy farms have been the focal point of environmental assessments in the dairy sector. The application of life cycle impact assessment (LCIA) has been used as a tool to facilitate the decision making in the dairy sector and increase its environmental performance. Several works have assessed the environmental impacts of the dairy sector proposing measures for the improvement of the sustainability of the dairy value chains (Battini et al., 2014; Hospido et al., 2003; Roy et al., 2009). Recovery of bioenergy (Kimming et al., 2015) and use of other renewable energy sources (Murgia et al., 2013), recycling of nutrients (Dolman et al., 2014) and wastewater treatment and valorisation (Gottschall et al., 2007), have been identified as key factors for the enhancement of the environmental profile of dairy farms. Recently, Kılış and Kılış (2017), developed a methodological approach for the comparison of different energy and biogas utilization schemes in a dairy farm following circular economy principles. New industrial symbiosis paradigms in Europe have demonstrated efficient management of materials, energy, water and waste flows mainly in industrial applications (WssTP, 2016), however, applications in the dairy supply chain are still premature. Monitoring of key performance indicators (KPIs) integrating environmental impact and related accountability allocation have been considered as main components for the development of an

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55 enhanced sustainability framework (Huysman et al., 2015) and as a basis for the development of
56 circularity metrics (Linder et al., 2017).

57 Circular economy (CE) is gaining increased attention over the conventional “make-use-dispose” models
58 (Ghisellini et al., 2016; Jawahir and Bradley, 2016). Fundamental principles of circular economy
59 strategies focus on the reduction, re-use and efficiency of resources utilization (Wu et al., 2014), while
60 boosting economic growth (Ellen MacArthur Foundation and the McKinsey Center, 2015) and therefore
61 directly linked with sustainable waste and resource management (Blomsma and Brennan, 2017), systems
62 thinking, re-design and “closing loops” of materials and energy flows (MacArthur, 2012). The dairy
63 sector traditionally features circular practices and there are many examples that demonstrate the potential
64 of circular dairy farming (Kılıkış and Kılıkış, 2017; Lybæk and Kjær, 2019). However, significant
65 challenges remain to achieve a truly circular dairy sector that is regenerative and closes nutrient, water,
66 carbon and waste cycles. The driving force for waste minimization in the dairy industry is the improved
67 yields of product, reduced impact on the environment and lower wastewater treatment costs (Barnett et
68 al., 2010). Waste-to-energy systems are seen as a mean to facilitate the transition to circular economy
69 (Pan et al., 2015) and are key solutions for the mitigation of the environmental impacts in the dairy
70 processing sector. Nowadays, with the evolving recycling technologies and solutions that are available on
71 the market, the nutrient conversion of dairy manure and milk processing residuals is becoming more
72 efficient and economically viable. As a result, the nutrient recycling is getting momentum due to its
73 environmental and economic benefits (Dolman et al., 2014). The Dairy sector is included in the priority
74 list of the recent political agenda of the European Union where the circular economy is an increasing area
75 of focus for the European businesses (EC, 2015). The anaerobic digestion (AD) has a major role in the
76 transition to circular economy due to its capability to minimise waste and retain the value of materials and
77 energy within the biological and technical cycles. Using the liquid by-product of anaerobic digestion to
78 restore natural capital to soil is a step forward in finding a way to produce fertiliser from a waste resource,
79 keeping nutrients in a cycle (Passeggi et al., 2012). Dairy factories generate significant amount of
80 wastewater from the various processing steps (i.e. reverse osmosis for milk concentration) and during
81 cleaning, heating, cooling or floor washing (Demirel et al., 2005). Dairy effluents constitute a good

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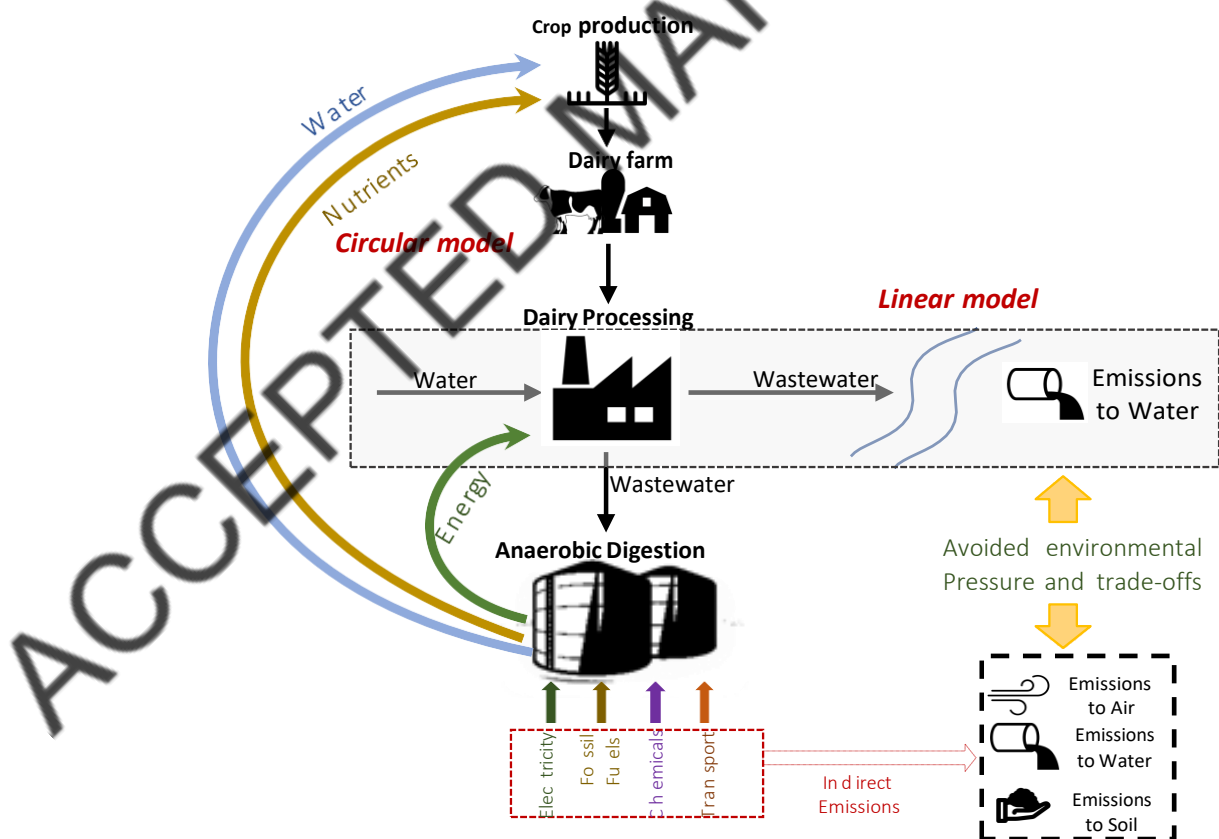
82 feedstock for anaerobic digestion processes, since they are characterised by significant organic and
83 microbiological load (Carvalho et al., 2013; Karadag et al., 2015; Prazeres et al., 2012). The techno-
84 economic viability of dairy effluents treatment by applying anaerobic digestion has been assessed in
85 various works (Carlini M et al., 2015; Traversi et al., 2013; Demirel et al., 2006; Gelegenis et al., 2007;
86 Zhong et al., 2015). However, these are usually are standalone studies, partially addressing individual
87 aspects of circular economy. At the same time optimization of the operating conditions remains the main
88 constraint for the widespread AD implementation in the dairy industry (An et al., 2010; Prazeres et al.,
89 2012) especially without the use of other feedstock.

90 Most of the LCA studies in dairy sector have been focusing on the dairy farm stage (Palhares and
91 Pezzopane, 2015; Zonderland-Thomassen and Ledgard, 2012) and only limited number of works have
92 considered the processing stage (Vasilaki et al., 2016). There is still a gap in the literature on sound LCA
93 based environmental impact assessment for dairy processing wastewater treatment. To the best of our
94 knowledge, the environmental and sustainability performance of a full scale AD treatment of dairy
95 effluents management strategy has not been systematically evaluated. The main objective of this study is
96 to evaluate the environmental performance of the AD dairy effluent treatment and to reveal its potential to
97 close the water, energy and nutrient circular economy loops at different levels of the dairy supply chain.
98 More specifically, the LCA was applied to assess the environmental performance of the AD, as well as its
99 benefit and costs ratios at each level. The approach was applied to a dairy processing facility coupled with
100 full-scale high-rate liquid AD unit treating the dairy processing effluents, located in South West of the
101 UK. The novelty of this study is to translate the LCA and MFA results into suitable circular economy
102 metrics for measuring the effectiveness of AD wastewater treatment on circularity performance, both in
103 terms of efficiency and scale. The findings of such analyses will facilitate decision makers and managers
104 towards improving sustainability of dairy industry.

105 **2. Materials and Methods**

106 CE is often associated with eliminating waste and closing the technological and biological material loops.
107 Figure 1 illustrates the ability of AD treatment to increase the circularity potential by reducing the direct
108 environmental pressure to receiving waters and valorise the embedded resources in the dairy effluents by

109 closing the water, energy and nutrient loops at different stages of the dairy supply chain. A linear model
 110 of the water usage in dairy factory would follow the take, make dispose concept, where water is taken
 111 directly or indirectly (by a public drinking water system) from the nature and after its use the generated
 112 wastewater is discharged untreated into receiving water bodies. The latter results in release of direct
 113 emissions to water bodies, which can cause eutrophication and have other adverse effects on the
 114 ecosystem. Nowadays, water use industries should meet certain regulation and legislation standards for
 115 discharging their effluents either directly into the environment or into a public sewer. In this sense, the
 116 pre-treatment/treatment of the effluents of a dairy factory can be considered already as a step towards
 117 circularity depending on the water quality level and further appropriate use of the treated effluents.
 118 Therefore, the scenario in which the wastewater is discharged directly into the environment (linear model)
 119 is more suitable to be used as Baseline (reference point) to which all circularity scenarios are relatively
 120 assessed. (Figure 1).



121
 122 **Figure 1 Circular economy loops in the dairy supply chain**

123 The anaerobic digestion process enables the recovery of embodied resources from the dairy wastewater
124 that can be recycled in a closed loop fashion at various stages of the supply chain. However, focusing
125 ultimately on closing the water, material and energy loops can result in bigger externalities and even lead
126 to negative net environmental performance. The LCA has been recognised as a valuable tool to capture
127 these trade-offs and justify the overall net environmental impact of a system change. Therefore, the
128 proposed methodology follows the key elements of the LCA, which generally consists of four phases:
129 goal and scope definition, life cycle inventory (LCI), life cycle impact assessment and interpretation of
130 results (ISO 14040, 2006). The results of the LCA identify the processes that have the highest impact in
131 the product lifecycle and facilitate the selection of suitable measures for the overall environmental impact
132 reduction. Although, the conventional LCA includes credits for recovered materials or substituted
133 resource inputs, it does not provide suitable and meaningful interpretation of the results from circular
134 economy perspective. Therefore, in the proposed methodology the LCA has been aligned to the circular
135 economy approach by including additional means of interpretation and two circularity indicators
136 expressing the material and environmental circularity performance.

137 **2.1. System description**

138 The studied dairy processing facility is located in the South West of United Kingdom. The raw milk is
139 sourced from local community dairy farms within 25 miles distance from the facility. The dairy farms
140 vary in size from small family farms with 80 cows to larger farms with up to 400 cows. The cows are fed
141 with grass during the summer months, while in the winter they are fed mainly with silage and cereals. The
142 dairy processing company processes about 42 million litres of milk annually and produces various fresh
143 and cultured dairy products for food manufacturers and service operators in the UK. It generates about
144 80,000 m³ of wastewater annually from various processing stages, such as milk receiving/storage,
145 pasteurisation, homogenisation, separation/clarification, cheese/butter/milk making, packaging and during
146 cleaning, heating/cooling or floor washing. The dairy plant wastewater contains milk components, and
147 acid and alkaline detergents used in the equipment cleaning. There are two wastewater streams leaving the
148 dairy processing facility – 1) trade effluent from spillages and wash-water rinses; and 2) wastewater
149 generated during soft cheese production (permeate of milk ultrafiltration). The permeate is characterised

150 by high chemical oxygen demand (COD) load ranging from 40.4 to 64.8 g/L, whereas the average COD
 151 concentration of the trade effluent is 15.0 g/L. The wastewater streams generated from the dairy facility
 152 are stored in two equalization tanks and the mixed flow is fed to the AD reactor resulting in 21.1 g/L
 153 COD and 0.4% Total Suspended Solids (TSS) in the feedstock. Thus, the AD operates with 3.29 kg
 154 COD/m³·d Organic Loading Rate (OLR) and average hydraulic retention time (HRT) of 6.90 days.

2.2. System boundaries

156 System boundaries are usually defined according to the goal of the study and include the relevant
 157 processes in the product system (ISO 14044, 2006). However, within the context of circular economy the
 158 boundaries are dependent also on the upstream and downstream processes. This is because the resource
 159 and energy recovery loops are crossing different levels in the supply chain in an open loop or closed loop
 160 fashion. Therefore, in order to better assess the AD circularity performance, three levels of system
 161 boundaries have been defined: **Level 1:** AD treatment plant; **Level 2:** Dairy processing factory; and **Level**
 162 **3:** The entire dairy supply chain – from the raw inputs in the farm stage to the dairy products distribution
 163 to the end customers (Figure 1). The analysis at Level 1 aim to evaluate the environmental efficiency of
 164 the AD unit, while at Level 2 and Level 3 reveal the scale of AD circularity improvements at dairy factory
 165 and supply chain level.

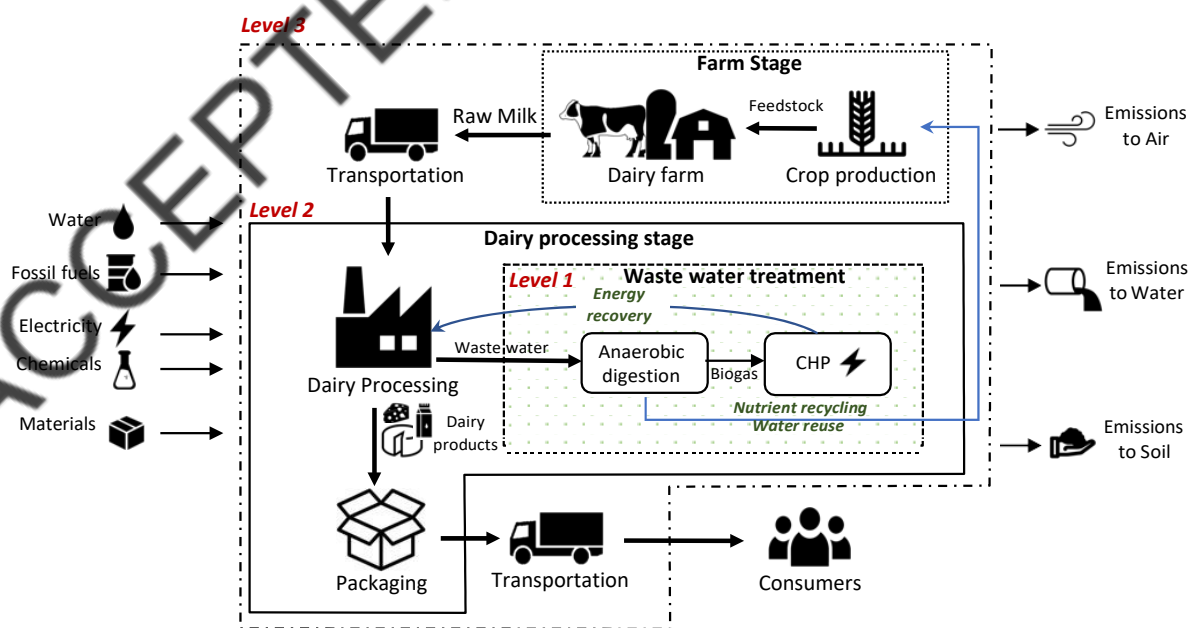


Figure 1 Multilevel system boundaries within the dairy supply chain

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169 **2.3. Definition of functional unit (FU)**

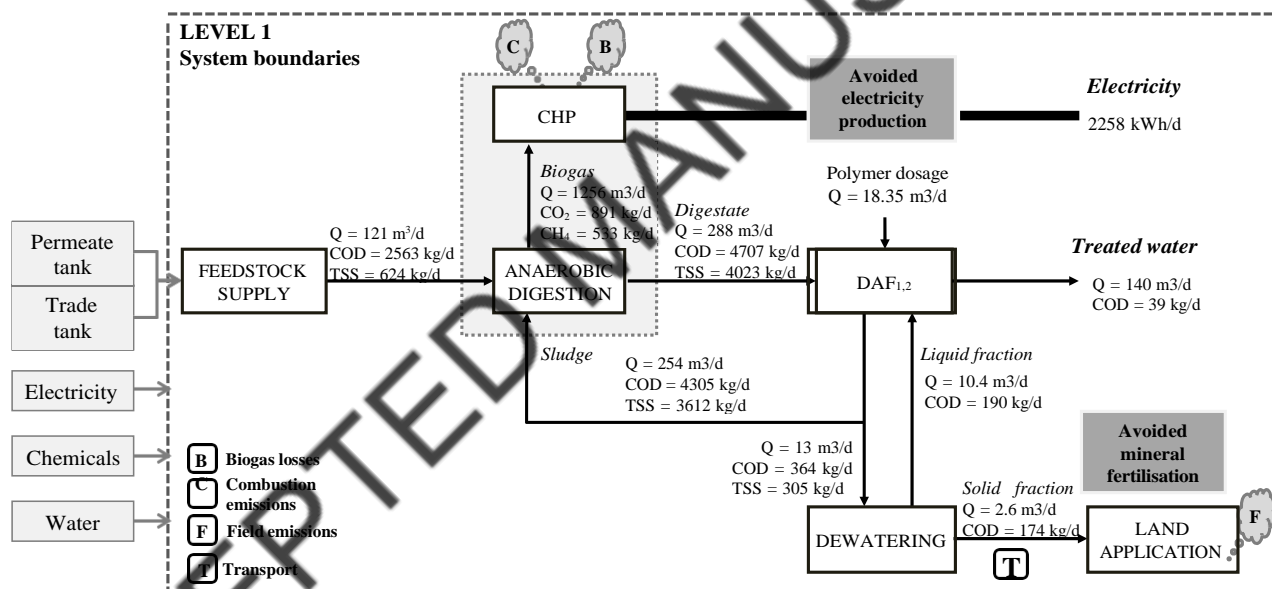
170 The main purpose of the functional unit is to quantify the performance characteristics for a target flow in
171 the system and to provide a reference to which all inputs and outputs of the system are normalised
172 (ISO14044 and ISO 14040, 2006). In Level 1, the m³ of wastewater on the input of the AD was selected
173 for a FU as a common choice in most of the LCAs applied on wastewater treatment processes (Berlin,
174 2002; Dolman et al., 2014; Salou et al., 2017). On a dairy factory or farm level, the most widely used
175 functional unit is based on input mass/volume of the raw milk or focuses on the nutritional / economic
176 aspects of the final dairy products (Cederberg and Mattsson, 2000; Flysjö et al., 2011). For the analysis at
177 level 2 and level 3, FU of 1 kg of fat and protein corrected milk (FCPM) was chosen as recommended by
178 the International Dairy Federation (IDF) guidelines (IDF, 2015). However, since the calculation of FCPM
179 using the standard fat (4%) and protein (3.3%) equals ~1 kg of raw milk, for simplicity, the “1 kg of
180 processed raw milk” was used.

181 **2.4. Life Cycle Inventory analysis**

182 Primary data were collected from the dairy processing company about the operations of the installed AD
183 plant, including: water, chemicals, energy consumption, energy generation, transport, and digestate
184 management. The type of data used for the LCA were based on experimental data and the measurement
185 records of main parameters (COD, TSS) (ii) complete mass balance of the process, and (iii) and data
186 extracted from the relevant literature about the parameters related to emissions characterization. The
187 construction of the AD plant has proven to have a minor contribution to total environmental impact thus
188 the construction stage of the plant has not been taken into account in the analysis (Mezzullo et al. 2013).
189 **Figure 2** shows the material and energy flows of the treatment of dairy effluents. The characteristics of
190 the treated effluent are shown in **Table 1**. The annual capacity of the mesophilic AD unit is 70,000 m³
191 whereas currently the average influent to the reactor is 121.3 m³/day. The operating parameters of the
192 system are summarised in **Table 2**.

193 The produced biogas from the AD plant is about $0.35 \text{ m}^3 \text{ CH}_4/\text{kg COD}_{\text{rem}}$ and consists of 64% CH_4 and
 194 36% CO_2 and is lead to a combined heat and power (CHP) engine where 2,258 kWh/day electricity on
 195 average is generated. The CHP unit has 105 kW electrical output and 32% electrical efficiency. The
 196 majority of the electricity generated (about 62%) is used for the operation of the AD plant, while the
 197 remaining electricity is used to cover the energy needs of the dairy facility or fed to the national grid. The
 198 AD effluent is characterised by 15.5 g/L COD and 1.3% TSS (average values).

199 The digestate is pumped out of the AD reactor into two Dissolved Air Flotation (DAF) units where it is
 200 thickened and polished, resulting into about $140 \text{ m}^3/\text{day}$ of treated effluent with 276 mg/L COD
 201 concentration. Approximately 92.5% of the thickened digestate is recirculated to the reactor and 7.5% is
 202 further thickened to a screw press (18% TSS).



204 **Figure 2.** Flowchart of the anaerobic digestion plant treating dairy processing wastewater

205 **Table 1.** Summary of input and output parameters in dairy AD unit.

Parameters	Units	Permeate	Trade	Effluent
Wastewater flow	m^3/d	24.0	97.0	140
COD	g/L	48.4	14.4	0.28
TSS	%	0.37	0.55	-

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208 **Table 2.** AD operating parameters.

Parameters	Units	Value
HRT	Days	6.7
OLR	Kg COD/m ³ day	3.9
SRT	Days	36
T	°C	30.7

209 The avoided impacts from the substitution of mineral fertilisers with the organic fertiliser, generated from
210 the AD were calculated based on the IPCC guidelines (IPCC, 2006). The electricity and heat produced
211 from biogas combustion in the CHP unit was assumed that substitutes an equivalent energy amount of
212 electricity and kerosene used in the dairy processing facility. The carbon emissions generated during the
213 valorisation process of the produced biogas from the AD have not been included in the analysis, since the
214 biogas is derived from organic waste streams. Thus, it does not add to the carbon dioxide load in the
215 atmosphere since the CO₂ emissions produced during combustion of biogas are offset by either the carbon
216 dioxide consumed by the biomass or by the avoided fugitive methane emissions.

217 In the expanded system boundaries (Level 2 and Level 3) the dairy processing stage has been considered
218 as a “black box” representing an aggregation of processes, including the wastewater treatment stage. The
219 total input and output energy and material flows for the operation of the dairy facility were collected from
220 the dairy company. The transport emissions for supplying raw milk, packaging materials and the
221 distribution of the dairy products were calculated based on the weight and the average distance to the
222 providers and retailers respectively. Secondary data obtained with SimaPro from Ecoinvent[®] databases
223 were used for the farm stage and all other upstream process to estimate the environmental impacts related
224 to the intermediate inputs from the technosphere.

225 The global inventory data is given in **Table 3**.

226 **Table 3.** Global annual inventory data of material and energy flows for the dairy supply chain

Inputs from technosphere	Amount	Unit	Data source
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1	Raw milk	42,000,000	kg/year	Primary data: Dairy factory
2	Total Water Usage	12,450	m ³ /year	
3	<u>Chemical Usage</u>			
4	Dairy processing			
5				
6	Disinfectant: mainly PAA	51,125	kg/year	
7	Detergent	4,200	kg/year	
8	Alkaline detergent: NaOH and KOH	152,010	kg/year	
9				
10	Acid: nitric and phosphoric acid	8,470	kg/year	
11	Enzyme: protease, lipase	11,210	kg/year	
12				Primary data: Dairy factory
13				
14	Wastewater treatment			
15				
16	Flocculant (polyvinylchloride)	27,000	kg/year	
17	Calcium carbonate	1,000	kg/year	
18	Iron (III) chloride, without water,	24,000	kg/year	
19	Sodium Hydroxide	36,000	kg/year	
20				
21				
22	<u>Energy use</u>			
23				
24	Dairy processing			
25	Fuel (Kerosene/light oil)	450,934	kg/year	
26				
27	Electricity	3,687,989	kwh/year	
28				
29	Wastewater treatment			Primary data: Dairy factory
30	Electricity	398,652	kwh/year	
31				
32	<u>Packaging materials</u>			
33				
34	Card Sleeve	18,942	kg/year	
35	Cardboard Divider	2,200	kg/year	
36	Cardboard Outer	15,330	kg/year	
37	Paper Label	36,546	kg/year	
38	Plastic Bucket	342,142	kg/year	
39	Plastic Carton	3,978	kg/year	
40	Plastic Film	12,860	kg/year	Primary data: Dairy factory
41				
42	Plastic HDPE(2) Bottle	28,188	kg/year	
43	Plastic HDPE(4) Lid	2,619	kg/year	
44				
45	Plastic label	643	kg/year	
46				
47	Plastic Lid	107,350	kg/year	
48	Plastic Liner	22,330	kg/year	
49	Plastic Pot	9,156	kg/year	
50				
51	<u>Transportation</u>			
52	Distribution of products	7,302,050	t-km	
53				
54	Average Distance to main distribution points	310	km	Primary data: Dairy factory
55				

	Total weight of generated products	23,555	t	
1	Chemical/ingredients inputs	81,725	t-km	
2				
3	Average distance from providers	360	km	
4	Packaging materials	251,755	t-km	
5				
6	Average distance from providers	418	km	
7				
8	Raw milk input	840,000	t-km	
9				
10	Average distance from providers	20	km	
11	Waste disposal	40,250	t-km	
12				
13	Average distance to landfill			
14		50	km	
15				
16				
17	Outputs to technosphere	Amount	Unit	Data source
18				
19	<u>Avoided energy production</u>			
20	AD Electricity Generation from CHP	824,039	kWh/year	Primary data: Dairy factory
21				
22	AD Heat Generation from CHP	1,236,000	kWh/year	Primary data: Dairy factory
23				
24	Avoided fertiliser production			
25	Generated sludge	805,000	kg/year	Primary data: Dairy factory
26				
27	N fertiliser substitution	397	kg/year	Vadenbo et al. 2017
28				
29	P fertiliser substitution	2287	kg/year	Vadenbo et al. 2017
30				
31	Land application emissions of the recovered fertilizer			
32	Direct N ₂ O	3.97	kg N-N ₂ O/year	
33	Indirect N ₂ O (atm. deposition)	0.79	kg N-N ₂ O/year	
34				IPCC 2006 guideline
35	Indirect N ₂ O (Leaching)	0.89	kg N-N ₂ O/year	
36				
37	Indirect NO ₃ (Leaching)	119.10	kg N-NO ₃ /year	
38	NH ₃ emissions	26.16	kg N-NH ₃ /year	
39				EMEP/EEA 2016 guideline
40	NO emissions	7.41	kg N-NO/year	
41	Phosphorus leached to ground water	0.07	kg P/ha*year	
42				
43	Phosphorus lost by runoff	2.13	kg P/ha*year	SALCA emission models
44				
45	Phosphorus emitted through erosion to rivers	0.0008	kg P/ha*year	
46				
47	Waste			
48	Wastewater	80,346	m ³ /year	
49				Primary data: Dairy factory
50	Packaging waste	58.4	t	
51				

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228 The Life Cycle Impact Assessment (LCIA) was conducted following the ReCiPe Midpoint (H) 1.12
229 method (Goedkoop et al., 2009). Based on the nature of the system eight environmental impact categories
230 were selected: climate change (CC), ozone depletion (OD), freshwater eutrophication (FE), ionising

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231 radiation (IE), agricultural land occupation (ALO), water depletion (WD), metal depletion (MD) and
232 fossil depletion (FD). The software SimaPro v8.0.5.13 was used for the computational implementation of
233 the inventories.

234 **2.5. Circularity performance assessment**

235 The potential of circularity and the benefits of the AD treatment in dairy sector have been highlighted in a
236 number of studies in the dairy sector (Demirel et al., 2005; Malaspina, F. et al., 1996; Strydom JP et al.,
237 1997). However, although there are several comprehensive sets of CE performance indicators on national
238 and regional level (EASAC, 2016), to date there is still no standardized and well-established method to
239 measure the circular economy performance on product level.. Quantitative indicators are essential to
240 evaluate how well an organization or product system performs in relation to the CE principles. One of the
241 main challenges in evaluating circularity is the selection of units that allow integration of the different
242 circularity aspects into a single value of circularity. In this paper, two dimensionless circularity metrics
243 based on MFA and LCA have been proposed to evaluate the material and environmental circularity
244 performance of the AD solution.

245 ***Material circularity performance indicator (MCPI)***

246 The material circularity performance metric is based on the Demand Minimisation Index (DMI) suggested
247 by Agudelo-Vera et al., 2012 (Equation 1) which enables to assess to what extent the baseline demand of
248 resource or energy flow is reduced at the level of the actual closing of the circularity loop. This results in
249 a value between 0 and 1, where 0 means that there is no reduction in the demand and 1 means that the
250 whole demand is covered.

(1)

$$251 \quad MCPI = \frac{\text{Baseline demand} - \text{Minimised demand}}{\text{Baseline demand}} \quad (2)$$

252 ***Environmental circularity performance indicator (ECPI)***

253 The main purpose of the LCA is to quantify both positive and negative environmental impact of a product
254 system throughout its life cycle. Therefore, it is often considered as complementary and in line with the
255 circularity assessment. However, the LCA is based on the conventional “cradle to grave” approach and

255 even that it includes credits for the displaced materials and resources, its interpretation is not fully in
256 consonance with circular economy concept. From LCA point of view, the anaerobic treatment is
257 considered ultimately as an end of life solution (“...grave”), focused on the reduction of direct
258 environmental pressure from the generated dairy effluent. On the other hand, the circular economy is
259 based on the “cradle to cradle” principle, which shifts the perception of wastewater from waste flow to a
260 source of valuable materials (Zijp et al., 2017) where the AD is seen as resource recovery solution
261 (“...cradle”). Therefore, from circular economy perspective, the AD treatment can be defined as a
262 multifunctional process with two main functions with relevant “environmental benefits”: 1) treatment of
263 the wastewater to reduce direct environmental pressure to the water body (primary function); and 2)
264 recovery of energy and valuable resources which brings indirect environmental benefits as a result of the
265 avoided virgin material and energy sources (secondary function). However, to fulfil these two functions
266 the AD requires the use of external resources such as energy, water and chemicals for its operation, which
267 on the other side are associated with negative indirect environmental impact – “environmental costs” (Fig.
268 1). In this sense, one can argue that an end of life solution is more circular when its overall
269 “environmental benefits” outweighs its “environmental costs” including the indirect and direct impact
270 generated within the foreground and background systems at the level where the circularity loops are
271 closed. Therefore, an Environmental Circularity Performance Indicator (ECPI) based on the ratio of the
272 total environmental benefits and costs is proposed to measure the circular environmental performance.
273 The ReCiPe Endpoint (H) 1.12 method has been applied in order to normalise and aggregate both the
274 direct and indirect environmental impact categories into one single score indicator. The ECPI indicator for
275 the circularity performance of the AD is defined according to Equation 2:

$$ECPI_{AD}^{Li} = \frac{EB_{direct}^{Li} + EB_{indirect}^{Li}}{EC_{direct}^{Li} + EC_{indirect}^{Li}}$$

279 ,where

280 EB_{direct}^{Li} is the direct (foreground) environmental benefit i.e. the reduced environmental pressure
281 (e.g. avoided eutrophication, emissions from sludge disposal);

282 $EB_{indirect}^{Li}$ is the indirect (background) environmental benefit i.e. avoided environmental impact
283 from recovered products (e.g. energy, fertilisers);

284 EC_{direct}^{Li} is the direct (foreground) environmental cost (e.g. emitted CH₄ and CO₂ emissions);

285 $EC_{indirect}^{Li}$ is the indirect (background) environmental cost (e.g. embodied emissions to the
286 resources used, transportation)

287 L_i , is the level at which the environmental assessment is performed

288 The indicator provides an aggregated metric of the environmental performance of a circular solution
289 whereas an output value less than 1 indicates a negative environmental circularity performance and more
290 than 1 indicates a positive performance.

291 2.6 Sensitivity analysis

292 The inherent uncertainties regarding the method used, the initial assumptions and the quality of the data
293 could affect the results. In order to address this issue, a sensitivity analysis of the main inputs has been
294 conducted. For this purpose, a $\pm 10\%$ change from the average of the main input parameters has been
295 simulated and the relevant effects on each impact category were calculated, assuming all other factors
296 remained fixed at their mean level. The sensitivity results aim to reveal the parameters that contribute the
297 most to the selected impact categories.

299 3. Results and discussion

300 3.1 Environmental assessment results

301 The characterisation impact assessment results for the three sub-system boundary levels are shown in
302 Table 1. The negative values at Level 1 indicate net positive environmental impact for these impact
303 categories meaning that the environmental benefits outweigh the environmental costs.

304

305 **Table 4 Characterisation impact assessment results**

Impact Category	Abbreviation	Unit	Level 1	Level 2	Level 3
			<i>FU 1 m³</i>	<i>FU 1 kg FPCM</i>	<i>FU 1 kg FPCM</i>
Climate change	CC	kg CO ₂ eq	-15.1	0.0991	1.92
Ozone depletion	OD	g CFC-11 eq	-4.50E-03	2.32E-05	1.06E-04
Freshwater eutrophication	FE	g P eq	-4.28	0.0240	2.76
Ionising radiation	IR	kBq U235 eq	-10.25	0.0242	0.0792
Agricultural land occupation	ALO	m ² a	-1.96	0.0192	1.44
Water depletion	WD	m ³	0.415	1.11E-03	0.0404
Metal depletion	MD	g Fe eq	3.1	2.30	61.9
Fossil depletion	FD	kg oil eq	13.7	0.0506	0.241

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307 ***Environmental performance assessment of the AD system (Level 1)***

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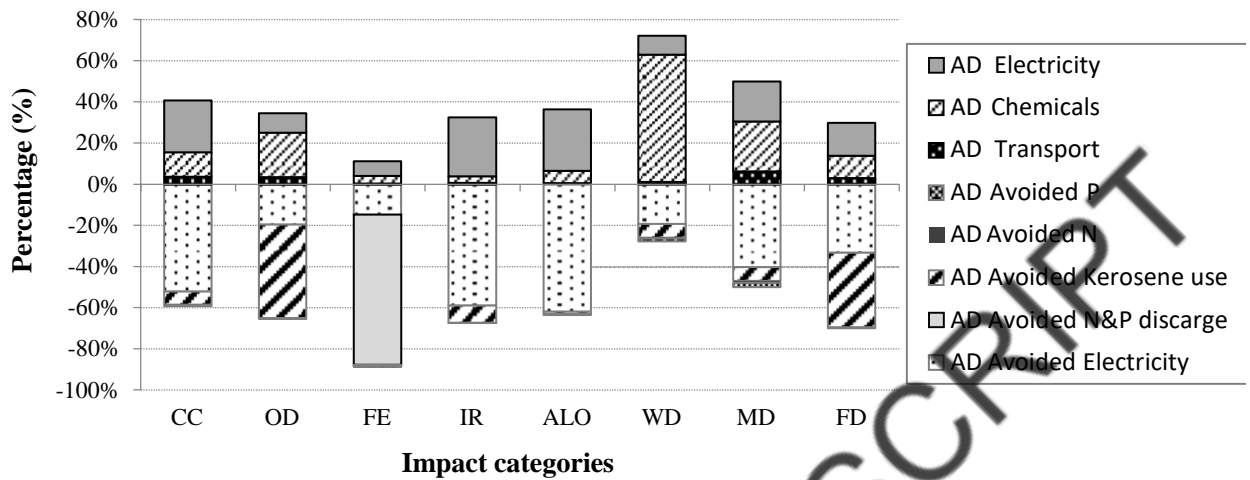
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Error! Reference source not found. shows the relative contributions to the impact categories of the environmental performance of the anaerobic digestion system. The environmental benefits resulting from the reduced N and P discharge to the receiving water bodies, valorisation of heat, energy and digestate that is applied as fertiliser are shown in the figure with negative contributions. The AD process for the treatment of the dairy effluent results in environmental benefits for most of the impact categories. The largest environmental benefit is in the eutrophication impact category (around 80%) due to the of N and P removal in the AD treatment process. Approximately 48% of the generated electricity is utilized for the operation of the anaerobic digestion facility (about 399 MWh/annum), while the surplus electricity (426 MWh/annum) is used to cover the dairy processing facility electricity needs. In the dairy processing facility, a kerosene boiler is used for heating purposes; thus, the CHP heat replaces equivalent heat produced from the boiler (kerosene). The electricity produced from the combustion of the biogas, is mainly responsible for the environmental benefits of the AD process with overall contribution ranging from 20% to 70%, for all impact categories. Additionally, the avoided impacts from the utilization of the heat produced in the CHP unit contribute to OD and FD impact categories (relative contributions equal to 41% and 32% respectively). On the contrary, the chemicals used in the AD plant are mainly responsible

323 for the negative environmental impacts for most of the impact categories and particularly for the OD and
 324 WD (24% and 64% respectively).



325

326 **Figure 3:** Relative contributions to each impact category from different activities involved in the level 1
 327 assessment.

328 There are only few studies available on the life-cycle based environmental analysis of AD process in the
 329 UK using waste as feedstock, with limited information on the operating characteristics and the mass
 330 balances of the systems. Whiting and Azapagic (2014), assessed the environmental impacts using the
 331 CML 2011 method of a UK AD-CHP plant operating with a mix of different agricultural wastes.
 332 Similarly, Styles et al. (2015) implemented CML 2010 method to determine the environmental impacts of
 333 AD installations in UK dairy farms. However, the results of the current work cannot be directly compared
 334 with the cited studies since different impact assessment methodologies or functional units were used.
 335 Nevertheless, all studies have concluded that the displacement of kerosene with the heat from AD
 336 contributes significantly to mitigate CO₂ emissions and fossil fuel depletion leading to an overall net
 337 negative climate change impact.

338 ***Environmental performance of the dairy processing facility (Level 2)***

339 The expansion of the system boundaries to the dairy facility (level 2) provided insight on the most
 340 significant contributors to its environmental profile, shown in Error! Reference source not found.. It can
 341 be seen that that the avoided direct eutrophication potential is almost equal to the indirect eutrophication

342 associated with the resources used in the dairy processing. The excess energy produced from the
343 anaerobic digestion unit, is equal to 188 ton of CO₂eq savings in the facility per annum. Benefits are also
344 observed in IR, MD and FD impact categories (19%, 14% and 12% respectively).

345 One of the most significant contributors to the majority of the impact categories is the energy
346 requirements of the dairy facility. About 3687 MWh of electricity is required annually for the processing
347 of the dairy products, whereas ~450 ton of kerosene are used for the heating requirements in the
348 processing stages. The electricity and fossil fuels consumption in the dairy plant are equal to 1.25 kWh/L
349 and 1.93 kWh/L of milk processed respectively. A wide range of electricity consumption has been
350 reported for the European dairy sector based on different dairy products ranging from 0.15 – 2.5 kWh per
351 kg of liquid milk processed for the production of milk and yoghurt products to 0.08 – 2.9 kWh per kg of
352 liquid milk processed for the production of cheese products (Expo and Sevilla, 2003). In terms of fossil
353 fuels consumption, the reported values range from 0.18 – 1.5 kWh per kg of liquid milk processed for the
354 production of milk and yoghurt products, to 0.15 – 4.6 kWh per kg of liquid milk processed for the
355 production of cheese products. Therefore, the energy requirements of the examined dairy facility (mixture
356 of milk, cheese, yoghurt products) are moderate compared to the respective ones of the European dairy
357 industry.

358 Additionally, the packaging materials are identified as ‘hotspot’ in the majority of the categories,
359 especially for ALO (relative contribution equal to ~69%) and OD (relative contribution equal to ~50%).
360 The packaging is responsible for around 30% of the carbon footprint emissions in the dairy facility. The
361 environmental impacts of packaging in dairy products have also been identified as an environmental
362 hotspot in other research works (González-García et al., 2013a; Vasilaki et al., 2016). Bio-based
363 packaging materials (i.e PLA, polyhydroxyalkanoates (PHA) etc.) have been proposed in the literature as
364 alternatives to conventional synthetic polymers towards the mitigation of the environmental impacts of
365 food packaging (Licciardello, 2017).

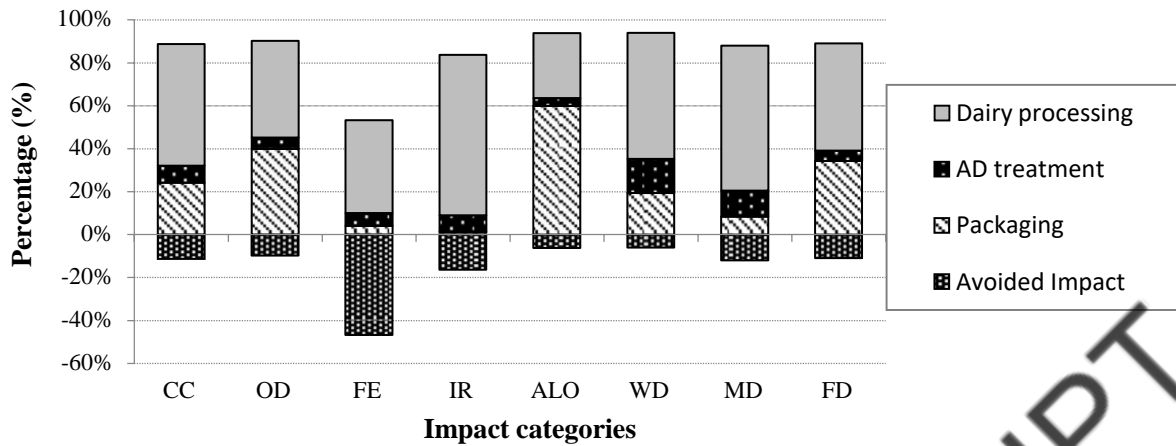


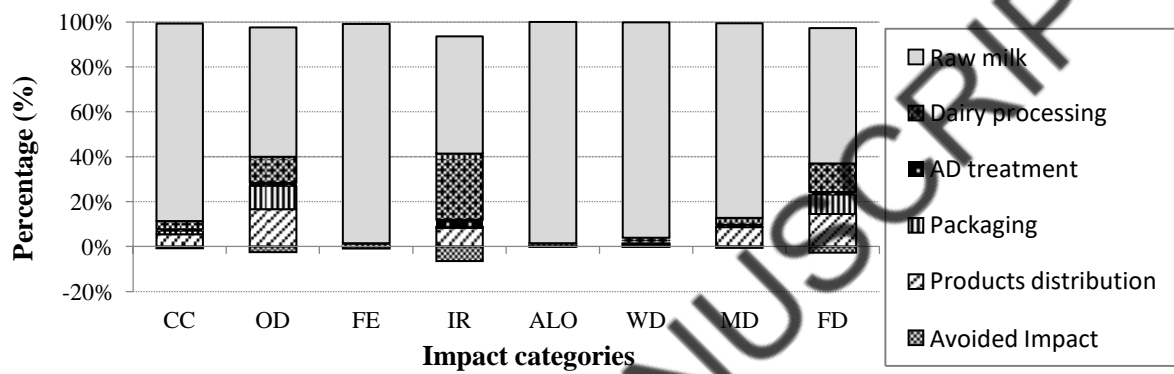
Figure 4: Relative contributions of activities involved at level 2 to each impact category

Direct water consumption in the plant mainly affects the WD impact category (20% relative contribution) with 4.2 L water consumed per L of milk processed. According to the European Commission Directorate (Expo and Sevilla, 2003), water consumption for the processing of milk and yoghurt products, varies from 0.8 to 25 L/kg of processed milk whereas the range for cheese products is even higher and equal to 1-60 L/kg of processed milk. Therefore, water consumption in the processing plant is relatively low compared to European average.

Environmental assessment of the entire dairy supply chain (Level 3)

The relative contribution of each sub-system on the entire dairy supply chain is shown on **Error! Reference source not found.** (dairy farm, processing plant, anaerobic digestion system) to the examined impact categories. The production of raw milk is the most significant contributor to all impact categories examined and almost the sole contributor for ALO, FE and WD (relative contribution equal to 96-99%). These findings correspond to the outcomes of other studies. Palmieri and Salimei (Palmieri et al., 2017) examined the effect of different cow feeding strategies on the environmental profile of cheese products and they demonstrated that irrespectively of the feeding strategy, raw milk is the main contributor of the environmental impacts in the dairy value chains. Previous studies assessing various dairy products have stressed the significance of the farm system (Fantin et al., 2012; Finnegan et al., 2015; González-García et al., 2013b) with contributions to the total carbon footprint of the products ranging from 81% to 93%. Even though the environmental impacts related to the production of raw milk affect significantly the

386 profile of the dairy end-products the dairy processing facility contributes significantly to OD, IR, FD, MD
 387 and CC impact categories (relative contributions equal to 12%, 34%, 13%, 3% and 4% respectively).
 388 Significant environmental impact is attributed also to the distribution of the final products. Dairy
 389 products' distribution to retailers accounts for 6% of the total carbon footprint and 17% of the OD, which
 390 is attributed to the GHG emissions emitted from the truck's fuel combustion and the long distribution
 391 routes.



392
 393 **Figure 5** Relative contributions of various activities involved at level 3 to each impact category.

394 ***Sensitivity analysis***

395 Fig. 6 shows the result of the sensitivity analysis performed regarding the main resource inputs
 396 (chemicals, transportation and energy) in the dairy supply chain and their endpoint impact at each system
 397 level. The results are based on 10% increase and 10% decrease in the average values of the individual
 398 input parameters. The energy has significant impact at all system levels with main contribution on Level 1
 399 and Level 2. This is interpreted by the high dependency of dairy facility processes and AD on energy use
 400 (electricity and kerosene). Transportation is the major contributor to Level 3 with highest impact in the
 401 whole dairy supply chain.

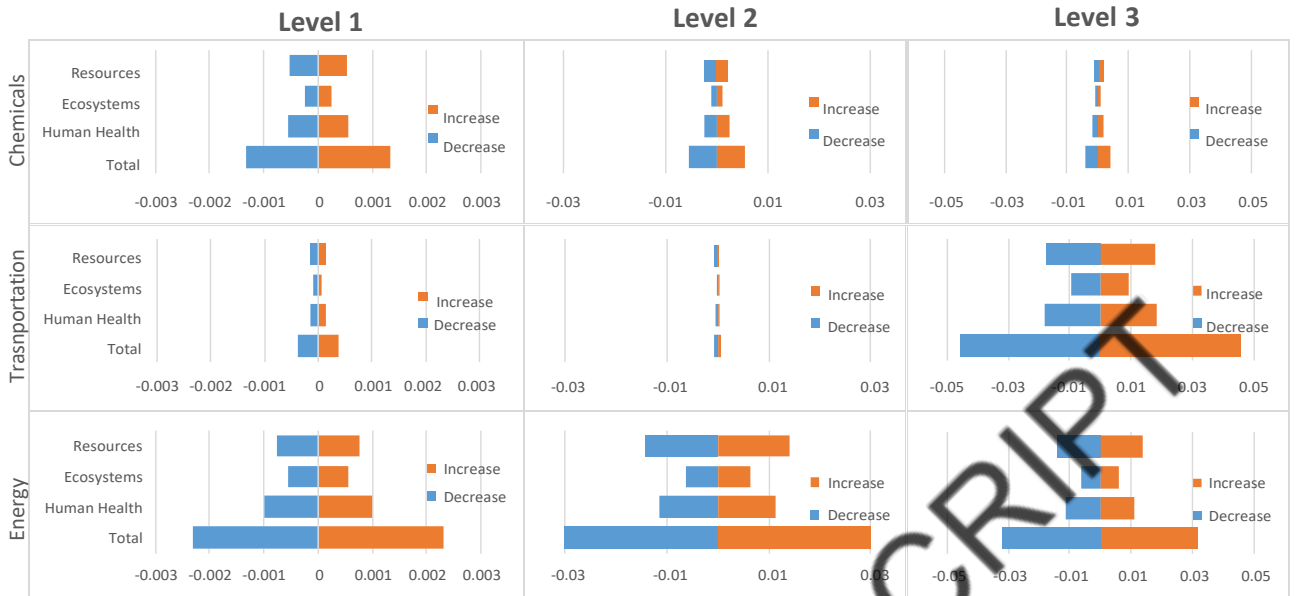


Figure 6 Multilevel sensitivity analyses

3.2. Circularity performance evaluation

3.2.1. Material circularity

Material circularity is a key element in the circular economy, which addresses the challenges of resource scarcity, whilst delivering the same functionality. Within this context, AD has many advantages, since apart from its primary function to treat the wastewater and remove organic content, it also produces a valuable by-product (biogas), that can be recovered and utilized as a fuel; and generate sludge that can be utilised as fertilizer. The recovered N and P fertilizers are key elements in the biological nutrient cycle of the circular economy, especially the P, which is a scarce and finite resource. In this case, it is considered that the recovered energy and materials from AD are recycled in a closed loop to the dairy factory (Level 2) and to the farm stage (Level 3). It is assumed that the solid fraction of the digestate is applied directly to the land as a fertiliser for crop production to feed the cows. As described in the methodology section, the material circularity indicator (MCPI) was calculated based on the demand minimization index taking into account the actual reduction of the baseline demand at the level that the circular loop is closed. Table 5 presents the MCPI results that related to the valorisation of the recovered energy and nutrients at the relevant system level; level 1 has not been considered here as it represent the AD solution itself. The

419 water pathway has not been included in the analysis, since the irrigation of pastures is not considered to
 420 be economically viable in the UK. The MCPI shows that the energy obtained from the anaerobic
 421 treatment leads to a reduction of 11.55% for electricity and 26.61% for kerosene consumption of the
 422 plant. At level 3 the nutrients derived from the dairy sludge account for about 0.033% N and 5.41% P
 423 fertilizer displacement at the farm. The MCPI indicators show that the AD solution has a considerable
 424 potential towards a more circular management in the dairy supply chain by minimising the baseline use of
 425 raw resources and providing a more self-sustaining system.

426 **Table 5** Material circularity performance indicator results

System boundary		MCPI			
level at which the CE loop is closed	Resource	Baseline	Minimised		
		Demand	demand	[-]	[%]
Level 2		Energy			
	Kerosene [kg]	450,934	330,934	0.266	26.61%
Level 3		Nutrients			
	N fertilizer [kg]	1,214,400	1,214,003	0.00033	0.03%
	P fertilizer [kg]	42,240	39,953	0.054	5.41%

427 3.2.2. Environmental circularity

428 When applying the LCA following the circular economy mind-set it is important, as mentioned above, to
 429 consider very carefully the system boundaries of the study. A standard application of the LCA and
 430 definition of system boundaries may not properly account for the environmental effects of the recovered
 431 materials and resources that cross the system boundary and are utilised in the upstream or downstream of
 432 the system or by other products or processes. A comprehensive LCA study that complies with the circular
 433 economy principles should extend the system boundary to the level that the circularity loops are closed
 434 considering processes and products beyond the initial life cycle, in order to correctly evaluate the
 435 circularity performance in the environmental domain for each potential scenario. Table 6 demonstrates the

436 application of this approach to the AD treatment solution based on the LCA analysis performed following
 437 the Endpoint (H) 1.12 method. Although the endpoint environmental assessment method is associated
 438 with increased uncertainty of the results, in this case it is considered reliable enough to best represent the
 439 aggregated environmental cost-benefit ratio.

440 **Table 6** Environmental circularity performance indicator calculation

System boundary level at which the CE loop is closed	Damage Category	Environmental benefits	Environmental costs	ECPI
		$EB_{direct}^{Li} + EB_{indirect}^{Li}$	$EC_{direct}^{Li} + EC_{indirect}^{Li}$	[-]
Level 2 (Energy CE pathway)	Human health [kPt]	23.55	17.04	1.38
	Ecosystems [kPt]	13.16	8.67	1.52
	Resources[kPt]	32.34	14.39	2.25
	Total [kPt]	69.04	40.10	1.72
Level 3 (Nutrient CE pathway)	Human health [kPt]	6.63	1.68	3.95
	Ecosystems [kPt]	1.96	0.87	1.05
	Resources [kPt]	0.35	1.65	0.21
	Total [kPt]	8.94	4.20	2.13

441
 442 The ECPI indicators represent the environmental benefit-cost ratio performance for the energy and
 443 nutrient circularity loops at the relevant system level. At Level 2, the ECPI shows that the AD treatment
 444 has a positive environmental circularity performance in all damage categories varying from 1.38 to 2.25.
 445 The single score ratio of the total impact result in a value of 1.72 meaning that the overall environmental
 446 benefits are ~1.7 times higher than the related environmental costs induced from the AD operation. At
 447 Level 3, a clear trade-off between the damage categories is observed. In Human health category the
 448 environmental benefits are about four times higher than the environmental costs, while for Resources the
 449 opposite is true. However, as a single score ratio the nutrient recycling AD scenario also has a positive
 450 benefit-cost ratio of 2.13.

451

452

453 **4. Conclusions**

454 A two-dimensional approach for measuring the circular economy performance of AD dairy processing
455 effluents was proposed. The results of the work demonstrate the potential to close the material and energy
456 circular economy loops at different levels of the dairy supply chain. The assessment on the AD system
457 level (Level 1), showed significant net positive impact in most of the impact categories (CC, OD, FE, IR,
458 ALO, FD). The analysis at dairy factory level (Level 2) revealed the main “hotspots” of the dairy
459 processing facility, and showed that the AD total GHG emissions can be reduced by about 13%. The
460 values obtained for the MCPI and ECPI circularity indicators reveal the importance of the application of
461 the AD treatment as an instrument of circular economy solutions for dairy effluent treatment. The
462 application of the indicators provides quantitative measure of the material and environmental performance
463 of the energy and nutrient circularity valorisation pathways. The latter can facilitate operational decisions
464 for the implementation of circular economy models aiming at retaining the material value within the
465 system taking into account any possible trade-offs within and between these two domains of CE.

466

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479 **References**

480 Agudelo-Vera, C.M., Mels, A., Keesman, K., Rijnaarts, H., 2012. The Urban Harvest Approach as an Aid
481 for Sustainable Urban Resource Planning. *Journal of Industrial Ecology* 16, 839–850.
482 <https://doi.org/10.1111/j.1530-9290.2012.00561.x>

483 An, C., He, Y., Huang, G., Liu, Y., 2010. Performance of mesophilic anaerobic granules for removal of
484 octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine (HMX) from aqueous solution. *Journal of Hazardous*
485 *Materials* 179, 526–532. <https://doi.org/10.1016/j.jhazmat.2010.03.035>

486 Battini, F., Agostini, A., Boulamanti, A.K., Giuntoli, J., Amaducci, S., 2014. Mitigating the
487 environmental impacts of milk production via anaerobic digestion of manure: Case study of a dairy farm
488 in the Po Valley. *Science of The Total Environment* 481, 196–208.
489 <https://doi.org/10.1016/j.scitotenv.2014.02.038>

490 Berlin, J., 2002. Environmental life cycle assessment (LCA) of Swedish semi-hard cheese. *International*
491 *Dairy Journal* 12, 939–953. [https://doi.org/10.1016/S0958-6946\(02\)00112-7](https://doi.org/10.1016/S0958-6946(02)00112-7)

492 Blomsma, F., Brennan, G., 2017. The Emergence of Circular Economy: A New Framing Around
493 Prolonging Resource Productivity. *Journal of Industrial Ecology* 21, 603–614.
494 <https://doi.org/10.1111/jiec.12603>

495 Carlini M, Castellucci S, Moneti M, 2015. Biogas production from poultry manure and cheese whey
496 wastewater under mesophilic conditions in batch reactor. *Energy Procedia* 82, 811–818.
497 <https://doi.org/10.1016/j.egypro.2015.11.817>

498 Traversi, D., Bonetta, S., Degan, R., Villa, S., Porfido, A., Bellero, M., Carraro, E., Gilli, G., 2013.
499 *Environmental Advances Due to the Integration of Food Industries and Anaerobic Digestion for Biogas*

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500 Production: Perspectives of the Italian Milk and Dairy Product Sector. *BioEnergy Research*. 6, 851–863.
501 <https://doi.org/10.1007/s12155-013-9341-4>

502 Carvalho, F., Prazeres, A.R., Rivas, J., 2013. Cheese whey wastewater: Characterization and treatment.
503 *Science of the Total Environment* 445–446, 385–396. <https://doi.org/10.1016/j.scitotenv.2012.12.038>

504 Cederberg, C., Mattsson, B., 2000. Life cycle assessment of milk production — a comparison of
505 conventional and organic farming. *Journal of Cleaner Production* 8, 49–60.
506 [https://doi.org/10.1016/S0959-6526\(99\)00311-X](https://doi.org/10.1016/S0959-6526(99)00311-X)

507 Demirel, B., Yenigün, O., 2006. Changes in microbial ecology in an anaerobic reactor. *Bioresource*
508 *Technology* 97, 1201–1208. <https://doi.org/10.1016/j.biortech.2005.05.009>

509 Demirel, B., Yenigun, O., Onay, T.T., 2005. Anaerobic treatment of dairy wastewaters: A review. *Process*
510 *Biochemistry* 40, 2583–2595. <https://doi.org/10.1016/j.procbio.2004.12.015>

511 Dolman, M.A., Sonneveld, M.P.W., Mollenhorst, H., De Boer, I.J.M., 2014. Benchmarking the economic,
512 environmental and societal performance of Dutch dairy farms aiming at internal recycling of nutrients.
513 *Journal of Cleaner Production*, 73, 245–252. <https://doi.org/10.1016/j.jclepro.2014.02.043>

514 European Academies Science Advisory Council, Deutsche Akademie der Naturforscher Leopoldina
515 (Eds.), 2016. Indicators for a circular economy, EASAC policy report 30. EASAC Secretariat, Deutsche
516 Akademie der Naturforscher Leopoldina, Halle (Saale). ISBN: 978-3-8047-3680-1.

517 EC, 2015. Closing the loop on the ground: 10 EU projects working towards a circular economy -
518 European Commission [WWW Document] URL
519 [https://ec.europa.eu/easme/en/news/closing-loop-ground-10-eu-projects-working-towards-circular-](https://ec.europa.eu/easme/en/news/closing-loop-ground-10-eu-projects-working-towards-circular-economy)
520 [economy](https://ec.europa.eu/easme/en/news/closing-loop-ground-10-eu-projects-working-towards-circular-economy) (accessed 6.17.18).

521 Fantin, V., Buttol, P., Pergreffi, R., Masoni, P., 2012. Life cycle assessment of Italian high quality milk
522 production. A comparison with an EPD study. *Journal of Cleaner Production*, 28, 150–159.
523 <https://doi.org/10.1016/j.jclepro.2011.10.017>

524 Finnegan, W., Goggins, J., Clifford, E., Zhan, X., 2017. Global warming potential associated with dairy
525 products in the Republic of Ireland. *Journal of Cleaner Production*, 163, 262–273.
526 <https://doi.org/10.1016/j.jclepro.2015.08.025>

527 Flysjö, A., Henriksson, M., Cederberg, C., Ledgard, S., Englund, J.-E., 2011. The impact of various
528 parameters on the carbon footprint of milk production in New Zealand and Sweden. *Agricultural Systems*
529 104, 459–469. <https://doi.org/10.1016/j.agsy.2011.03.003>

530 IDF, 2015. IDF A common carbon footprint approach for the dairy sector. The IDF guide to standard
531 lifecycle assessment methodology Bulletin of the International Dairy Federation, No. 479/2015. [WWW
532 Document] URL [https://store.fil-idf.org/product/a-common-carbon-footprint-approach-for-the-dairy-](https://store.fil-idf.org/product/a-common-carbon-footprint-approach-for-the-dairy-sector-the-idf-guide-to-standard-life-cycle-assessment-methodology/)
533 [sector-the-idf-guide-to-standard-life-cycle-assessment-methodology/](https://store.fil-idf.org/product/a-common-carbon-footprint-approach-for-the-dairy-sector-the-idf-guide-to-standard-life-cycle-assessment-methodology/) (accessed 3.15.19).

534 Freibauer, A., Mathijs, E., Brunori, G., Damianova, Z., Faroult, E., Gomis, J.G. i, O'Brien, L., Treyer, S.,
535 2011. Sustainable Food Consumption and Production in a Resource-constrained World. *EuroChoices* 10,
536 38–43. <https://doi.org/10.1111/j.1746-692X.2011.00201.x>

537 Gelegenis, J., Georgakakis, D., Angelidaki, I., Mavris, V., 2007. Optimization of biogas production by co-
538 digesting whey with diluted poultry manure. *Renewable Energy* 32, 2147–2160.
539 <https://doi.org/10.1016/j.renene.2006.11.015>

540 Ghisellini, P., Cialani, C., Ulgiati, S., 2016. A review on circular economy: the expected transition to a
541 balanced interplay of environmental and economic systems. *Journal of Cleaner Production*, Towards Post
542 Fossil Carbon Societies: Regenerative and Preventative Eco-Industrial Development 114, 11–32.
543 <https://doi.org/10.1016/j.jclepro.2015.09.007>

544 Godfray, H.C.J., Beddington, J.R., Crute, I.R., Haddad, L., Lawrence, D., Muir, J.F., Pretty, J., Robinson,
545 S., Thomas, S.M., Toulmin, C., 2010. Food Security: The Challenge of Feeding 9 Billion People. *Science*
546 327, 812–818. <https://doi.org/10.1126/science.1185383>

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60
61
62
63
64
65

547 Goedkoop, M., Heijungs, R., Huijbregts, M., De Schryver, A., Struijs, J., Van Zelm, R., 2009. ReCiPe
548 2008. A life cycle impact assessment method which comprises harmonised category indicators at the
549 midpoint and the endpoint level 1.

550 González-García, S., Castanheira, É.G., Dias, A.C., Arroja, L., 2013a. Environmental life cycle
551 assessment of a dairy product: the yoghurt. *The International Journal of Life Cycle Assessment* 18, 796–
552 811. <https://doi.org/10.1007/s11367-012-0522-8>

553 González-García, S., Hospido, A., Moreira, M.T., Feijoo, G., Arroja, L., 2013b. Environmental Life
554 Cycle Assessment of a Galician cheese: San Simon da Costa. *Journal of Cleaner Production* 52, 253–262.
555 <https://doi.org/10.1016/j.jclepro.2013.03.006>

556 Gottschall, N., Boutin, C., Crolla, A., Kinsley, C., Champagne, P., 2007. The role of plants in the removal
557 of nutrients at a constructed wetland treating agricultural (dairy) wastewater, Ontario, Canada. *Ecological*
558 *Engineering* 29, 154–163. <https://doi.org/10.1016/j.ecoleng.2006.06.004>

559 Hawkesworth S., Dangour A.D., Johnston D., Lock K., Poole N., Rushton J., Uauy R., Waage J., 2010.
560 Feeding the world healthily: the challenge of measuring the effects of agriculture on health. *Philosophical*
561 *Transactions of the Royal Society B: Biological Sciences* 365, 3083–3097.
562 <https://doi.org/10.1098/rstb.2010.0122>

563 Hospido, A., Moreira, M.T., Feijoo, G., 2003. Simplified life cycle assessment of galician milk
564 production. *International Dairy Journal* 13, 783–796. [https://doi.org/10.1016/S0958-6946\(03\)00100-6](https://doi.org/10.1016/S0958-6946(03)00100-6)

565 Huysman, S., Sala, S., Mancini, L., Ardente, F., Alvarenga, R.A.F., De Meester, S., Mathieux, F., Dewulf,
566 J., 2015. Toward a systematized framework for resource efficiency indicators. *Resources, Conservation*
567 *and Recycling* 95, 68–76. <https://doi.org/10.1016/j.resconrec.2014.10.014>

568 IPCC, 2006. 2006 IPCC Guidelines for National Greenhouse Gas Inventories. Intergovernmental Panel on
569 Climate Change.

570 ISO 14040 (2006) Environmental management—life cycle assessment—principles and framework.
571 International Organization for Standardization, Geneva

572 ISO 14044 (2006) Environmental management—life cycle assessment—requirements and guidelines.
573 International Organization for Standardization, Geneva

574 Barnett J.W., Robertson S.L. and Russell J.M., 2010. Environmental issues in dairy processing.
575 Environment Portfolio, New Zealand Dairy Research Institute, Private Bag 11029, Palmerston North

576 Jawahir, I.S., Bradley, R., 2016. Technological Elements of Circular Economy and the Principles of 6R-
577 Based Closed-loop Material Flow in Sustainable Manufacturing. Procedia CIRP, 13th Global Conference
578 on Sustainable Manufacturing – Decoupling Growth from Resource Use 40, 103–108.
579 <https://doi.org/10.1016/j.procir.2016.01.067>

580 Karadag, D., Koroğlu, O.E., Ozkaya, B., Cakmakci, M., 2015. A review on anaerobic biofilm reactors for
581 the treatment of dairy industry wastewater. Process Biochemistry 50, 262–271.
582 <https://doi.org/10.1016/j.procbio.2014.11.005>

583 Kimming, M., Sundberg, C., Nordberg, Å., Baky, A., Bernesson, S., Hansson, P.-A., 2015. Replacing
584 fossil energy for organic milk production – potential biomass sources and greenhouse gas emission
585 reductions. Journal of Cleaner Production, 106, 400–407. <https://doi.org/10.1016/j.jclepro.2014.03.044>

586 Kirilova, E.G., Vaklieva-Bancheva, N.G., 2017. Environmentally friendly management of dairy supply
587 chain for designing a green products' portfolio. Journal of Cleaner Production 167, 493–504.
588 <https://doi.org/10.1016/j.jclepro.2017.08.188>

589 Kılış, Ş., Kılış, B., 2017. Integrated circular economy and education model to address aspects of an
590 energy-water-food nexus in a dairy facility and local contexts. Journal of Cleaner Production 167, 1084–
591 1098. <https://doi.org/10.1016/j.jclepro.2017.03.178>

592 Licciardello, F., 2017. Packaging, blessing in disguise. Review on its diverse contribution to food
593 sustainability. Trends in Food Science & Technology 65, 32–39.
594 <https://doi.org/10.1016/j.tifs.2017.05.003>

595 Linder, M., Sarasini, S., Loon, P. van, 2017. A Metric for Quantifying Product-Level Circularity. Journal
596 of Industrial Ecology 21, 545–558. <https://doi.org/10.1111/jiec.12552>

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
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18
19
20
21
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46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

597 Lybæk, R., Kjær, T., 2019. Pre-assessment of the Circular Economic Benefits and Challenges of Biogas
598 Production in Denmark when Utilizing Sand Bedding in Dairy Cow Stables. *Journal of Cleaner*
599 *Production* 219, 268–277. <https://doi.org/10.1016/j.jclepro.2019.01.241>

600 Ellen MacArthur Foundation and the McKinsey Center for Business and Environment. 2015. *Growth*
601 *within: A circular economy vision for a competitive Europe*. Dublin, Ireland: Foundation for
602 *Environmental Economics and Sustainability*

603 Malaspina, F., Cellamare, C.M., Stante, L., Tilche, A., 1996. Anaerobic treatment of cheese whey with a
604 downflow-upflow hybrid reactor. *Bioresource Technology* 55, 131–139. [https://doi.org/10.1016/0960-](https://doi.org/10.1016/0960-8524(95)00187-5)
605 [8524\(95\)00187-5](https://doi.org/10.1016/0960-8524(95)00187-5)

606 Mezzullo, W.G., McManus, M.C., Hammond, G.P., 2013. Life cycle assessment of a small-scale
607 anaerobic digestion plant from cattle waste. *Applied Energy*, 102, 657–664.
608 <https://doi.org/10.1016/j.apenergy.2012.08.008>

609 Murgia, L., Todde, G., Caria, M., Pazzona, A., 2013. A partial life cycle assessment approach to evaluate
610 the energy intensity and related greenhouse gas emission in dairy farms. *Journal of Agricultural*
611 *Engineering* 44. <https://doi.org/10.4081/jae.2013.279>

612 Palhares, J.C.P., Pezzopane, J.R.M., 2015. Water footprint accounting and scarcity indicators of
613 conventional and organic dairy production systems. *Journal of Cleaner Production* 93, 299–307.
614 <https://doi.org/10.1016/j.jclepro.2015.01.035>

615 Palmieri, N., Forleo, M.B., Salimei, E., 2017. Environmental impacts of a dairy cheese chain including
616 whey feeding: An Italian case study. *Journal of Cleaner Production* 140, 881–889.
617 <https://doi.org/10.1016/j.jclepro.2016.06.185>

618 Pan, S.Y., Du, M.A., Huang, I.T., Liu, I.H., Chang, E.E., Chiang, P.C., 2015. Strategies on
619 implementation of waste-to-energy (WTE) supply chain for circular economy system: a review. *Journal of*
620 *Cleaner Production* 108, 409–421. <https://doi.org/10.1016/j.jclepro.2015.06.124>

621 Passeggi, M., López, I., Borzacconi, L., 2012. Modified UASB reactor for dairy industry wastewater:
622 performance indicators and comparison with the traditional approach. *Journal of Cleaner Production* 26,
623 90–94. <https://doi.org/10.1016/j.jclepro.2011.12.022>

624 Prasad, P., Pagan, R., Kauter, M., Price, N., Crittenden, P., 2004. Eco-efficiency for the Dairy-Processing
625 Industry. The UNEP Working Group for Cleaner Production in the Food Industry, St. Lucia and
626 Sustainable Business, Sydney, Australia.

627 Prazeres, A.R., Carvalho, F., Rivas, J., 2012. Cheese whey management: A review. *Journal of*
628 *Environmental Management* 110, 48–68. <https://doi.org/10.1016/j.jenvman.2012.05.018>

629 Roy, P., Nei, D., Orikasa, T., Xu, Q., Okadome, H., Nakamura, N., Shiina, T., 2009. A review of life
630 cycle assessment (LCA) on some food products. *Journal of Food Engineering* 90, 1–10.
631 <https://doi.org/10.1016/j.jfoodeng.2008.06.016>

632 Salou, T., Le Mouél, C., van der Werf, H.M.G., 2017. Environmental impacts of dairy system
633 intensification: the functional unit matters! *Journal of Cleaner Production*, 140, 445–454.
634 <https://doi.org/10.1016/j.jclepro.2016.05.019>

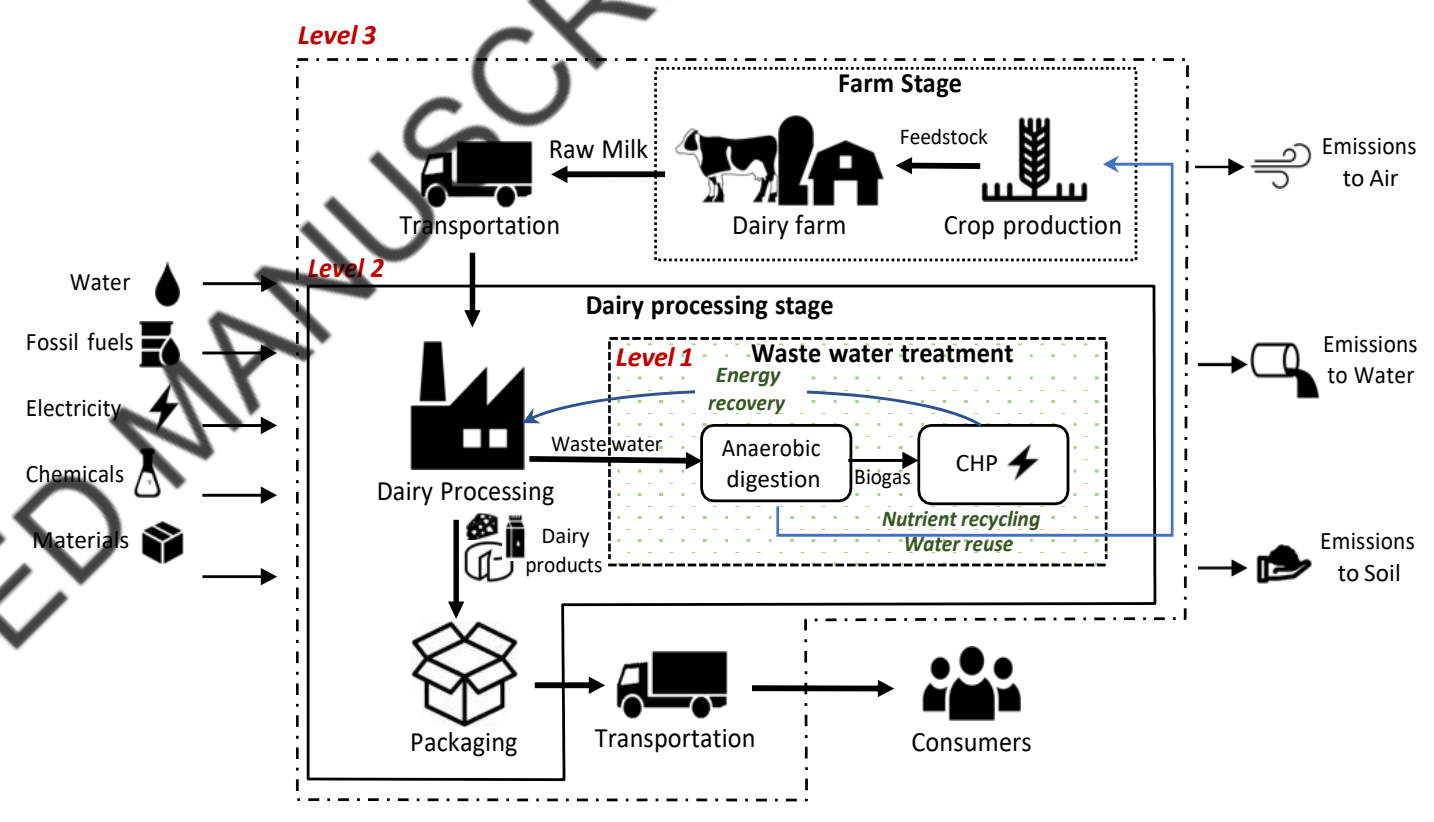
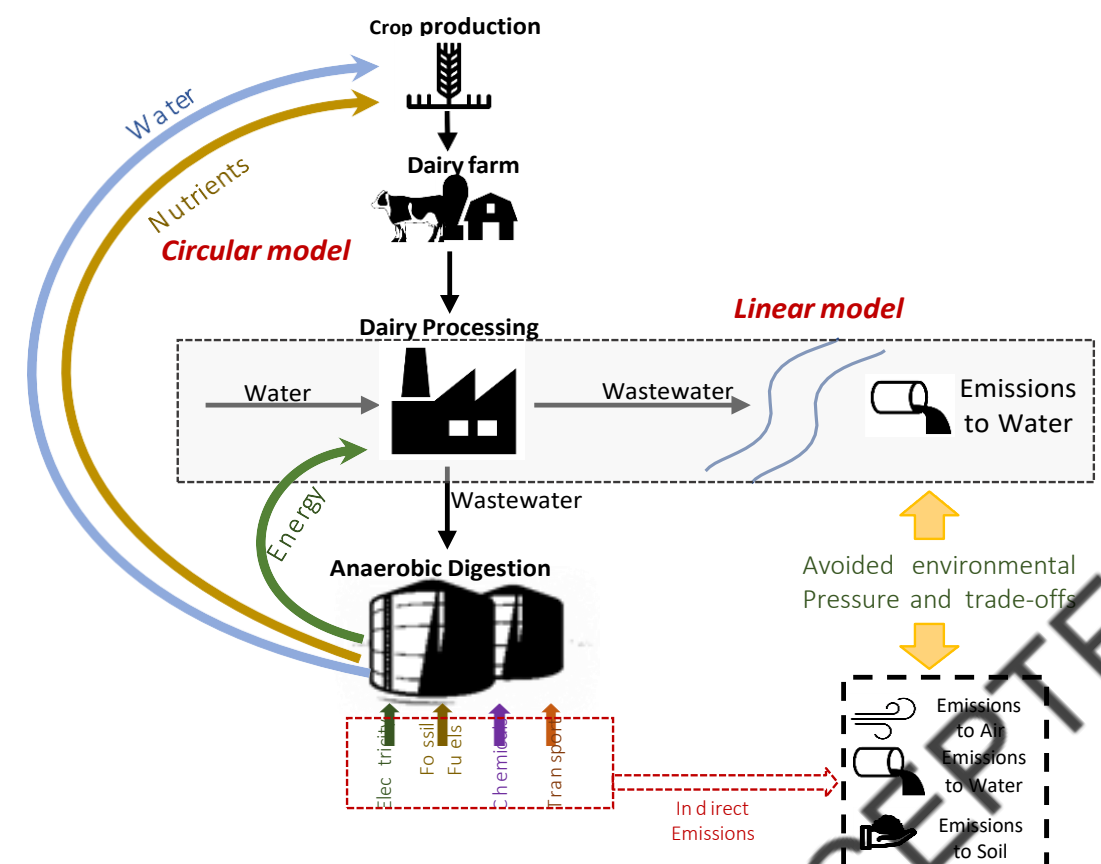
635 Strydom JP, Britz TJ, M.J., 1997. Two-phase anaerobic digestion of three different effluents using a
636 hybrid bioreactor. *Water Salination* 23, 151–6.

637 Styles, D., Gibbons, J., Williams, A.P., Stichnothe, H., Chadwick, D.R., Healey, J.R., 2015. Cattle feed or
638 bioenergy? Consequential life cycle assessment of biogas feedstock options on dairy farms. *GCB*
639 *Bioenergy* 7, 1034–1049. <https://doi.org/10.1111/gcbb.12189>

640 MacArthur, E. 2012. *Towards the Circular Economy Vol.1: Economic and Business Rationale for a*
641 *Circular Economy*. Ellen Macarthur Foundation.

642 Vasilaki, V., Katsou, E., Ponsá, S., Colón, J., 2016. Water and carbon footprint of selected dairy products:
643 A case study in Catalonia. *Journal of Cleaner Production* 139, 504–516.
644 <https://doi.org/10.1016/J.JCLEPRO.2016.08.032>

- 645 Weaver, C.M., 2010. Role of dairy beverages in the diet. *Physiology & Behavior, Beverages and Health*
1
2 646 100, 63–66. <https://doi.org/10.1016/j.physbeh.2010.01.020>
3
4
5 647 Whiting, A., Azapagic, A., 2014. Life cycle environmental impacts of generating electricity and heat from
6
7 648 biogas produced by anaerobic digestion. *Energy* 70, 181–193.
8
9 649 <https://doi.org/10.1016/j.energy.2014.03.103>
10
11
12 650 Wijnands, J.H.M., Verhoog, A.D., 2016. Competitiveness of the EU food industry : ex-post assessment of
13
14 651 trade performance embedded in international economic theory (No. 2016– 018). LEI Wageningen UR,
15
16 652 Wageningen.
17
18
19 653 WssTP, 2016. The Value of Water - Multiple Waters, for multiple purposes and user for multiple
20
21 654 purposes and users, Strategic Innovation and Research Agenda (SIRA).
22
23
24 655 Wu, H., Shi, Y., Xia, Q., Zhu, W., 2014. Effectiveness of the policy of circular economy in China: A
25
26 656 DEA-based analysis for the period of 11th five-year plan. *Resources, conservation and recycling* 83, 163–
27
28 657 175. <https://doi.org/10.1016/j.resconrec.2013.10.003>
29
30
31
32 658 Zhong, J., Stevens, D.K., Hansen, C.L., 2015. Optimization of anaerobic hydrogen and methane
33
34 659 production from dairy processing waste using a two-stage digestion in induced bed reactors (IBR).
35
36 660 *International Journal of Hydrogen Energy* 40, 15470–15476.
37
38 661 <https://doi.org/10.1016/j.ijhydene.2015.09.085>
39
40
41 662 Zijp, M.C., Waaijers-van der Loop, S.L., Heijungs, R., Broeren, M.L.M., Peeters, R., Van
42
43 663 Nieuwenhuijzen, A., Shen, L., Heugens, E.H.W., Posthuma, L., 2017. Method selection for sustainability
44
45 664 assessments: The case of recovery of resources from waste water. *Journal of Environmental Management*
46
47 665 197, 221–230. <https://doi.org/10.1016/j.jenvman.2017.04.006>
48
49
50
51 666 Zonderland-Thomassen, M.A., Ledgard, S.F., 2012. Water footprinting – A comparison of methods using
52
53 667 New Zealand dairy farming as a case study. *Agricultural Systems* 110, 30–40.
54
55 668 <https://doi.org/10.1016/j.agsy.2012.03.006>
56
57
58
59 669



ACCEPTED MANUSCRIPT

Avoided environmental Pressure and trade-offs

Indirect Emissions

Highlights

- Implementation of circular economy concept in a dairy processing industry
- LCA assessment of the anaerobic treatment of dairy processing effluents
- Material and environmental indicators for measuring circularity performance

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