

Chapter 6: Possible effects of Clofibrat acid on gene expression

6.1 Introduction

Aquatic organisms are exposed to a variety of pollutants, all of which have the potential to have a detrimental effect. Clofibric acid, which has also been found to be present in the environment, has so far (within the realms of this thesis) been shown to have effects in fish (some more convincingly so than others) at the physiological and biochemical levels. This chapter attempts to find the mechanistic effects of clofibric acid on adult fathead minnows (if present), by looking at the molecular level.

It was decided that two genes would be considered for examination after exposure of fish to clofibric acid. The decision of which genes to look at was primarily made by identifying those genes already found to be up- or down-regulated by fibrates in humans (there were no data available from fish). It was also important to consider whether the genes had been previously identified in fish (more specifically in the fathead minnow). The chosen genes were Lipoprotein Lipase (LPL) and Glucose-6-Phosphate Dehydrogenase (G6PDH). LPL is a gene involved in lipid metabolism and is a known natural target for fibrates (via PPAR α) in humans, and G6PDH is involved in fatty acid synthesis, steroid biosynthesis and defence against xenobiotic toxicity and oxidative stress.

LPL is a glycoprotein, and one of the family of triacylglycerol lipases, which include pancreatic and hepatic lipases. As previously stated, lipoprotein lipase is a key enzyme in the metabolism of lipids (Murthy *et al* 1996). In mammals, the enzyme is synthesised in the parenchymal cells of adipose and several other tissues (excluding the liver), and is secreted into the capillary endothelium, where it is bound to the cell surface and where the lipase acts on triacylglycerols in plasma lipoproteins, to release fatty acids for uptake and

storage or use by other organs such a muscle, for oxidation (Liang *et al*, 2002).

In humans, LPL activity is regulated by diet and a number of hormones, including epinephrine, norepinephrine, thyroid stimulating hormone, adrenocorticotropic hormone (ACTH), growth hormone and glucagons. Liang *et al* (2002) have also shown in the red sea bream, LPL expression seems to also be regulated by the nutritional state of the fish.

Babin and Vernier, (1989) have previously purified trout LPL, and found its properties and molecular weight to be similar to those of the mammalian enzyme. Kwon *et al* (2001) have also previously cloned and sequenced the LPL gene from Rainbow trout. In fish, significant LPL expression occurs in the liver (Oku *et al*, 2002; Liang *et al*, 2002) (unlike the situation in mammals) and also in vitellogenic ovaries (Kwon *et al*, 2001). It has been suggested by Kwon *et al*, (2001) that LPL is involved in the storage and mobilisation of egg yolk constituents during vitellogenesis. The tissue specific mechanisms of LPL gene expression in fish may therefore be different from that in mammals and consequently the literature regarding LPL may not be completely comparable across species.

Peroxisome proliferators-activated receptors (PPARs), as previously mentioned in Chapter 1, were identified in the 1990s in rodents and named after their ability to facilitate peroxisome proliferation. PPARs are key transcriptional factors that catalyze and coordinate different biochemical events in order to achieve energy homeostasis. Fibrates are well known activators of PPARs, specifically PPAR α , for the regulation of β -oxidation of fatty acids. They have been shown to control the expression of various genes that are crucial for lipid and glucose metabolism. PPAR α activation decreases triglyceride levels by amplifying the expression of lipoprotein lipase (LPL).

Glucose-6-phosphate dehydrogenase (G6PDH), the key enzyme of the pentose phosphate pathway (previously discussed in Section 1.6.1.1), is crucial in the supply of riboses for DNA and RNA synthesis by production of 6-phosphogluconate, particularly in proliferating cells. However, a more important function, the production of NADPH, the major cytoplasmic reducing compound, is now acknowledged (Winzer *et al*, 2002). G6PDH is a regulatory enzyme in NADPH-dependant xenobiotic biotransformation and defences against oxidative stress (Figure 6.1).

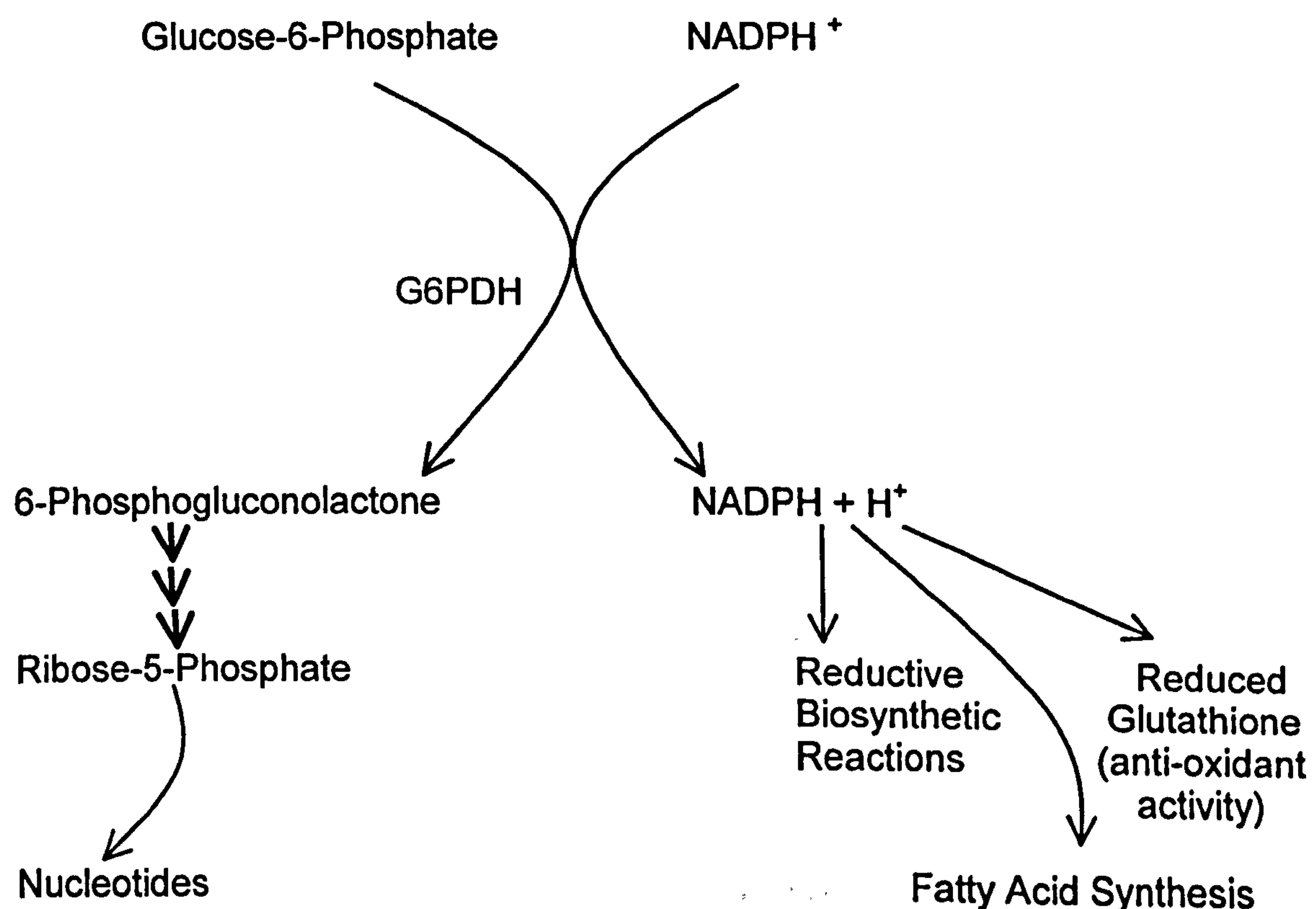


Figure 6.1: The role of glucose-6-phosphate dehydrogenase (G6PDH) in cellular metabolism (adapted from Salati and Amir-Ahmady, 2001).

G6PDH expression is regulated in the liver and adipose tissue by hormonal and nutritional factors. Its activity correlates with the rate of fatty acid biosynthesis. Winzer *et al* (2002) suggest that G6PDH in fish is sensitive to inactivation by xenobiotics in a sex-dependant manner. Physiological

conditions resulting in oxidative stress can also result in changes in G6PDH activity, and its gene expression is essential for protection of the cell against even mild oxidative stress.

6.2 Protocol

The effect of clofibric acid on both LPL and G6DPH mRNA expression levels was investigated. The livers of adult female and male fathead minnow collected from Experiments 4 and 5 (see Chapter 5 for experimental design) were used in this study. Morphometric and biochemical results from these experiments can also be seen in Chapter 5.

6.2.1 Computer analysis of known sequences

Gene sequence searches for fathead minnow LPL, G6PDH and β -actin were carried out using the Entrez search and retrieval system at NCBI (<http://www.ncbi.nlm.nih.gov/entrez/>). If the specific fathead minnow sequences were not found, sequences from related fish species that were present in the database were aligned using a web-based multiple alignment tool (BCM search launcher - <http://searchlauncherbcm.tmc.edu/multi-align/>) so that primers could be designed to the conserved regions (it was important that similar taxa were used in the alignment to ensure a more accurate consensus was achieved). Alignments can be seen in Appendix 2 and 3.

6.2.2 Primers construction

Primers were designed for the lipoprotein lipase and β -actin genes (chosen as a house-keeping gene) by aligning sequences from other species of fish and finding conserved regions. The primers were designed by eye due to

the complexity of the alignments. For LPL, primers were designed for RT-PCR across introns (sequence can be seen in Appendix 2), whereas for β -actin, the only suitable primer binding sites were found within exon 2, so the primers were designed to this region (sequence can be seen in Appendix 3). Although exonic primers potentially lack the specificity of primers designed across introns (as they can also amplify from the genomic gene sequence), genomic contamination was previously shown not to be an issue (see Section 2.3.1.2).

The general rules of primer design were followed. In some primers, when the identity of a base was unclear in the consensus sequence of the alignment, a degenerate base was used to allow for this.

Rules for primer design:

- Keep the G-C content about 50%
- Ensure there is a GC clamp at the 3' end
- Try and design primers between 16-24 base pairs long
- Avoid complementary sequence, both within and between primer pairs (to avoid secondary structure formation)

The primers for the G6PDH gene were designed using PRIMER3 program (part of the Lasergene suite of genetic analysis software) using a submitted sequence for the fathead minnow liver G6PDH gene (Gen Bank Accession number: AF206637), which can be seen in Appendix 1. All primers were then synthesised and purchased from Sigma – Genosys Ltd (Pampisford, UK). They were assessed by PCR-amplification of fathead cDNA at 1, 2 and 3 mM MgCl₂ concentrations (screening cycling conditions: 95°C for 15 mins, 35 cycles of 94°C for 30 sec, 58°C for 30 secs, 72°C for 30 secs, and a final 72°C for 5 mins). Primers that failed to amplify were disregarded. Remaining primers

were then screened by amplifying from a fathead mRNA dilution series on the Rotor Gene (Corbett Research, Cams). It was found that the LPL primers were problematic, giving amplification reaction efficiencies greater than 1. It was thought that this might be due to the degeneracy of certain bases within the primer sequence. These primers were then re-synthesised without degeneracy and these were then screened as before. All reaction efficiencies using these non-degenerative primers were less than 1. The optimal primer pair for each gene can be seen in Table 6.1 below.

Table 6.1: Primer pairs used for Real Time RT-PCR for liver samples from Experiments 4 and 5.

Gene	Primer sequence
LPL	FHLPL4r – AAC AGG TGG ATG GAG CGC FHLPL7f – GAC ATC TAC CCC AAT GGA GG
G6PDH	FHGAP3r – ACC CCA TCC CAG CGT TCA TTC FHGAP3f – GAG AAG CCC GCA TCC ACC AG
β-Actin	FHBACT2f – GAT ATG GAG AAG ATC TGG C FHBACT2r – GTT GGC TTT GGG GTT CAG G

6.2.3 Quantification of LPL and G6PDH gene expression by quantitative real time RT-PCR

Livers from control and dosed fathead minnows (*Pimephales promelas*) from both Experiment 4 (exposed to 1mg/l clofibric acid) and Experiment 5 (control, 10µg/l and 1mg/l clofibric acid) were carefully dissected and weighed using RNase free conditions and immediately snap-frozen as previously described in Section 2.1.4. Total RNA extraction was performed on the liver using TRI Reagent, as previously described in Section 2.3.1. The RNA concentrations were determined spectrophotometrically (Genequant,

Pharmacia, Cambridge, UK) and the integrity of the RNA checked by running it on agarose gels (Section 2.3.1.3). After concentrations were determined, all samples were then diluted with double autoclaved MilliQ water to give concentrations of 10ng/ μ l total RNA in each sample.

Real time RT-PCR was used for relative quantitative determination of LPL, G6DPH and β -actin mRNA expression in adult fathead minnow exposed to clofibric acid (control and dosed), as briefly discussed in Section 2. β -actin was used as a house-keeping gene to allow all other results to be 'normalised' to the level of expression of β -actin.

6.2.3.1 Experiment 4

130 total RNA samples extracted from livers dissected from fathead minnows exposed to clofibric acid in Experiment 4 (Section 2.1.4 and 5.4) were diluted to 10ng/ μ l using double autoclaved, sterile deionised water.

10 μ l reactions were run in strips of 4x100 μ l PCR tubes, where 10ng of RNA was used as a template in each reaction with master mix (Qiagen QuantiTect SyBR Green RT-PCR master mix - consisting of HotStarTaq DNA polymerase, Quantitect SYBR Green RT-PCR buffer (containing TrisCl, KCl, (NH₄)₂SO₄, and 5mM MgCl₂, pH 8.7 (20°C), dNTP Mix (contains dATP, dCTP, dGTP, and dTTP/dUTP; ultrapure), SYBR Green and ROX fluorescent dyes, QuantiTect RT mix) and RNase free water (PCR grade)) and 5 picomols of each of the primers. All samples were analysed in duplicate

RNA sample 100 was used as the standard for this experiment and so was diluted using double autoclaved, sterile deionised water to 0.5, 2, 8, 20, 50 ng/ μ l (Figure 2.14). These dilutions were used to produce the standard curve for quantification of all 3 genes.

Amplification and detection of samples and standards were performed on the Rotor Gene 3000 system (Corbett Research, Cambs) using the following thermal cycling conditions:

50°C for 30 minutes	(Reverse transcription)
95°C for 15 minutes	(Taq activation/initial denaturing step)
94°C for 10 sec	(denature)
58°C for 10 sec	(annealing)
72°C for 10 sec	(extension)

} repeated for 50 cycles

At the end of this cycling, a melt curve was calculated which consisted of measurements of fluorescence at every 1°C from 60-99°C. This curve shows the specificity of the PCR and reveals the presence, if any, of primer-dimer (Figure 2.15).

Quantities of mRNA in the samples were quantified relatively by comparison with a standard curve derived from a dilution series of an RNA reference (sample 100). The level of β-actin in each reaction was used to normalise for the variability in RNA quality or quantity, among the samples. Results are reported as the ratio of LPL/β-actin and G6DPH/β-actin.

6.2.3.2 Experiment 5

84 total RNA samples previously extracted from livers dissected from fathead minnows exposed to clofibrate acid in Experiment 5 (Section 2.1.4 and 5.5) were diluted to 10ng/μl using double autoclaved, sterile deionised water as described previously. In this experiment, however, PCR reactions were carried in 96-well plates in 20μl reactions using the same constituent concentrations as specified previously (Section 6.2.3.1) and again, all samples were analysed in

duplicate. RNA sample 100 from Experiment 4 was again used as the standard for this experiment and so was diluted using double autoclaved, sterile deionised water to give the range 0.19, 0.78, 3.125, 12.5, 50, 200 ng/ μ l. Amplification and detection of samples and standards were performed on the ABI Prism 7900HT Sequence detection system (Applied Biosystems, USA) using thermal cycling conditions as described above. Quantities of mRNA in each of the samples were calculated as described previously.

6.2.4 Statistics

Differences between males and females within control and clofibrate acid-exposed experimental groups were analysed using SigmaStat (version 2.03). In cases of normality, differences were determined by a t-test, and where normality test failed, non-parametric statistical tests were used.

6.3 Results

6.3.1 Levels of expression of selected genes in the livers of fish from Experiment 4

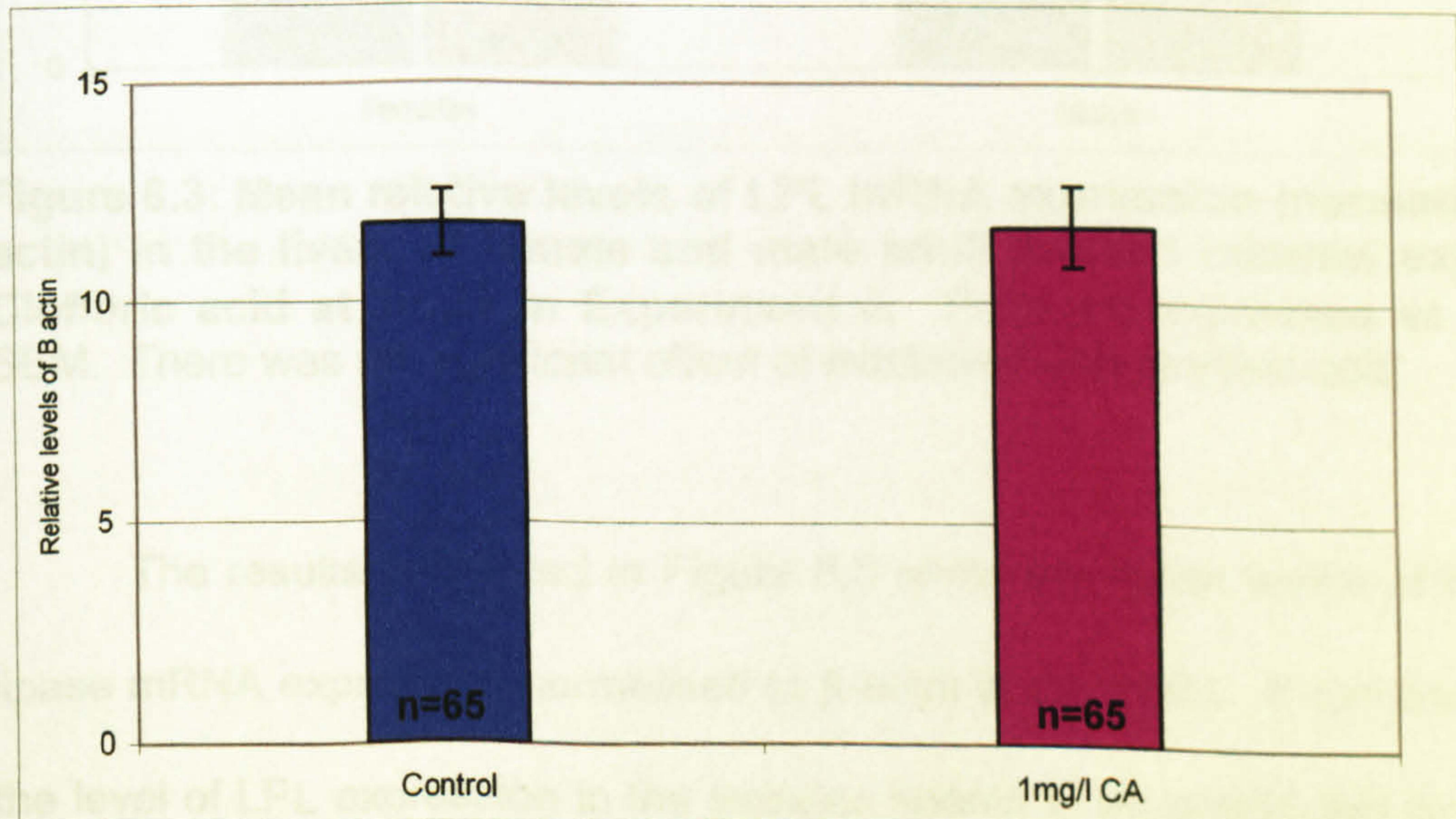


Figure 6.2: Mean relative levels of β -Actin expression in the livers of adult fathead minnow exposed to 1mg/l Clofibrate acid from Experiment 4 ($p=0.3$). Data are expressed as means \pm SEM. There was no significant effects of treatment with clofibrate acid.

Figure 6.2 shows the levels of β -actin mRNA expression from the livers collected from Experiment 4. It has been shown that gene expression of β -actin (and many other house-keeping genes) may change in the face of physiological challenges (Kazeto *et al*, 2004), thereby negating its usefulness in standardising for loading differences, RNA quality, etc. However, in this experiment, β -actin levels amongst control and dosed fish were shown to be very similar, and were well within the expected individual and procedural variance. Therefore, I feel confident that β -actin expression, at least in this study, was a useful and accurate internal standard for normalisation of the data.

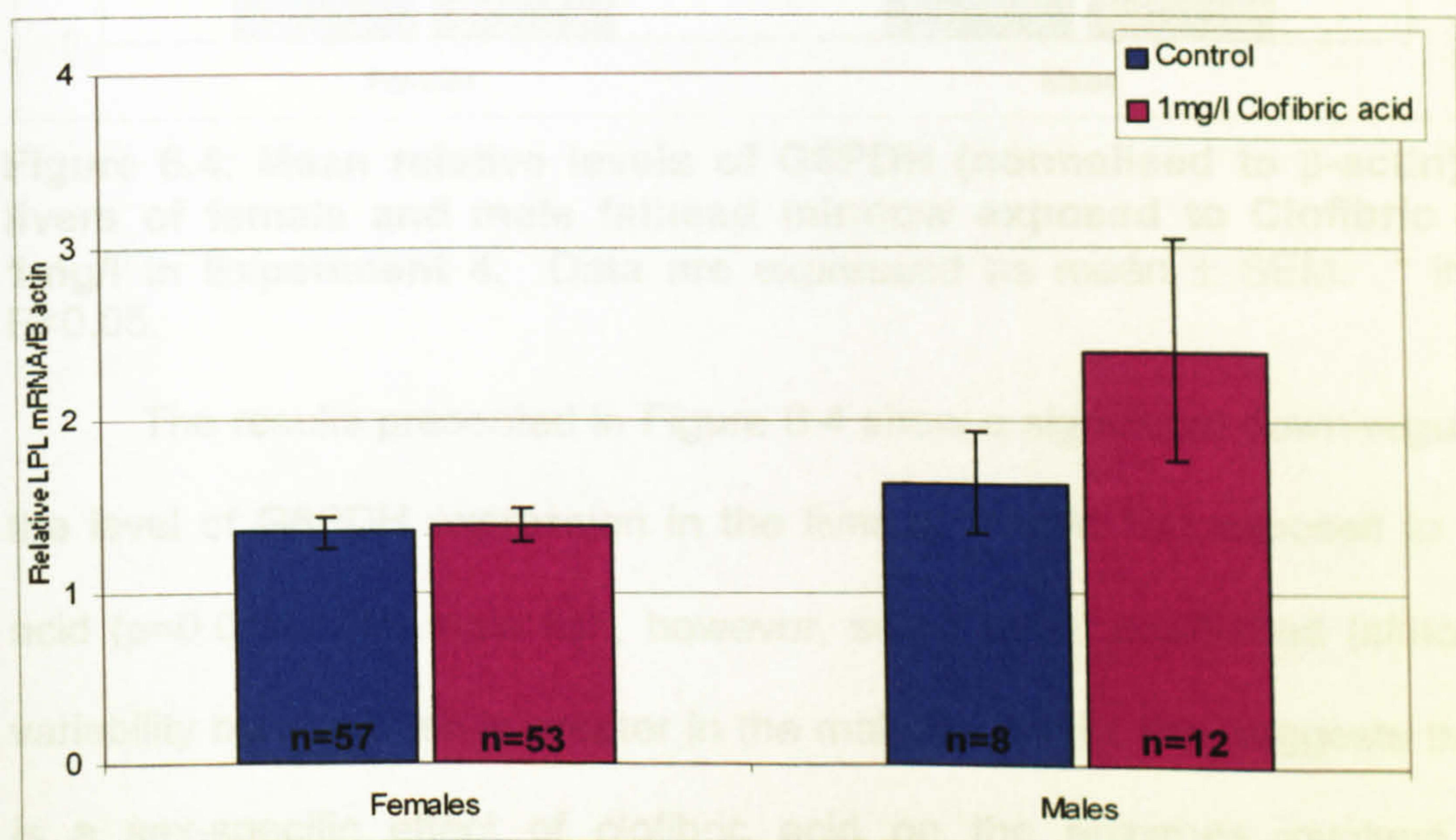


Figure 6.3: Mean relative levels of LPL mRNA expression (normalised to β -actin) in the livers of female and male adult fathead minnow exposed to Clofibrate acid at 1mg/l in Experiment 4. Data are expressed as means \pm SEM. There was no significant effect of treatment with clofibrate acid.

The results presented in Figure 6.3 show the mean levels of lipoprotein lipase mRNA expression normalised to β -actin in the livers. It can be seen that the level of LPL expression in the females seems to be unaffected by exposure to clofibrate acid, whereas in the males the levels give the impression that they

are up-regulated (although not significantly). These results indicate that there may be a sex-specific effect on LPL expression after exposure to clofibric acid.

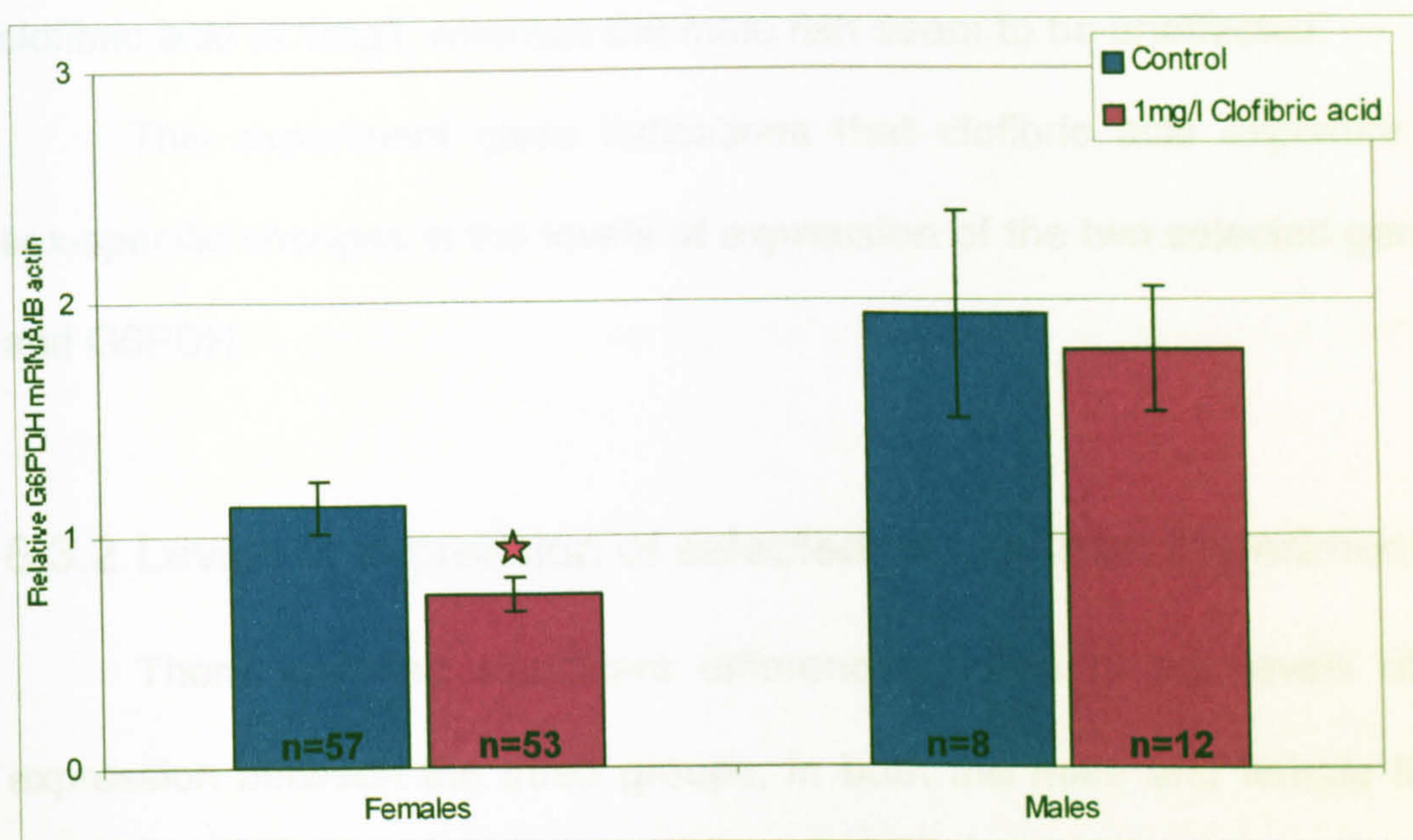


Figure 6.4: Mean relative levels of G6PDH (normalised to β -actin) in the livers of female and male fathead minnow exposed to Clofibric acid at 1mg/l in Experiment 4. Data are expressed as mean \pm SEM. * indicates $P<0.05$.

The results presented in Figure 6.4 show a significant down-regulation in the level of G6PDH expression in the liver of female fish exposed to clofibric acid ($p=0.01$). The male fish, however, seem to be unaffected (although the variability between fish is greater in the males). Again, this suggests that there is a sex-specific effect of clofibric acid on the enzymes involved in lipid metabolism in fathead minnow, with the effect in this case on female fish.

6.3.1.1 Summary of the results from Experiment 4

In summary, the results obtained from Experiment 4 indicate that the levels of LPL mRNA expression in fathead minnow exposed to 1mg/l clofibric acid are affected in a sex-specific manner. In the males, the levels of LPL mRNA seem to show an upward trend in response to exposure, whereas the female fish seem to be unaffected by this same treatment. The results obtained

for G6PDH expression, on the other hand, tell a different story: the female fish have significantly lower levels of G6PDH mRNA expression when exposed to clofibric acid at 1mg/l, whereas the male fish seem to be unaffected.

This experiment gives indications that clofibric acid exposure causes sex-specific changes in the levels of expression of the two selected genes, LPL and G6PDH.

6.3.2 Levels of expression of selected genes from Experiment 5

There were no significant differences found in the levels of β -actin expression between the three groups, in both the male and female fish, from this study (data not shown). This agrees with the results from Experiment 4, and therefore shows that β -actin is suitable as a house-keeping gene in these studies.

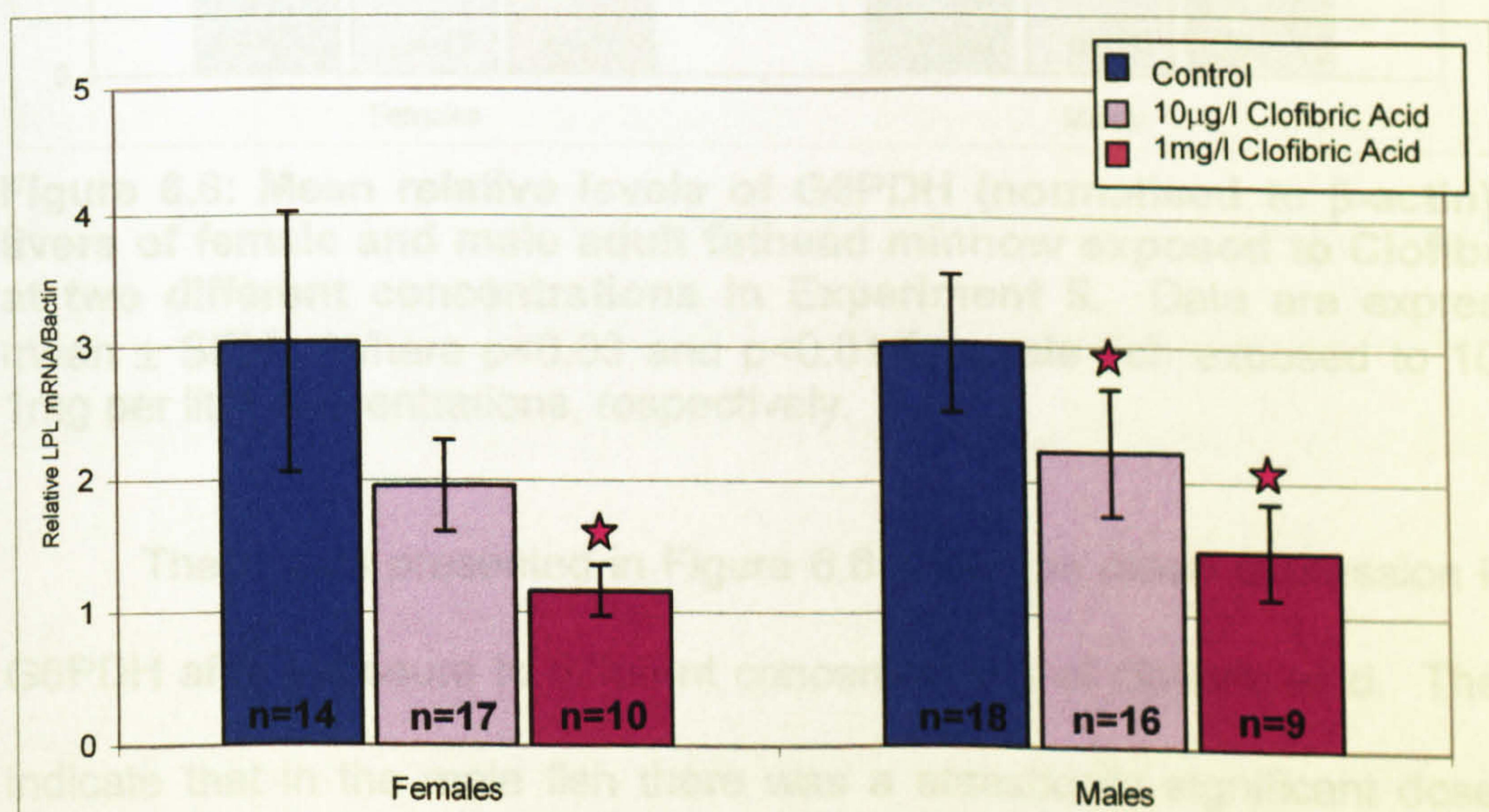


Figure 6.5: Mean relative levels of LPL mRNA expression (normalised to β -actin) in the livers of female and male adult fathead minnow exposed to Clofibric acid at two different concentrations in Experiment 5. Data are expressed as mean \pm SEM. Significant differences are shown, where p<0.03 (males) and p<0.01 (males and females) for the 10 μ g and 1 mg per litre concentrations, respectively.

The results presented in Figure 6.5 show that there seems to be a dose-dependant decrease on the mean LPL expression levels of both male and female fish in response to clofibric acid exposure for 3 weeks. In the male fish at both doses of clofibric acid, significant differences were seen when compared to the controls ($p<0.03$ and 0.01). In the case of the females, there was a significant difference between the control and highest dose only of clofibric acid ($p<0.01$).

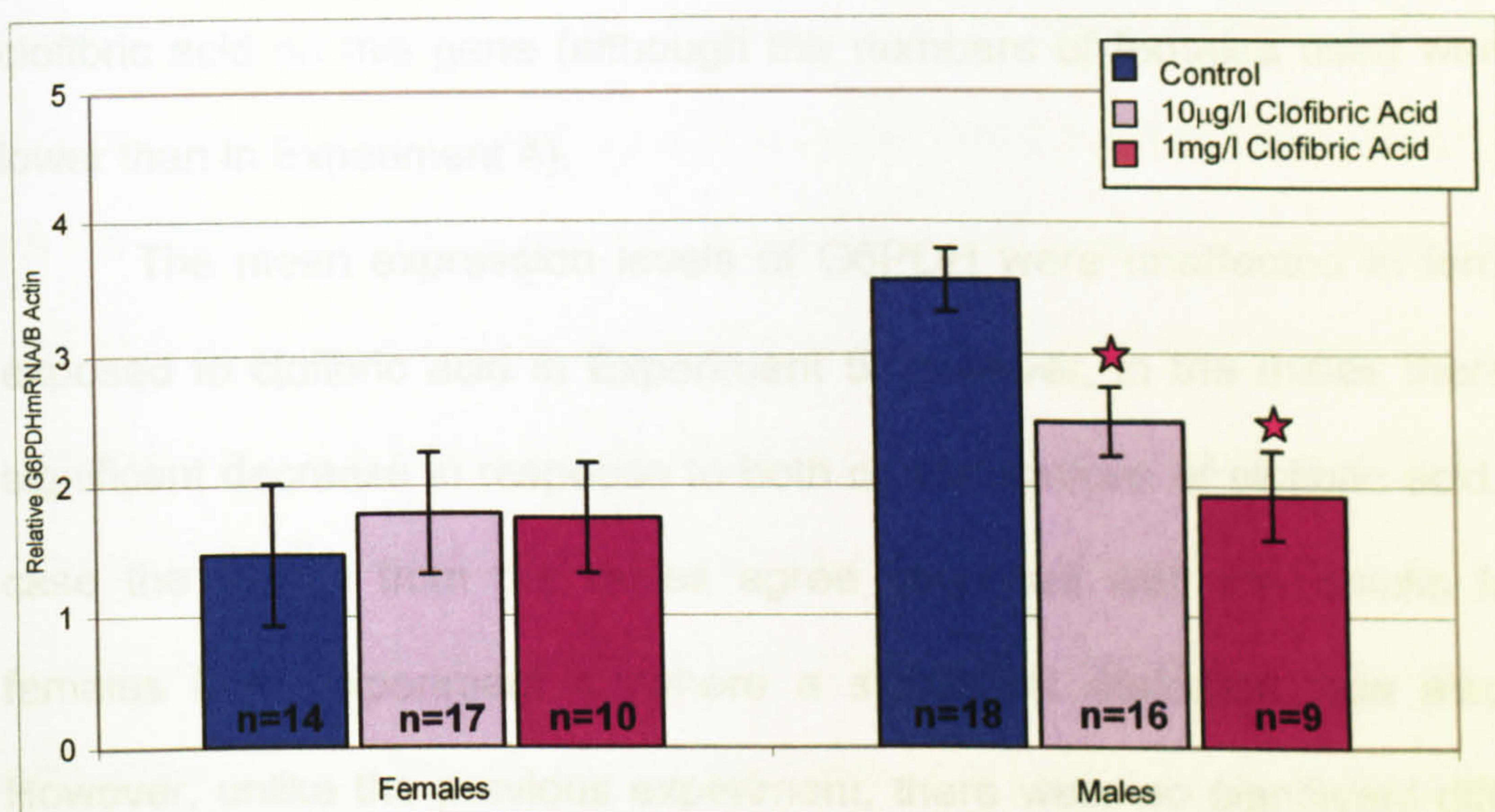


Figure 6.6: Mean relative levels of G6PDH (normalised to β -actin) in the livers of female and male adult fathead minnow exposed to Clofibric acid at two different concentrations in Experiment 5. Data are expressed as mean \pm SEM. Where $p<0.03$ and $p<0.01$ for male fish exposed to $10\mu\text{g}$ and 1mg per litre concentrations, respectively.

The results presented in Figure 6.6 show the mean expression levels of G6PDH after exposure to different concentrations of clofibric acid. The results indicate that in the male fish there was a statistically significant dose-related decrease in mean G6PDH levels with increasing exposure to clofibric acid. In the females, however, no difference was seen with exposure to clofibric acid.

6.3.2.1 Summary of the results from Experiment 5

In summary, the results obtained in Experiment 5 indicate that clofibric acid had significant effects of the levels of expression of LPL in both males and females (with less effect at lower concentrations in the females). These results differ from the results from Experiment 4, where a possible up-regulation was seen in the males and no effect was seen in the females. However, this up-regulation was not significant, and the number of males used in Experiment 5 were higher and therefore may give be a better representation of the effect of clofibric acid on this gene (although the numbers of females used were much lower than in Experiment 4).

The mean expression levels of G6PDH were unaffected in female fish exposed to clofibric acid in Experiment 5, however, in the males there was a significant decrease in response to both concentrations of clofibric acid. In this case the results from the males agree very well with the results from the females from Experiment 4 (where a significant decrease was also seen). However, unlike the previous experiment, there were no significant differences seen in the expression levels in females (although the numbers of fish were lower). The apparent difference in the response of the males may be seen due to the fact that the numbers of fish Experiment 5 were higher than in Experiment 4.

6.4 General summary of the effects of clofibric acid on gene expression in the fathead minnow

The results presented in Chapter 6 indicate there are definite effects of clofibric acid exposure on levels of gene expression in the fathead minnow. It can be seen, however, that there are differences in the results between the two

experiments, although they were both carried out in the same way. The only obvious differences between experiments were in numbers of fish used and the sex ratio within each tank (~5:1 females: males in Experiment 4 and 1:1 in Experiment 5). The importance of this is discussed below.

The LPL results in Experiment 4 indicate that the levels of expression were not significantly different in the males or females exposed to clofibric acid, however, there was an indication in the males of a slight upward trend. On the other hand, in Experiment 5, there were significant and dose-dependant decreases in the males and females exposed to clofibric acid. It is considered that the results from the female fish in Experiment 4 are probably more reliable, as the numbers were higher and therefore there was less variability, whereas with the males, the numbers were lower and may be less reliable than those obtained in Experiment 5. Additionally, these differences in effect on LPL expression between experiments may be explained by the sex ratio of the fish in each experiment. In Experiment 5, there were higher numbers of male fish within each tank and consequently more likelihood of the female fish being encouraged to spawn more regularly, therefore inducing more reproductive cycles and consequently more fluctuations in the vitellogenin production/oogenesis. It has previously been shown that the levels of LPL expression change in the ovary during oogenesis (Kwon *et al*, 2001) and this may be true for LPL of the liver also; this could account for the relatively high variability between fish seen within Experiment 5.

Fibrates have been previously shown to raise LPL expression in rats (Staels and Auwerx, 1992) and female mice (Toda *et al*, 2003), although no induction has been shown in rabbits (Alegret *et al*, 1998), and in fact a decrease in LPL expression has been noted previously in hamster hepatocytes after

exposure to a fibric acid (Guo *et al*, 2001). Alegret *et al* (1998) suggest that the differences seen may be explained by inter-species differences in peroxisome proliferation.

It was noted in both experiments that the levels of expression of G6PDH were higher in males than females, and this agrees with the results of Winzer *et al* (2002), who found that male fish generally show higher G6DPH activity than females. In Experiment 4, the levels of G6PDH were significantly down-regulated in female fish and seemed to be unaffected in male fish exposed to clofibric acid at 1mg/l. However, the results obtained in Experiment 5 show a significant decrease (which was dose-dependant) in the expression levels of G6PDH of male fish, and the female fish in this case seem to be unaffected. Again, as with the LPL results, the numbers of females used in Experiment 4 were higher and consequently results may therefore be more of a true representation of the effects of clofibric acid on the female fish. The males, however, have higher numbers in Experiment 5, and these results also agree with the significant results found in the female and the slight decrease seen in the males from Experiment 4.

G6PDH activity has previously been shown in male rats exposed to clofibric acid to be decreased (Cleary *et al*, 1987), however, they do state that there have been no consistent effects previously seen on G6PDH activity. Winzer *et al* (2002) showed a decrease in the level of G6PDH in female flounder hepatocytes with exposure to xenobiotics, and Bucher, *et al*, (1993) also observed a reduction in G6PDH activity in bullheads exposed to paper mill effluent (however sex was not specified). G6PDH in female flounder has been previously shown to be inhibited by their levels of 17-β-oestradiol, although the mechanisms of action are not clearly identified yet (Winzer *et al*, 2002), and this

inhibitory effect of oestradiol is probably the reason that G6PDH levels are on average lower in females compared to males. It may also explain why no effects were seen in Experiment 5, while a significant effect was seen in Experiment 4: the fish were more likely to be spawning with the presence of more males, which might lead to higher, and more fluctuating 17- β -oestradiol levels. This would lead to more variable G6PDH levels, which could potentially 'mask' any effects of drug exposure in this experiment.

A reduction in G6PDH level increases an organism's susceptibility to xenobiotics, and this may result in the decreased production of NADPH, which is necessary for biosynthetic reactions or biotransformation activities (Winzer *et al*, 2002). This makes females far more susceptible to xenotoxicity, as they also need NADPH for the production of VTG.

In conclusion, it can be seen from these experiments that significant effects of clofibric acid, at different concentrations, occurred on the levels of expression of selected genes. Generally, the effect of clofibric acid seems to be to reduce the expression of LPL and G6PDH in the liver of exposed fish. However, these were not completely reproduced in each experiment, and possible reasons for this have been discussed. The results nevertheless highlight the need for further investigation, taking into account sex ratios, spawning and E2 levels. It may also be informative to look at tissue-specific effects of clofibric acid exposure, as only levels in the liver were measured in these experiments.

Chapter 7: General Discussion

It is now well established that pharmaceuticals and synthetic chemicals from personal care products are being released into the environment. The exact effects that each drug has (both individually and in combination) on ecosystems, biota, and humans, however, are not understood. It is therefore clear that much more research is critically needed.

Prior to the work reported this thesis, there were no known experiments (excluding basic toxicological studies) carried out to look at the effects of clofibric acid or clofibrate on lipid metabolism in fish.

In 2001, Pfluger and Dietrich asked a pertinent question, as to whether clofibric acid can adversely affect cholesterol synthesis, and thus steroidogenesis, and consequently influence endocrine regulation, in aquatic species. Given the complexity of lipid metabolism and its pathways, there may potentially be various mechanisms that could make fish susceptible to exposure to clofibric acid. The results obtained in this thesis suggest that clofibric acid may have complex effects, reflecting several different mechanisms of action at various stages of lipid metabolism. Effects of fibrates on mammals have been seen from the physiological level down to the molecular levels. The chapters within this thesis have presented the first evidence (albeit preliminary) of clofibric acid having effects on both adult and embryo fish, at various levels of organisation.

The studies I conducted involved two different species, and for one of these species, fish of different ages. None of the results I obtained (perhaps with the exception of those concerned with spermatogenesis and some of those involving gene expression) showed major, very pronounced effects of clofibric acid. Further, although some of the effects I observed might have been what I expected if the drug had the same effects (through the same mechanisms) in

fish as it does in mammals, these effects were not always reproducible, raising a question mark as to their robustness. However, my research does show that there are effects of clofibrate acid in fish on pathways which were not only previously unexpected (for example, steroidogenesis and spermatogenesis), but which may potentially have a much larger impact on the viability of fish populations. These effects all occurred at concentrations of clofibrate acid below those previously shown to have any biological effects on fish.

It may be the case, however, that the seemingly minor effects I found were not statistically significant purely because of the sensitivity of methodologies employed for determining these relatively small differences, rather than them not actually being due to real effects. Consider, for example, the effects of clofibrate acid on lipid metabolism in mammals. The effects on lipid levels in humans (a reduction of the concentration of triglycerides by 30-60% and cholesterol by 20-25%) are changes seen in people who have elevated levels in the first place. It is this group of people who take fibrates. Investigating these effects in a different organism, with different levels and route of exposure, and in one whose lipid levels are not elevated above normal by disease, is a completely different scenario - (although it has been shown that in hamsters with normal lipid levels there are significant decreases in both cholesterol and triglyceride levels (Guo *et al* 2001)). In fish, which have been shown to be hyperlipidemic and hypercholesterolemic (when compared with human standards), the effect of clofibrate acid on triglyceride and cholesterol levels may be lower than in humans, and as such may be close to or below the sensitivity of the assays employed.

Another point to consider is whether these minor effects of clofibrate acid on fish are genuinely minor, or is it possible that the doses I used produced

effects that are just at the bottom of the dose response curve, in which case increasing the dose would increase the magnitude of those effects. And if so, are the higher doses environmentally relevant? Further work is needed to assess these possibilities.

The most significant results from this thesis must be put into context of any relevant previous literature. The reduction in sperm production and steroid synthesis have shown that fish exposed to clofibrate acid are deleteriously effected. Similarly to this, studies previously have reported reduced libido and impotence in humans (Martindale, 1999), perhaps due to a reduction in sex steroids, as well as arrested spermatogenesis in dogs and monkeys (Schulman *et al*, 2002). Phthalates (also PPs) have also been shown to induce pathological changes in rodents, and other fibrates have been shown to alter gene expression in Leydig cells of rats, and inhibit progesterone synthesis in human cell lines (Gazouli, *et al*, 2002). In mice Bezafibrate has also been shown to exhibit antiandrogenic properties (Gazouli, *et al*, 2002). The fact that the same results were found in fish and mammals, and even that the potential mechanism (the reduction in testosterone) is the same for fish and mammals, provides much stronger evidence that these are true effects in fish. However, the more expected and well-characterised therapeutic effects that Clofibrate acid has in humans and mammals (namely the raising of HDL/reduction in triglyceride and cholesterol, etc) have not been found in this study. The failure to reproduce these effects in fish may highlight true biological differences between fish and mammals, or may be purely due to differences in routes of entry.

Upon reflection, what would I do differently if I were to start my PhD now? I would probably start my studies at the molecular level, using a larger range of concentrations for the exposure experiments – to determine if there were dose-related effects of exposure to clofibric acid. I would also probably more clearly establish the base-line values of the parameters of interest, and their inter-individual variability, before exposing any fish to the drug of interest. For example, more knowledge of the normal levels of cholesterol, triglyceride and lipoproteins in fish, and to what degree and why these change, would have helped a lot in both experimental design and the interpretation of results. More established techniques to measure these parameters would have also been very useful. It is important, however, to remember that assessment of the effects of specific drugs is complicated by their mechanisms of action – which show species-, tissue-, and sometimes cell-specific effects. Consequently, those effects seen in one species may not be seen in others.

It is always very easy to say in hindsight what one would have done differently – however, I believe that the choices I made at the time were correct. They have provided thought-provoking results that can, in the future, be used to take this piece of science further.

I have shown significant effects on a number of very important endpoints, in both adults and embryos, during fixed-term exposures. There is no way of estimating by how much these changes might be magnified over time and successive generations, or what cumulative effect they may have over the entire lifecycle, and consequently what effect they will have on overall fitness. It may be that pronounced effects occur only at high (environmentally unrealistic) concentrations of clofibric acid, but nevertheless much lower (environmentally

realistic) concentrations may still produce significant effects which, although small, could still be important (and adverse).

It is interesting that work designed to ascertain the impact of environmental levels of clofibric acid on fish has highlighted the need to reconsider the impact of these same levels on man (particularly as the drug has been shown to be present in drinking water): what effect will it be having on our young men's sperm counts, and on unborn embryos? Although such statements may seem unnecessarily alarmist to some, others would not dismiss the possibility that low level, but constant, exposure of people might affect some of them (especially in the light of the continued decline of human sperm counts).

In summary, some effects of clofibric acid on fish have been determined, although further, more in-depth studies are required to determine if these effects are reproducible. Multigenerational studies would ideally be carried out, to determine if clofibric acid has any population effects. The fact that effects have been observed when clofibric acid concentrations were well below the previously determined NOEC and LOEC levels, indicates that populations in the wild could well be affected. Long-term exposure to seemingly low concentrations may be very harmful, causing sublethal effects manifested in later generations.

It is important, also, to consider that it may not be exposure to an individual drug which causes a population decline – aquatic organisms are probably never exposed to just one drug (such as clofibric acid), but instead receive simultaneous exposure to a poorly defined, complex mixture of many drugs. I investigated the effects of just one drug, but in the future I hope research will lead to the determination of the effects of the cocktail of drugs fish that are exposed to. These drugs in combination with even more complex

multiple stressors such as disease, predation or competition and habitat destruction, need to be considered, to assess if these individual contributors add to the overall stress a population can be under.

It is clear from the results of the embryo experiment that further work must be carried out to further characterise and quantify the effects that were reported here and ascertain if the effects could be detrimental at the population level. It would be sensible to examine where exactly clofibrate acid is located when it is taken up by the egg/embryo. This could be done by using radio-labelled (or dye-conjugated) clofibrate acid. Another interesting parameter to look at, in light of the differences in yolk opacity observed, would be the changes (if any) in the lipid profile within the developing embryo yolk; this could be done using gels (similar to the ones used in the adult experiments). Using histological techniques, the development of the spleen could also be examined (i.e. examining changes in sinusoid space, as was done by Handy *et al*, 2002), or by determining white blood cell counts, to see if there is any effect on the immune response with exposure to clofibrate acid.

Perhaps the most important endpoints for further study, in view of the results obtained, would be to look at the effects of clofibrate acid on gene expression in embryos after exposure to clofibrate acid, specifically looking at genes (especially G6PDH) involved in lipid metabolism (for example, acyl CoA, APO E, and A1) and the detoxification pathways (e.g. cytochrome p450 genes). Changes in gene expression in the embryo would be expected to have pronounced effects, as they will be occurring during development, when the embryo is extremely sensitive and gene expression is consequently tightly regulated.

The results from the adult studies also highlight the need for further investigation. It would be interesting to ascertain the reproductive effects of the

lowering of the testosterone and sperm counts, and consequently a pair breeding experiment would be useful to determine if there were any effects on fertility, by measuring the numbers of eggs spawned/fertilised. The sperm quality could be further assessed by measuring sperm motility and viability using the Computer Assisted Sperm Analysis (CASA) system (Hobson Sperm tracker; Kime *et al*, 2001). Additionally, gonads of both sexes could also be examined histological to determine if the decreased sperm count is due to any gross effects on cellular organisation (such as abnormal development of Leydig, Sertoli and germ cells, for example). The effect of clofibric acid on the next generation could also be assessed – by looking at juveniles, which where spawned, hatched, and ‘grown on’ in the presence of clofibric acid.

It could also be very important to look in more detail at the molecular level. As well as looking at other genes involved in lipid metabolism (see above), it might be very informative to also look at the genes involved in spermatogenesis and steroid metabolism. An extremely powerful way of doing this would be to use DNA microarrays to quantify changes in gene expression using RNA extracted from exposed versus non-exposed fish. In this way, changes in expression of thousands of genes, including whole biochemical pathways, can be assessed at the same time, allowing multiple mechanisms in many different pathways to be identified rapidly and efficiently.

It is true also that the experiments described in this thesis were conducted somewhat unrealistically (in an environmental sense), since only a single compound was assessed. In reality, wildlife are often exposed to multiple concentrations of many pollutants with multiple mechanisms of action. Therefore, once reproducible effects have been established, mixture effects could be assessed.

With respect to pharmaceuticals in the environment, it is now clear that the most important conclusion is that an improvement in the clearance of these drugs from the environment is essential. Although costly, removal/degradation of these contaminants biologically or chemically during wastewater treatment is required, to reduce their concentrations in effluents discharged into the aquatic environment. This, in turn would reduce the long-term and largely unknown cost not only to wildlife, but also ourselves.

Personal Publications

Although none of the research reported in this thesis has currently been published as papers, I intend to write a paper or two, based on the most convincing results, in the near future.

PAPERS

Harries, J.E., **Runnalls, T**, Hill, E., Harris, C.A., Maddix, S., Sumpter, J.P., Tyler, C.R. (2000) Development of a reproductive performance test for endocrine disrupting chemicals using pair-breeding fathead minnows (*Pimephales promelas*). *Environmental Science and Technology*, 34: 3003-3011.

Handy, R.D, **Runnalls, T**, Russell, P.M. (2002) Histopathologic biomarkers in three-spined sticklebacks, *Gasterosteus aculeatus*, from several rivers in Southern England that meet the freshwater fisheries directive. *Ecotoxicology*, 11(6):467-79.

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Brian, J.V., Harris, C., Scholze, M., Backhaus, T., Booy, P., Lamoree, M., Pojana G., Jonkers, N., **Runnalls, T.**, Bonfa, A., Marcomini, A. and Sumpter, J.P. (2005) Prediction Of The Response Of Freshwater Fish To A Mixture Of Estrogenic Chemicals. *Environmental Health Perspectives*, 113 (6): 721-728.

MEDIA INTEREST IN MY RESEARCH

Steve Farrar (2004) Cast-off drugs harm wildlife. *The Times Higher Education Supplement*. 21st May 2004.

POSTERS

Runnalls, T.J. and Sumpter J.P., (2004) Pharmaceuticals in the Environment – the effect of Clofibrate acid on fish. SETAC Europe, 14th Annual Meeting, Prague, Czech Republic.

Brian, J.V., Harris, C., Scholze, M., Backhaus, T., Booy, P., Lamoree, M., Pojana, G., Bonfa, A., **Runnalls, T.**, Marcomini, A., Sumpter, J. (2004) Assessing the joint action of estrogenic chemicals in freshwater fish. SETAC Europe, 14th Annual Meeting, Prague, Czech Republic.

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Van Aerle, R., Runnalls, T.J., and Tyler, C.R. (2000) Gonadal Sex differentiation in fathead minnow, *Pimephales promelas*. 4th International Symposium on Fish Endocrinology, Seattle, Washington, USA.

Tyler, C.R., Hutchinson, T., Harries, J., Runnalls, T., Pounds, N., Maddix, S., van Aerle, R., Harris, C., and Sumpter, J.P. (1999) Reproductive effects of environmental estrogens in the fathead minnow (*Pimephales promelas*). SETAC 20th Annual meeting, Philadelphia, PA.



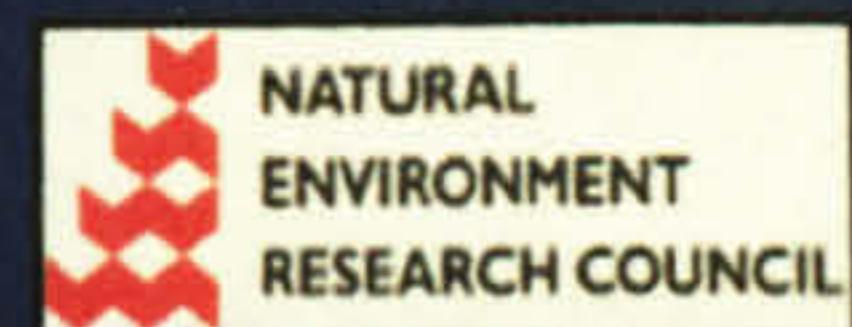
PHARMACEUTICALS IN THE ENVIRONMENT

- the effects of clofibric acid on fish

Tamsin Runnalls and John Sumpter

Ecotoxicology Research Group, Department of Biological Sciences, Brunel University, Uxbridge,
Middlesex, UB8 3PH, UK

Tamsin.Runnalls@Brunel.ac.uk



Abstract

Pharmaceuticals in the aquatic environment are an emerging issue and the risks they pose are mostly unknown. We have been studying possible effects of clofibric acid, the major metabolite of a lipid-lowering drug, in fish. These were investigated using different experimental approaches and endpoints. When fathead minnow embryos were exposed to clofibric acid, effects were seen on eggshell, time to hatch, hatchability, mortality and viability. Adult fathead minnow were similarly exposed and, in females, significant effects on liver size and HDL cholesterol were observed. The action of clofibric acid on gene expression was also examined: Glucose-6-Phosphate Dehydrogenase (G6PDH) mRNA was significantly down-regulated in females, whilst Lipoprotein Lipase (LPL) mRNA appeared to be up-regulated in males. Vitellogenin (VTG) gene expression decreased, whereas eggshell protein (ZP3) mRNA levels remained unchanged. Clofibric acid also had a significant effect on sperm count in exposed fish; this may be, in part, due to a decrease in plasma testosterone concentration (also reduced in exposed fish). These results show that clofibric acid may have detrimental effects on fish populations.

Introduction

Various classes of prescription drugs have been detected at concentrations up to $\mu\text{g/l}$ levels in sewage, surface and ground water. They include steroids, anti-inflammatory, anti-cancer, lipid regulators, anti-epileptic, antibiotics and painkillers. Little is currently known about the environmental impact of these drugs. As they are developed to have a biological effect, their potential to cause effects within the environment is great.

Clofibric acid (CA) is the major metabolite of the lipid-lowering drugs Clofibrate, Etofibrate and Etofyllinclofibrate. It is non-biodegradable, very persistent and frequently found at $\mu\text{g/l}$ concentrations in the environment. CA was originally detected in the environment as a consequence of the routine analysis for acidic pesticides, and has since been frequently detected in surface waters, ground waters, sewage effluent and even drinking waters.

Materials and Methods

Exposure studies were carried out using both adult and embryo fathead minnow (figure 1) to determine the effects of CA.

➤ **Adults**: various assays were employed to determine potential effects of CA (nominal concentrations: 1mg/l and 10 $\mu\text{g/l}$):

- Serum was run on agarose gels (HYDRAGEL HDL/LDL cholesterol kit - SEBIA) and visualised using a cholesterol specific enzymatic method. Quantification of HDL cholesterol within lipoproteins was achieved using densitometry
- Livers and gonads from exposed and control female fish were used to determine VTG and ZP3 mRNA levels respectively, using Hybridisation Protection Assays (MLT, Cardiff)
- Sperm counts were carried out on male gonads
- Plasma was analysed for levels of testosterone using a radioimmunoassay
- Quantitative real-time RT-PCR was performed on livers samples using Qiagen QuantiTect SYBR green RT-PCR kit to compare expression levels of LPL and G6PDH mRNAs between control and dosed fish

➤ **Embryos**: various end points were monitored including time to hatch, hatchability, mortality and viability when exposed to 1mg/l, 1 $\mu\text{g/l}$ and 1 ng/l CA.

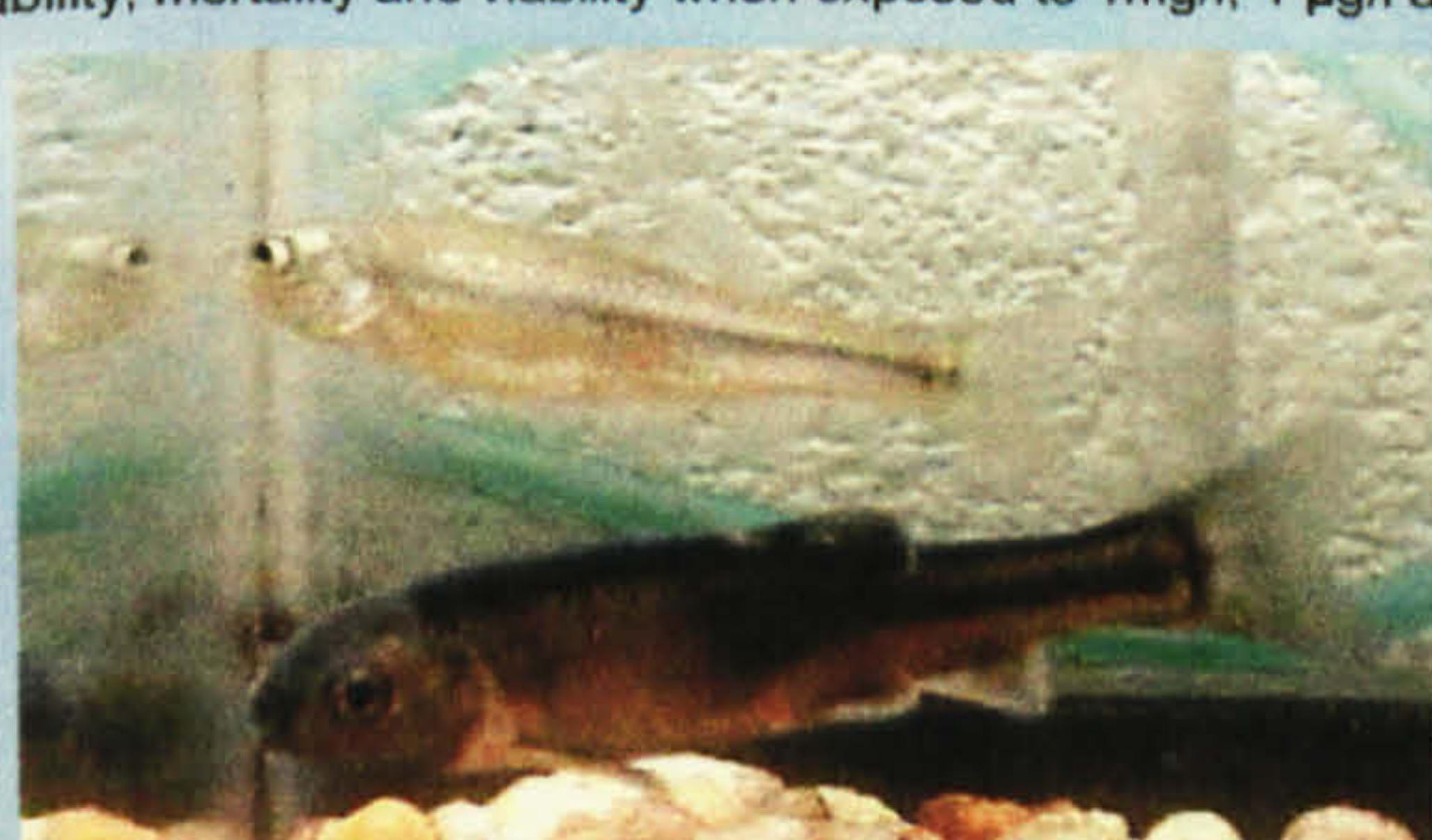


Figure 1: The fathead minnow (*Pimephales promelas*): males are the larger and darker fish, whereas the females are smaller and lighter

Results and Discussion

➤ **Adults:**

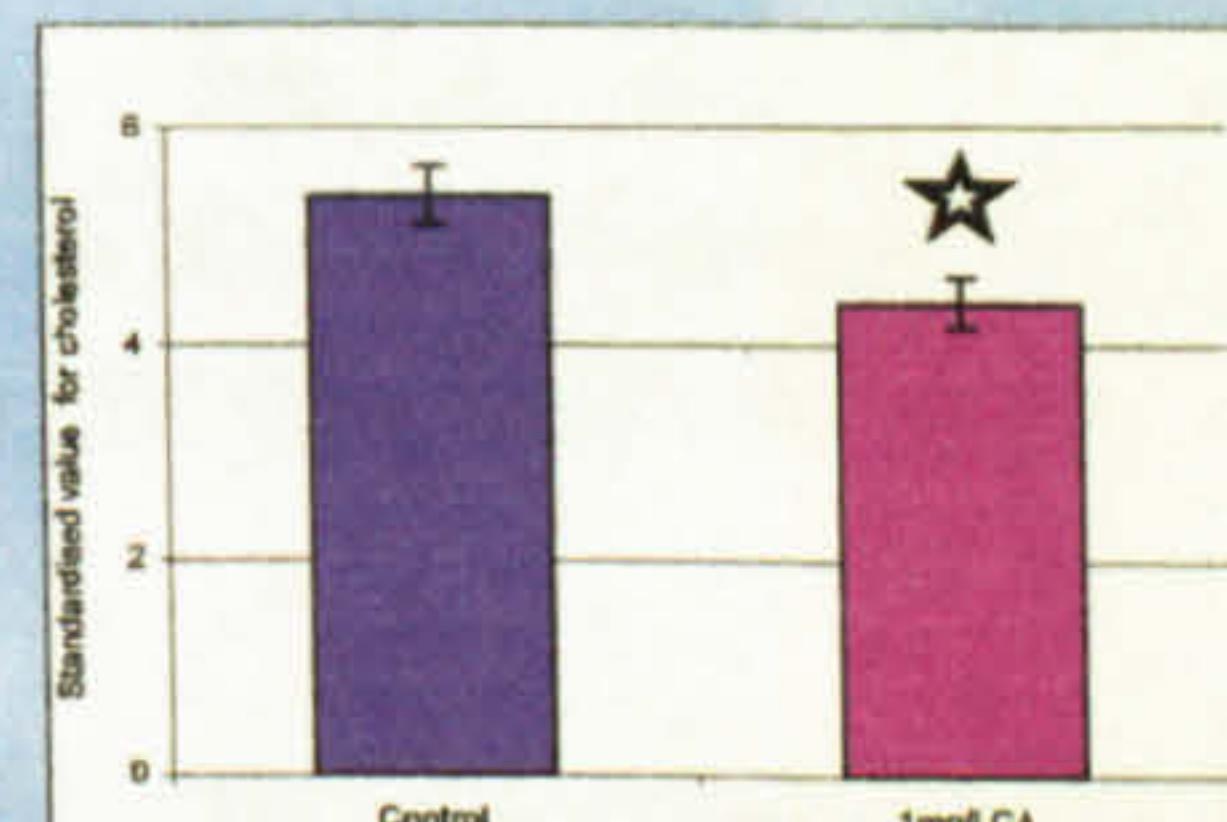


Figure 2: Effect of clofibric acid on HDL cholesterol levels in females ($p=0.006$)

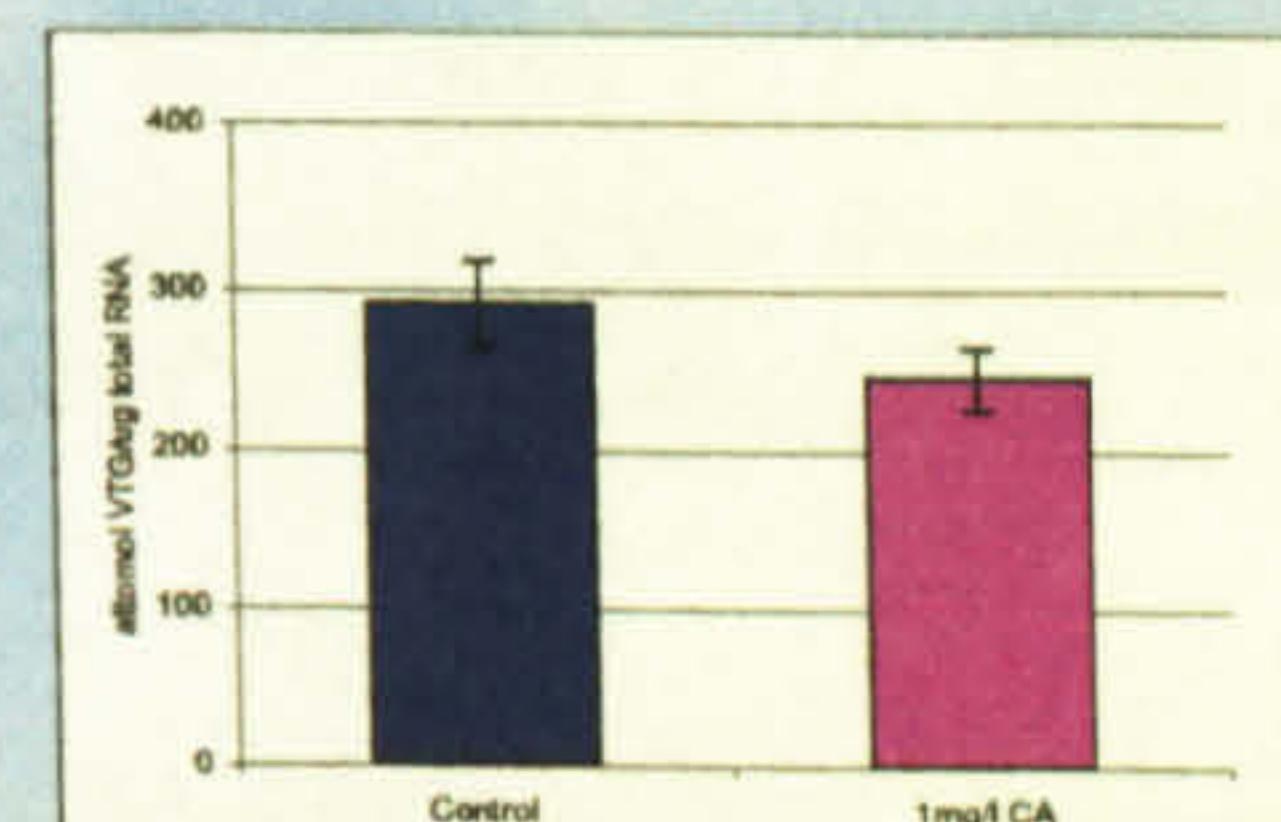


Figure 3: Effect of clofibric acid on VTG mRNA levels in females

- Significant effects on HDL cholesterol in lipoproteins were observed in females (figure 2)
- Vitellogenin (VTG) gene expression in females decreased (figure 3), although not significantly, whereas eggshell protein (ZP3) mRNA levels in females remained unchanged (data not shown)

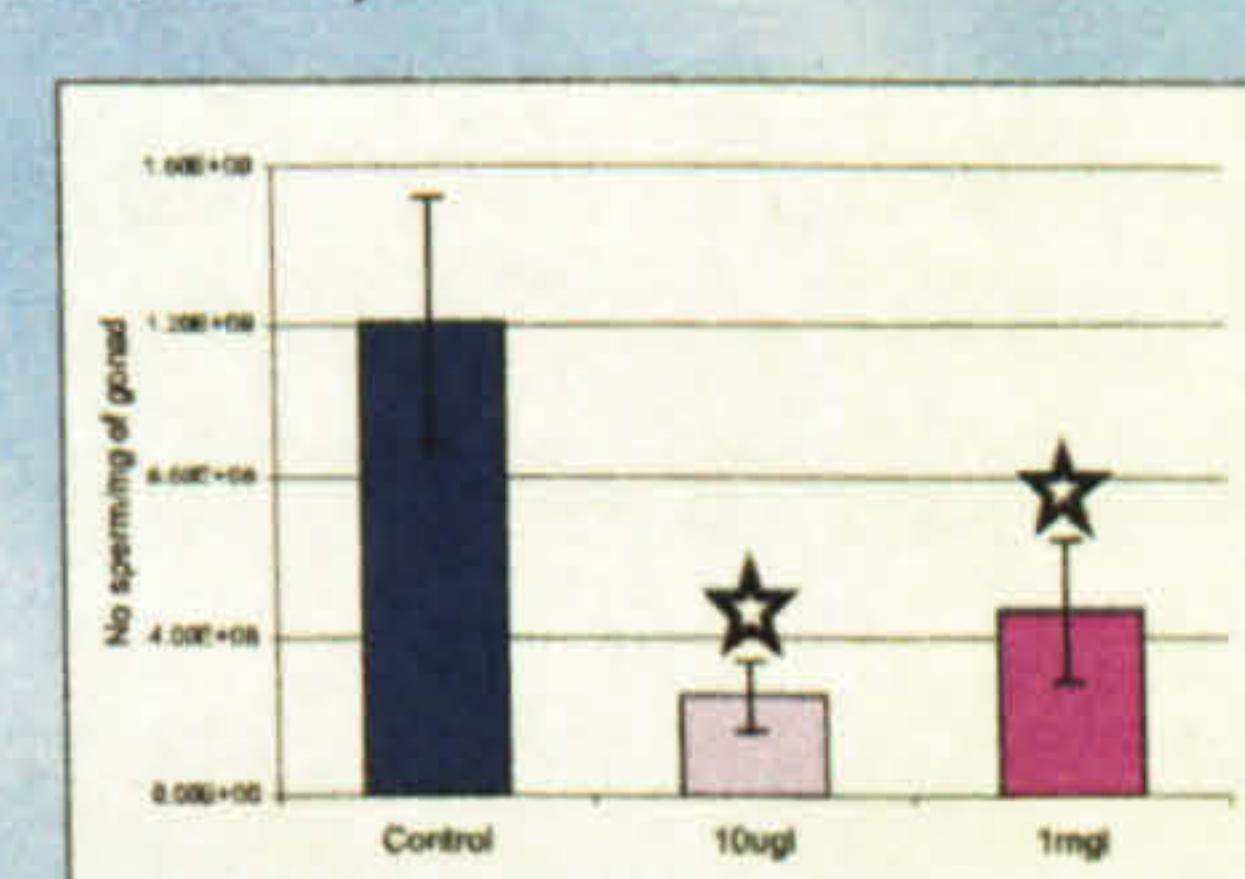


Figure 4: Effect of clofibric acid on sperm counts of males ($p=0.01$ and 0.04)

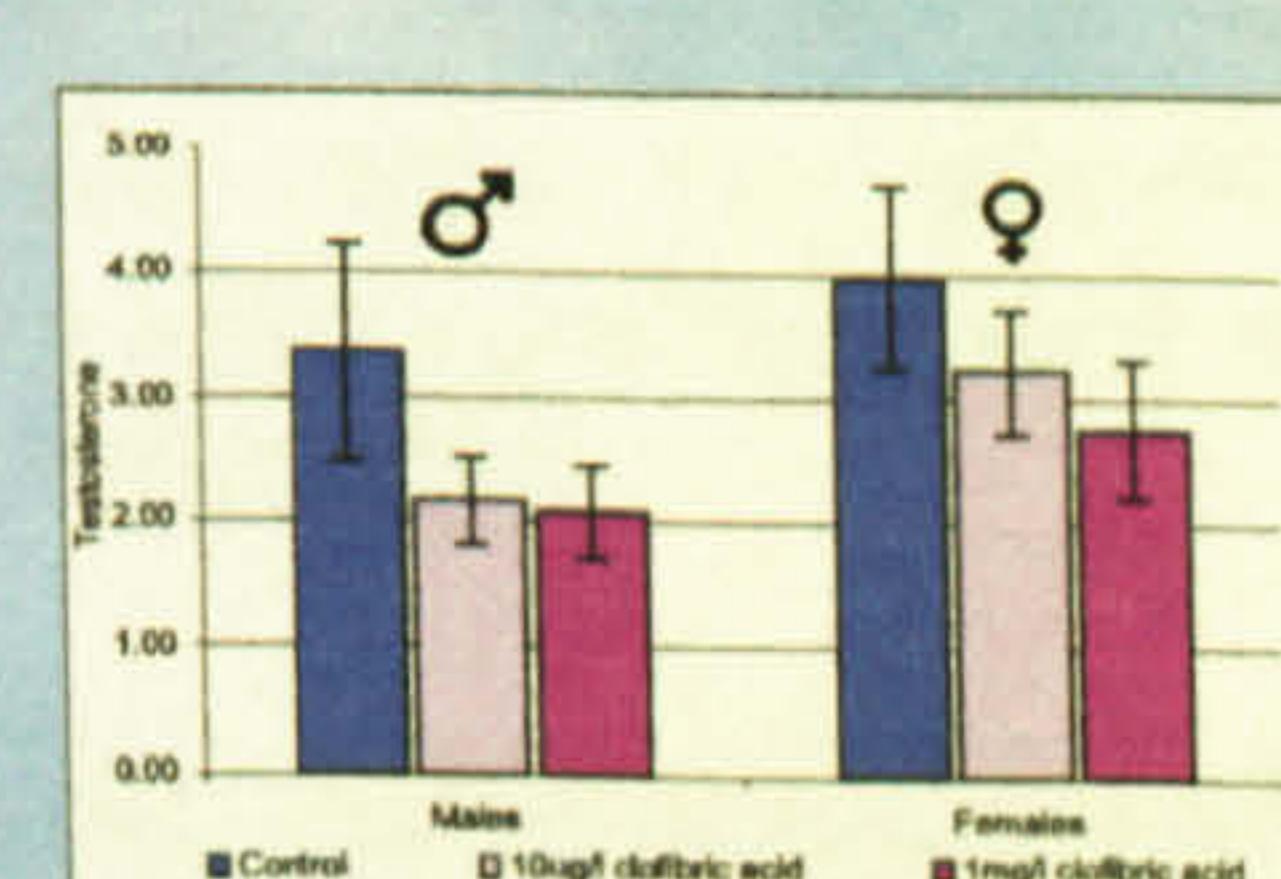


Figure 5: Effect of clofibric acid on plasma testosterone levels in males and females

- Significant effects on sperm count in exposed fish were seen (figure 4); this may be, at least in part, due to a decrease in plasma testosterone concentration in dosed fish (figure 5)

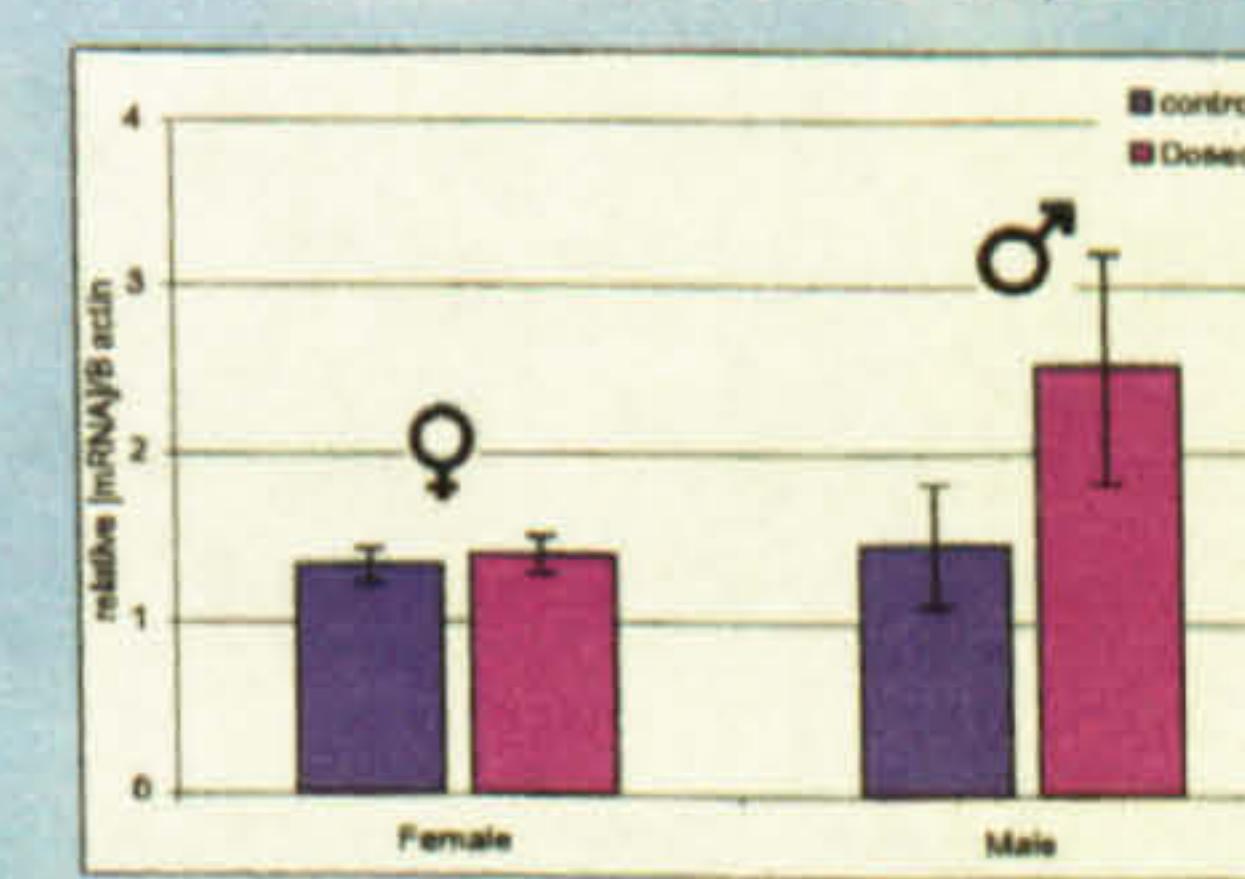


Figure 6: Effect of 1mg/l clofibric acid on LPL mRNA levels

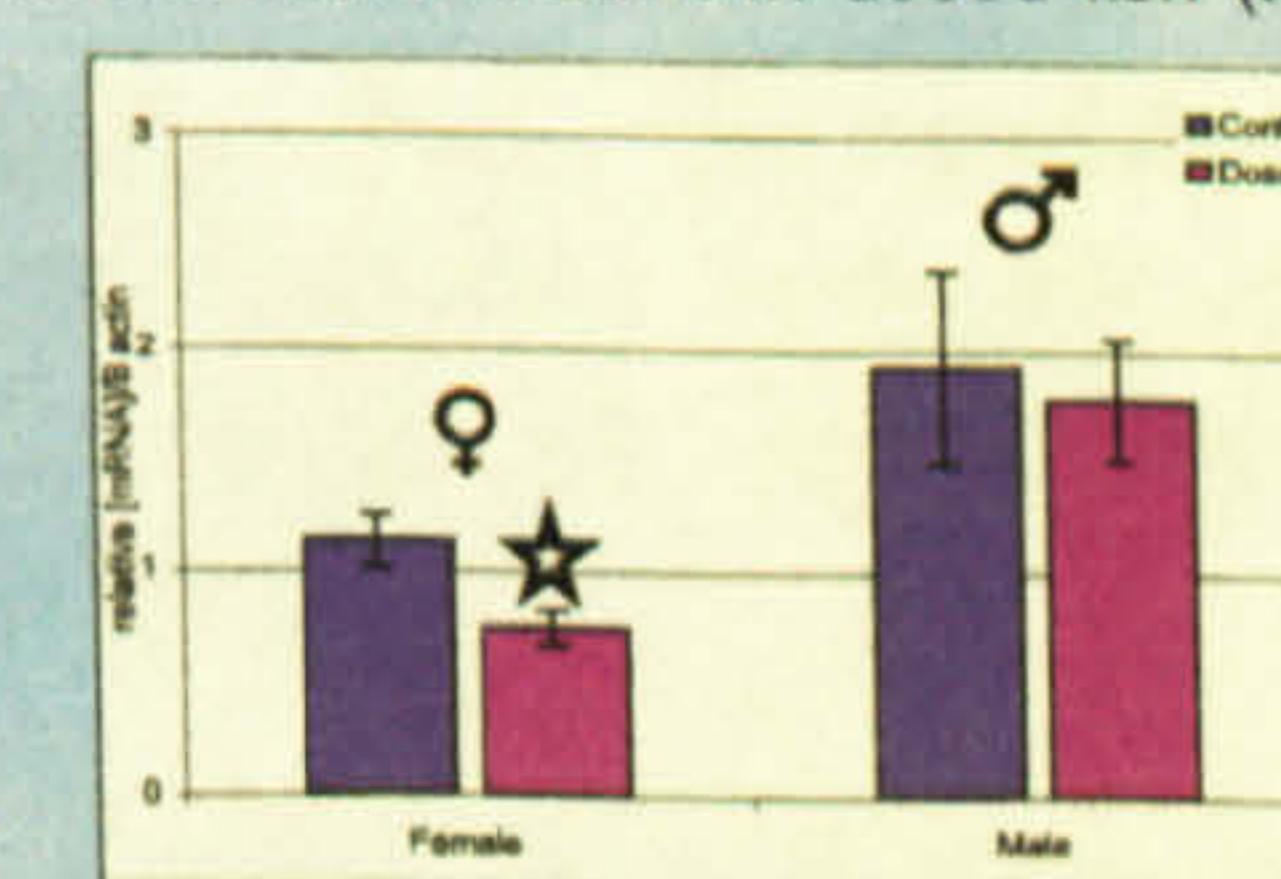


Figure 7: Effect of 1mg/l clofibric acid on G6PDH mRNA levels ($p=0.01$)

- Effects on gene expression: LPL mRNA appeared to be up-regulated in males (figure 6), whilst G6PDH mRNA was significantly down-regulated in females (figure 7)

- Significant effects were seen on HSI in females (data not shown)

➤ **Embryos:**

- Effects were also observed on egg shell, time to hatch, hatchability, mortality and viability (data not shown)

As can be seen, significant effects have been found at relatively low doses (further work needs to be carried out to establish the LOECs for these endpoints)

Conclusions

- Although these are preliminary findings, they indicate a trend, and perhaps a sex-specific effect in some cases. The presence of clofibric acid in the aquatic environment may have effects not only at a physiological level but also at the molecular level.
- In mammals, Clofibrate is known to reduce elevated plasma concentrations of triglycerides by 30-60% and, to a lesser extent, cholesterol (by 20-25%). It has been found to do this by acting through PPAR receptors to alter expression of a number of genes - we have quantified the degree of expression of two of these genes (Lipoprotein Lipase and Glucose-6-Phosphate Dehydrogenase) in fathead minnow and have shown that this mechanism of action may also pertain to fish.
- This work shows that clofibric acid has the potential to have significant effects on all stages of fish development, the cumulative effect of which may have a profound effect on fish fitness when chronically exposed in the natural environment.

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APPENDIX

Appendix 1: Sequence used for designing primers for G6PDH real-time PCR in the PRIMER3 program.

LOCUS AF206637 **1412 bp mRNA linear VRT 20-APR-2000**
DEFINITION *Pimephales promelas* liver glucose-6-phosphate-1-dehydrogenase mRNA, partial cds.
ACCESSION AF206637
VERSION AF206637.2 GI:7629274
SOURCE *Pimephales promelas*
ORGANISM *Pimephales promelas* Eukaryota; Metazoa; Chordata; Craniata; Vertebrata; Euteleostomi; Actinopterygii; Neopterygii; Teleostei; Ostariophysi; Cypriniformes; Cyprinidae; *Pimephales*.
REFERENCE 1 (bases 1 to 1412)
AUTHORS Dasmahapatra,A.K., Wimpee,B.A. and Dellinger,J.
TITLE G6PD mRNA in fathead minnow liver
JOURNAL Unpublished
REFERENCE 2 (bases 1 to 516)
AUTHORS Dasmahapatra,A.K., Wimpee,B.A.B. and Dellinger,J.
TITLE Direct Submission
JOURNAL Submitted (18-NOV-1999) School of Allied Health, NIEHS Marine and Freshwater Biomedical Sciences Center, University of Wisconsin-Milwaukee, 600 East Greenfield Avenue, Milwaukee, WI 53204, USA
REFERENCE 3 (bases 1 to 1412)
AUTHORS Dasmahapatra,A.K., Wimpee,B.A.B. and Dellinger,J.
TITLE Direct Submission
JOURNAL Submitted (20-APR-2000) NIEHS Marine and Freshwater Biomedical Sciences Center, University of Wisconsin-Milwaukee, 600 East Greenfield Avenue, Milwaukee, WI 53204, USA
REMARK Sequence update by submitter
COMMENT On Apr 20, 2000 this sequence version replaced gi:[6601558](#).
FEATURES
source Location/Qualifiers
1..1412
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/mol_type="mRNA"
/db_xref="taxon:90988"
/tissue_type="liver"
CDS
<1..>1412
/note="G6PD"
/codon_start=1
/product="glucose-6-phosphate-1-dehydrogenase"
/protein_id="[AAF19030.2](#)"
/db_xref="GI:7629275"

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IVYHDVTKNIKHQCMSTKGWNRIIVEKPFHDLQSSEELSSHLSLFTTEEQIYRIDHY
LGKEMVQNLMLVLRFGNRIFGPIWNRDSVACVVLTFKEPFGTQGRGGYFDDFGIIRDVM
QNHLQLMQLSLVAMEKPASTSSNDVRDEKVVLKCIEAVSLSDVVLGQYVGDPDGEGDA
KLGYLDDSTVPKGSTQATFATAVLYVKNERWDGVPFILRCGKALNERKAEVRLQFTDV
PGDIFDSQCRRNELVVRVQPNEAIYAKMMSKKPGVYFSPEETELDLTYHSRYRDVKLP
DAYERLILDVFCGQMHFVRSDELREAWRIFTPLLHQIEKEKTPPIKYKGSRGP
AEAD
ELVQKVGFRYEGTYKWVNPH"

ORIGIN
1 ctggccaaga agaaaaatcta cccaaactttg tggtgggtgt tcagagacgg tcttctcccc
61 gaacagacat attcgtggg ctttgctcg tccgatctga ccgtggatgc catacgatcg
121 gcctgcatgc cttacatgaa ggtggatcgac tctgatgcag agccgcctcg cgcgttcttc
181 agtcgaaact cttacatcg ggaaagtac gtggatgaat cgtcttcga taacctgaac
241 actcacctgc tgtccctgcc cgaggaggcc gggccaacc ggctttcta cctgc
301 **ccgeccatcg tctatcatga** tgtcaccaag aacatcaaac atcaatgcat gagcacc
361 **gcttggaca** ggatcatcg ggagaagccg tttggccatg acttgcagag ctcagaggag
421 ttatccagtc atctgttttc tctcttcacc gaggaacaga tttaccggat agaccattac
481 ctaggcaaag agatggtcca gaacctaatg gtgctcagat **tttggaaaccg** gatatttggc
FHGAP2r 541 cccatatggc accggggacag tgtggcatgt **gtgggtctga** ctttcaaaga gcccgg
FHGAP1f 601 **acccaggccc** qqqqaggata tttcgatgt **tttggtatca** ttcggatgt catgc
661 cacctgctgc agatgctctc tctgggtgcc **atggagaacgc** cccatccac cactcta
721 gatgttaggg acgagaagg **qaagggtttg** aagtgcattg aagcagtctc tctgtcagac
781 gtggcttgg gtcagttacgt gggagatcca gatggagaag gagatgctaa acttggttat
841 ctagatgact caactgtccc taaaggctcc actcaagcta catttgccac tgca
FHGAP2f
FHGAP3f

	FHGAP3r	FHGAP1rb
901	tatgtaaa <ins>ga</ins> atgaacgctg ggatgggg <ins>ttt</ins> cccttcatcc tccgggt <ins>gtgg</ins> aaaggctctg	
961	aatgagagga aggcagaggt gaggctgcag ttcaccgacg ttccctggaga catcttgac	
FHGAP1r	1021 tctcagtgc <ins>a</ins> ggcggAACGA <ins>gttggggc</ins> cgcgtgc <ins>agc</ins> ccaacgaggc catctacgcc	
1081	aagatgatga gcaagaaacc tggagttcac ttcagtcgg aagagacgga gctggacctg	
1141	acttaccaca gcagatacag <ins>ggat</ins> tttaag ctccctgatg cttatgagcg actgatttg	
1201	gatgtctttt gtggtcagat gcatttt <ins>gtt</ins> cgc <ins>atgt</ins> gacg aattgaggga agcctggaga	
1261	atcttcactc ctcttcttca tcagatcgag aaagagaaga cgccacccat caaatacaaa	
1321	tacgg <ins>agtc</ins> gtggcctgc agaggctgac gagctggtgc agaagg <ins>tggg</ins> ct <ins>ttcgctat</ins>	
1381	gagggtacat acaagtgggt caaccacac ag	

Where: **intron/exon boundies;** **reverse primers;** **forward primers;** **over lap between primers.**

Appendix 2: LPL sequence alignment using the ClustalW Multiple sequence alignment on the BCM search launcher.

Including bream genomic sequence did not give good alignment (compared to β -actin), possibly because of the large number of exons, or diversity between genes. Intron/exon boundaries, therefore, are from the submitted genomic sequence for bream (AB054063) and are marked on the bream cDNA sequence (|). Marked on the alignment below are the primers binding sites for the original degenerate primer pairs, non-degenerate primers were subsequently designed and screened from these.

1 rainbowtrout1	1	15 16	30 31	45 46	60 61	75 76	90	
2 rainbowtrout2		-----CCAC	TCTTATCACAAATCA	GATTACTTTAGAC	CAGAGCAGTGGTCCT	ACTGACCTATAACAAT	TGGATCTACCACTCC	79
3 zebrafish1		CATAAGAAA	CTCCAC	TCTTATCACAAATCA	GATTACTTTAGAC	CAGAGCAGTGGTCCT	ACTGACCTATAACAAT	90
4 zebrafish2		-----	-----	-----	-----	-----	-----	0
5 catfish		-----	-----	-----	-----	-----	-----	0
6 atlanticsalmon2		-----	-----	-----	-----	-----	-----	0
7 bream		-----GACGCACTGA	ACTCAGTCTGAGAGG	GAAATATACAGTCTT	CTGCAAAGAGATCCG	ACTGATTAAATTAAC	TGCTGAAGCGATTCT	85
8 atlanticsalmon1		-----	-----	-----	-----	-----	-----	0
9 winterflounder1		-----	-----	-----	-----	-----	-----	0
10 winterflounder2		-----	-----	-----	-----	-----	-----	0
1 rainbowtrout1	91	105 106	120 121	135 136	150 151	165 166	180	
2 rainbowtrout2		TAAACCAACATACAA	GAACATTTCATTCG	GGACTTGACGGAGAA	ACCTCTGGATCAACA	ATAATATGGAAAAG	AAAATACCTTTTGG	169
3 zebrafish1		TAAACCAACATACAA	GAACATTTCATTCG	GGACTTGACGGAGAA	ACCTCTGGATCAACA	ATAATATGGAAAAG	AAAATACCTTTTGG	180
4 zebrafish2		-----	-----	-----	-----	-----	-----	0
5 catfish		-----	-----	-----	-----	-----	-----	0
6 atlanticsalmon2		-----	-----	-----	-----	-----	-----	0
7 bream		TCTGTGAAAGGACA-	-AACCTGTGAA---	AGAAATGAAAGCGTG	GCGAGTTGTGTTCT	GTACTTTCTGGTGT	GAA-TGCAGTCGTGC	168
8 atlanticsalmon1		-----	-----	-----	-----	-----	-----	0
9 winterflounder1		-----	-----	-----	-----	-----	-----	0
10 winterflounder2		-----	-----	-----	-----	-----	-----	0
1 rainbowtrout1	181	195 196	210 211	225 226	240 241	255 256	270	
2 rainbowtrout2		TCACCGTTGGATAA	TTCTGGCAAATATCT	GTGTATCTTTCAA	GTACACCAGAACAAA	CACTTTTGGTAACA	GCAACTCCACTGAAT	259
3 zebrafish1		TCACCGTTGGATAA	TTCTGGCAAATATCT	GTGTATCTTTCAA	GTACACCAGAACAAA	CACTTTTGGTAACA	GCAACTCCACTGAAT	270
4 zebrafish2		-----	-----	-----	-----	-----	-----	0
5 catfish		-----	-----	-----	-----	-----	-----	0
6 atlanticsalmon2		-----	-----	-----	-----	-----	-----	0
7 bream		AGCATGTGACGTCCC	TGGAAGAAGAGCTCT	CCGATTCTATTTTG	[GTA---ACTTCCTC	GACCTCTGAAA---	--GACTTGATTGAAC	249
8 atlanticsalmon1		-----	-----	-----	-----	-----	-----	0
9 winterflounder1		-----	-----	-----	-----	-----	-----	0
10 winterflounder2		-----	-----	-----	-----	-----	-----	0
1 rainbowtrout1	271	285 286	300 301	315 316	330 331	345 346	360	
2 rainbowtrout2		GGCTTGAGGACTACA	CAGACATTGTATCCA	AGTTCTCCCTGAGAA	CTGCTGAGATACCGG	ATGATGACTTGTGCT	ACATCGTCCCGGCC	349
3 zebrafish1		GGCTTGAGGACTACA	CAGACATTGTATCCA	AGTTCTCCCTGAGAA	CTGCTGAGATACCGG	ATGATGACTTGTGCT	ACATCGTCCCGGCC	360
4 zebrafish2		-----	-----	-----	-----	-----	-----	0
5 catfish		-----	-----	-----	-----	-----	-----	0
6 atlanticsalmon2		-----	-----	-----	-----	-----	-----	0
7 bream		ACAAGGATGACGCCA	ATCAAACCGTTGCCA	AATTCTCCCTCCGGA	AACCGTCCCACCTCTG	ATGACGACCTGTGCT	ACATCGTCCCTGGCA	339
8 atlanticsalmon1		-----	-----	-----	-----	-----	-----	0
9 winterflounder1		-----	-----	-----	-----	-----	-----	0
10 winterflounder2		-----	-----	-----	-----	-----	-----	0
1 rainbowtrout1	361	375 376	390 391	405 406	420 421	435 436	450	
2 rainbowtrout2		AGCCCTCAACCATCC	CAAAGTGTGAATTCA	ACCCCTGGGTATAAGA	CGTTCTGGTCATT	ATGGATGGACGGTCA	CGGGGCTGTTTGAGA	439
3 zebrafish1		AGCCCTCAACCATCC	CAAAGTGTGAATTCA	ACCCCTGGGTATAAGA	CGTTCTGGTCATT	ATGGATGGACGGTCA	CGGGGCTGTTTGAGA	450
4 zebrafish2		-----	-----	-----	-----	-----	-----	0
5 catfish		-----	-----	-----	-----	-----	-----	0
6 atlanticsalmon2		-----	-----	-----	-----	-----	-----	0
7 bream		AACCTGACTCCCTGG	CAGCCTGCACCTTC	ACAGCTCCCTCCAAAA	CCTTCCTAGTGATCC	ACGGATGGACG	[TTGA GCGGCATGTTGAAA	429
8 atlanticsalmon1		-----	-----	-----	-----	-----	-----	0
9 winterflounder1		-----	-----	-----	-----	-----	-----	0
10 winterflounder2		-----	-----	-----	-----	-----	-----	0
						FHLPL-1f		
1 rainbowtrout1	451	465 466	480 481	495 496	510 511	525 526	540	
2 rainbowtrout2		GCTGGTCCCTAACG	TGGTGACAGCGCTGT	ACAAGAGGGAGCCCA	AGGCCAACGTC	TGA	CACGGGCGCAGCAGC	529
3 zebrafish1		GCTGGTCCCTAACG	TGGTGACAGCGCTGT	ACAAGAGGGAGCCCA	AGGCCAACGTC	TGA	CACGGGCGCAGCAGC	540
4 zebrafish2		-----	-----	-----	-----	-----	-----	0
5 catfish		-----	-----	-----	-----	-----	-----	0
6 atlanticsalmon2		-----	-----	-----	-----	-----	-----	0
7 bream		GCTGGTGGCGAACG	TGGTGTCAGCGCTGT	ACGAGAGAGAGCAGA	CGGCCAACGTC	TGA	CCTCGGCACAGAAC	519
8 atlanticsalmon1		-----	-----	-----	-----	-----	-----	0
9 winterflounder1		-----	-----	-----	-----	-----	-----	0
10 winterflounder2		-----	-----	-----	-----	-----	-----	0
1 rainbowtrout1	541	555 556	570 571	585 586	600 601	615 616	630	
2 rainbowtrout2		ACTACCTCACCTCCG	CTGCCAACACCAAGC	TGGTGGGCAAAGACG	TGGCCAAGTTGTTA	ATTGGCTGCAGAAA	CACTCGACTATCCCT	619
3 zebrafish1		ACTACCTCACCTCCG	CTGCCAACACCAAGC	TGGTGGGCAAAGACG	TGGCCAAGTTGTTA	ATTGGCTGCAGAAA	CACTCGACTATCCCT	630
4 zebrafish2		-----	-----	-----	-----	-----	-----	0
5 catfish		-----	-----	-----	-----	-----	-----	0
6 atlanticsalmon2		-----	-----	-----	-----	-----	-----	0
7 bream		ACTACGTGGTGGCTG	CTCAGAACACCAAAG	CAGTGGGACAGGAGA	TCGCTCGCTTCATCG	ACTGGATCGAG	GAAA CCACCAACATGCCTC	609
8 atlanticsalmon1		-----	-----	-----	-----	-----	-----	0
9 winterflounder1		-----	-----	-----	-----	-----	-----	0
10 winterflounder2		-----	-----	-----	-----	-----	-----	0

FHLPL-2f

FHLPL-3f

FHLPL-1r **FHLPL-2r**

	1171	1185	1186	1200	1201	1215	1216	1230	1231	1245	1246	1260
1 rainbowtrout1	ACCTCAAGACCCGCG	AGATGATGCCCTTCA	AACTTTCCATTACC	AAGTGAAGGTGCATT	TCTTCAGC-AGTGAG	AAACTGGCCTACACT						1248
2 rainbowtrout2	ACCTCAAGACCCGTG	AGATGATGCCCTTCA	AACTTTCCATTACC	AAGTGAAGGTGCATT	TCTTCAGC-AGTGAG	AAACTGGCCTACACT						1259
3 zebrafish1	-----	-----	-----	-----	-----	-----	ACGACGCAGC	GAGTCAGTGAGCGAG	AACGCCGCATA-ACT			39
4 zebrafish2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	6
5 catfish	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	0
6 atlanticsalmon2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	0
7 bream	ACACCAAGACGAGAG	CCTCAATGCCTTTCA	GAG] [TTTACCACTACC	AGCTGAAGATCCACT	TCTCCAGT-AAAGTG	AACCGCTCAGAGATG						123
8 atlanticsalmon1	ACACCAAGACCAGGG	CCTCCATGCCGTTCA	GAGTCTATCACTACC	AGGTGAAGATCCACT	TCTCCAGC-AAGGTG	AACAGGTCTGAGATG						16
9 winterflounder1	-----	-----	-----	-----	-----	-----	ACG	CTGGACAGAGCCCAG	AGTCCACGTTCCAG	AAATAACCAGAAAAC		4
10 winterflounder2	-----	-----	-----	-----	-----	-----	ACG	CTGGACAGAGCCCAG	AGTCCACGTTCCAG	AAATAACCAGAAAAC		4

	1261	1275	1276	1290	1291	1305	1306	1320	1321	1335	1336	1350	
1 rainbowtrout1	GAGCAGCCCATGAAA	ATCTCTCTGTACGGA	ACCCATGATGAGAAG	ATTGA-CAT--ACCT	TACACCA-TGCCCTT	CCTGAACACCAACAG							133
2 rainbowtrout2	GAGCAGCCCATGAAA	ATCTCTCTGTACGGA	ACCCATGATGAGAAG	ATTGA-CAT--ACCT	TACACCA-TGCCCTT	CCTGAACACCAACAG							134
3 zebrafish1	TCGTATAGCATA---	-----	CATTATACG-A	AGTTATCAGTCGACG	GT--A-CCG--GACA	TATGCCCGGGATT	C-GGCCATTACGGC						11
4 zebrafish2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	
5 catfish	-----	AG	ATTCCCTGTATGGC	TATTCTGGGGAGAAG	GAGAA-CAT--CCCC	TACATTA-TGCCTGC	TTTGAAAACCAACAC						7
6 atlanticsalmon2	-----	-----	-----	--GCACGAGGCGTAG	TTGGA-CA---ACTT	TACAAAAATAGACTT	TT-----ACAAGCGG						4
7 bream	GAGCCGTCACTCACC	GTCTCCCTGTACGGG	ACAAACGGAGAGGCT	GAAAA-CCTGGAGCT	TAAACT [GAAGGAGAA	AATTGCGACAAATAA							132
8 atlanticsalmon1	GAGCCTTCACTGACA	GTCTCCCTGATTGGA	ACCAAAGGAGAGGTT	GAAGA-CCTCAAAC	GACACTAAAGGAGAA	GATAACAACCAATAA							25
9 winterflounder1	TTCCAGACGCTGAAA	--CCTCCAGCTCA-T	CCCAACCGTGCCTG	ATCAATCCTCCTCCA	AACGAAAAAGTGCTC	G-----CATCAAGAG							13
10 winterflounder2	TTCCAGACGCTGAAA	--CCTCCAGCTCA-T	CCCAACCGTGCCTG	ATCAATCCTCCTCCA	AACGAAAAAGTGCTC	G-----CATCAAGAG							3

1 rainbowtrout1	1351	1365	1366	1380	1381	1395	1396	1410	1411	1425	1426	1440	
2 rainbowtrout2		CACTGTGTCATTCT	GCTGACCACTGACGT	GGACGTTGGGAGCT	GCTTATGGTCAAGCT	GCGCTGG	--GAGAA	AGATGCCT	-----AT	1416			
3 zebrafish1		CACTGTGTCATTCT	GCTGACCACTGACGT	GGACGTTGGGAGCT	GCTTATGGTCAAGCT	GCGCTGG	--GAGAA	AGATGCCT	-----AT	1427			
4 zebrafish2		CGGGGTGTCCTTCT	GTTGACCAACGGATGC	AGACATCGGAGAGCT	GCTGATGGTAAACT	TCTCTGG	--GAGAA	AGACACGC	-----TC	195			
5 catfish		-GGGGTGTCCTTCT	GTTGACCAACGGATGC	AGACATCGGAGAGCT	GCTGATGGTAAACT	TCTCTGG	--GAGAA	AGACACGC	-----TC	81			
6 atlanticsalmon2		CACCATCTCTTCT	ACTCACCAACGGATGT	AGACATTGGTGAAGCT	GCTGATGGTGAAGCT	GCTCTGG	--GAGAA	GGACACCA	-----TC	155			
7 bream		TAAAATA-CATTC	CCGGGATTTAAAGTGAC	AAACATTTTCTT	TCATTGG--AAAGT	TCTT	--CATTAA	TATTGA	-----A-	123			
8 atlanticsalmon1		GACCCATTCTTCT	GCTGGTAGCGGAGAA	AGACATTGGTGAACCT	CCTGATGGTGAAGTT	AAATGGGAGGAGAC	AAACGGTGGTCAAC	1414					
9 winterflounder1		GAGGTAGTGCTCTC	GTTGATCCATGTG-	-----GGTCCCCC	GC-----GGTACGGT	CAAGTGGGAGGAGTC	TACTAGCTGGTCTGC	346					
10 winterflounder2		GAGGTAGTGCTCTC	GTTGATCCATGTG-	-----GGTCCCCC	GC-----GGTACGGT	GCATTGGTTGGAGA	AG-TGCCG	-----AG	202				
										202			
1 rainbowtrout1	1441	1455	1456	1470	1471	1485	1486	1500	1501	1515	1516	1530	
2 rainbowtrout2		TTC-----A	GCTGGT-----CTGA	CATT-----GGGGCAA	CAA--AAACT	-----T	CCACATC-CGCAAA	TGCGTGTCAAGGCCG		1480			
3 zebrafish1		ATC-----A	GCTGGC-----CTTG	G-----GGGAAA	CAA--AAACT	-----T	CCACATC-CGCAAA	TGCGTGTCAAGGCCG		1491			
4 zebrafish2		ATC-----A	GCTGGC-----CTTG	G-----GGAACTC	CGA--CACCT	-----T	CCACATC-CGCAAC	TGCGCATCAAGCTG		256			
5 catfish		CTC-----A	GTTGGC-----CTTG	G-----GGAACTC	CGA--CACCT	-----T	CCACATC-CGCAAC	TGCGCATCAAGCTG		142			
6 atlanticsalmon2		-TC-----A	ACTGTA-----CTAT	G-----GGAAAAG	AAAT-CACCTGGT	-T	GATCATA-ATTGGGA	TTTGTATCAAAGTCGT		216			
7 bream		CTCCAACATGCTGAA	GATGGTTCTTCTCTG	GTGGTCCGGTGA	CTCAGCGTCAACAT	GGAGGT-CACAAA	TCCGCATTCGGCAG		186				
8 atlanticsalmon1		CTCGTTCATGCTCAA	CATGGTCTTCTCTG	GTGGTCAGGTGA	TGCGGAGTCTGA	GGAGGT-CACAAA	TACGCATCCGAGTCG	435					
9 winterflounder1		TTGTT-----A	CATGGGCAGC-CTTT	T---TCTAGTTCA	CTTCCCAGCC-ATCA	CGACATTTCACAAAC	GTTATGCCCTCTCCT	278					
10 winterflounder2		TTGTT-----A	CATGGGCAGC-CTTT	T---TCTAGTTCA	CTTCCCAGCC-ATCA	CGACATTTCACAAAC	GTTATGCCCTCTCCT	278					
										278			
1 rainbowtrout1	1531	1545	1546	1560	1561	1575	1576	1590	1591	1605	1606	1620	
2 rainbowtrout2		---GGGAGACT-CAA	T-CCAGGGTGT-CT	TCAGCGCTAAAGA	--TGG--AGAGTA	T-----	GCCTA CCTCAT--CAGAGGA			1545			
3 zebrafish1		---GGGAGACT-CAA	T-CCAGGGTGT-CT	TCAGCGCTAAAGA	--TGG--AGAGTA	T-----	GCCTA CCTCAT--CAGAGGA		1556				
4 zebrafish2		---GAGAGACA-CAG	T-CAAAATCAT-CT	TTAGTGCAAAGA	--AAG--TGAATT	T-----	TCCTA CTTTC--CCGTGGA		321				
5 catfish		---GAGAGACA-CAG	T-CAAAATCAT-CT	TTAGTGCAAAGA	--AAG--TGAATT	T-----	TCCTA CTTTC--CCGTGGA		207				
6 atlanticsalmon2		CATGTGAGTCT-CTC	TTCAACACTGAAACT	CCTGCTCTCAGCA	--AGG--TGAGTT	T-----	GCCTA CCTCAC--CCGTGGA		281				
7 bream		---GCGAGACC-CAA	C-AAAA] [GATGGT-GT	TCTGTCAAAGACC	CTGAAAGTCAGAAGT	TAACACAAGAGGTCA	CCTTTGTTAAATGTA		1587				
8 atlanticsalmon1		---GGGAGACG-CAG	A-AAAAGATGGA-GT	TCTGCATTAAAGATC	CTCATGCTCTGAGCT	TACACAGGAAGTTA	CGTTGTTAAATGCA	519					
9 winterflounder1		--GGTGAGATTTCAG	TTTTGAAGGCT-CT	TGAGC--ACAGA	--AAG--TAAATT	TCTTCTGCTTCTC	CCGCTT-CACACGA	353					
10 winterflounder2		--GGTGAGATTTCAG	TTTTGAAGGCT-CT	TGAGC--ACAGA	--AAG--TAAATT	TCTTCTGCTTCTC	CCGCTT-CACACGA	353					
										353			
1 rainbowtrout1	1621	1635	1636	1650	1651	1665	1666	1680	1681	1695	1696	1710	
2 rainbowtrout2		AAA-----GACGA	CGTAGTCTTGTCAA	ATC--AAAAGAGGA	--CAACATGAGC	---	CGGAAAGAGAAA	AC---GATGCACAGA		1614			
3 zebrafish1		AAA-----GACGA	CGTAGTCTTGTCAA	ATC--AAAAGAGGA	--CAACATGAGC	---	CGGAAAGAGAAA	AC---GATGCACAGA	1625				
4 zebrafish2		GGT-----GAGGC	GGCGTCTTCGTGAA	AGA--CAAAGAGGC	--CCAGTCGAGC	---	CGCAAAAACCA	AG---ATTGCACAA	390				
5 catfish		GGT-----GAGGC	GGCGTCTTCGTGAA	AGA--CAAAGAGGC	--CCAGTCGAGC	---	CGCAAAAACCA	A-----	265				
6 atlanticsalmon2		ATA-----GAGAC	CAAGTTCTCAGTGC	TTC--AGTTGAGGAG	--CCAGAGGAGG	---	ATCTGTGTTACC	TG---GTGCCAGGCA	348				
7 bream		AGGATGCATGGAGGA	CGAACCAAACAAA	CTCTAAAGAG	[AGTAA	CTCTGGAGAACACT	GACCTCGAAGAAAAG	ACCCACTCCACC	1677				
8 atlanticsalmon1		A-----CGA	CGAA-TGGAGGAA	AAA	CTTCAAAAGAGTGA	A-----			553				
9 winterflounder1		AT-----GCGCC	GGACTTCCAGGAC	GTT---TGGGGGTT	TGGCCTCCAGGCC	---CAGAAGGACTTC	TTCACGCTTCCAC	428					
10 winterflounder2		AT-----GCGCC	GGACTTCCAGGAC	GTT---TGGGGGTT	TGGCCTCCAGGCC	---CAGAAGGACTTC	TTCACGCTTCCAC	428					
										518			
1 rainbowtrout1	1711	1725	1726	1740	1741	1755	1756	1770	1771	1785	1786	1800	
2 rainbowtrout2		CTCAAGATGCAG	--GGAAG	--CCACTT-C	AAGAATAACATTGCA	TGAAGCAACACTTGA	GTAGAGTGCACACTG	--AACCACACAGCCA		1696			
3 zebrafish1		CTCAAGATGCAG	--GGAG	--CCACTT-C	AAGAATAACATTGCA	TGAAGCAACACTTGA	GTAGAGTGCACACTG	--AACCACACAGCCA	1707				
4 zebrafish2		ATGAAGATGCAC	--GGCAG	--TTCATT-C	AAACAGAACACCGAG	TAAAGACCCACGTCA		--AAATCCACCTCC-	456				
5 catfish		-TTGACTTATT	--CCAAC	--TTACTGAC	ACACAGTAGCATCAC	AAAAGAGTC			265				
6 atlanticsalmon2		AAAAGGATACA	--GTCAC	--ACAGTG	--TGGCTCAAATTCAG	CTTACCTACGTTGCG	--CATCATACTACATGG	--CTGGTCGGTGGCA	396				
7 bream		-TGAACATCCCTCG	TCCAT	--TCAAGCAC	TTAGTAATAATTAA	TACACCTGTAACCTA	TTTACTCTC	--ACTCTCACCGCCT	407				
8 atlanticsalmon1									1761				
9 winterflounder1		ATGGAGCTCCAGGAC	TCATCGTTCTTCC	CAGCTCAGACTGATC	TTCAGCAGGTCCCCA	ATCTCCTCTCGGT	AAAACAGGAAGGTG	553					
10 winterflounder2		ATGGAGCTCCAGGAC	TCATCGTTCTTCC	CAGCTCAGACTGATC	TTCAGCAGGTCCCCA	ATCTCCTCTCGGT	AAAACAGGAAGGTG	518					
									518				
1 rainbowtrout1	1801	1815	1816	1830	1831	1845	1846	1860	1861	1875	1876	1890	
2 rainbowtrout2		AAGCT-TGATGGA-G	CCTGGGAGAATACAC	TGG--AGAGAA	AGGGGATGTTAGT-C	TCTGACCTGTGCTG	GTAGAGTGCACACTG	--AACCACACAGCCA		1773			
3 zebrafish1		AAGCT-TGATGGA-G	CCTGGGAGAATACAC	TGG--AGAGGA	AGGGGATGTTAGT-C	TCTGACCTGTGCTG	GTAGAGTGCACACTG	--AACCACACAGCCA	1784				
4 zebrafish2		-----GCTGGG-A	TCTGGCACAG-AATC	TGC-----CAGG	-----GAGGAA	CTGGAATGGTAGA-A	TCTGACCTTCCCCGT	ATATA-----TTCCG	528				
5 catfish		-----							265				
6 atlanticsalmon2		GGTCT-GTTGAG-A	GCTGGATCCACAA	TGG--T-AGCGGCC	CTG---TTTGAGC-G	TGTGCCAATGCCAA	CGTTA-----TAGTG		396				
7 bream		TGTAT-ATATCCACA	AAAGTATGTTGTA	TTT-----AAAAAAA	CATCTATGTGAATAC	TGTAAAATATTTAT	AAAGAGCTGTTTTA		482				
8 atlanticsalmon1									1847				
9 winterflounder1		TNGGTCAGGTCAGA	CCGATGACGTTTCA	TGGCTGAACACAGAT	AAANCAGGTAGG	--TCAGATGTC	ANGTGGATTATNGTA	553					
10 winterflounder2		TNGGTCAGGTCAGA	CCGATGACGTTTCA	TGGCTGAACACAGAT	AANC-AGGTAGG	--TCAGATGTC		606					
								589					
1 rainbowtrout1	1891	1905	1906	1920	1921	1935	1936	1950	1				

1 rainbowtrout1	2071	2085	2086	2100	2101	2115	2116	2130	2131	2145	2146	2160	
2 rainbowtrout2	CCTTCATCTCACT	GTGTTT	--ATGTATA	CCTACTAAATCAAAT	AAATTAAATTCCCTAAC	GCCATA	AAAAAAAAAAAAA	AAAAAAA	-----	-----	-----	-----	2022
3 zebrafish1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	655
4 zebrafish2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	265
5 catfish	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	396
6 atlanticsalmon2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	602
7 bream	GGAATAATGTCCTCT	TTGTGTCCATGGAGG	TCTGCAAAGGGAAC	ATGAGAAATTCTTCT	TCATAGTAAGATGAA	GGTGTTTATGCAAT	-----	-----	-----	-----	-----	-----	2115
8 atlanticsalmon1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	553
9 winterflounder1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	615
10 winterflounder2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	589
	2161	2175	2176	2190	2191	2205	2206	2220	2221	2235	2236	2250	
1 rainbowtrout1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	2022
2 rainbowtrout2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	1982
3 zebrafish1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	655
4 zebrafish2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	265
5 catfish	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	396
6 atlanticsalmon2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	602
7 bream	AGGGCCCACATTATT	TACCTTTAIGTGT	TCTAACCTGACAGCA	TTCACCACTGTGGC	CTCAGTACTGTATGT	GAETGTGGGACCTGT	-----	-----	-----	-----	-----	-----	2205
8 atlanticsalmon1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	553
9 winterflounder1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	615
10 winterflounder2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	589
	2251	2265	2266	2280	2281	2295	2296	2310	2311	2325	2326	2340	
1 rainbowtrout1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	2022
2 rainbowtrout2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	1982
3 zebrafish1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	655
4 zebrafish2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	265
5 catfish	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	396
6 atlanticsalmon2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	602
7 bream	ATCAGTGTGCCTTAT	ACTGTAAACAGTCAC	CAGAGTAAGCTGTT	GCTGTTTGTCTAAC	TTAGCGCTTGTGTC	TGTTTTCACCTCGG	-----	-----	-----	-----	-----	-----	2295
8 atlanticsalmon1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	553
9 winterflounder1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	615
10 winterflounder2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	589
	2341	2355	2356	2370	2371	2385	2386	2400	2401	2415	2416	2430	
1 rainbowtrout1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	2022
2 rainbowtrout2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	1982
3 zebrafish1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	655
4 zebrafish2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	265
5 catfish	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	396
6 atlanticsalmon2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	602
7 bream	CATGTTTATTGTTG	TTATTCACTTAAC	TGACTTTACTGA	AAACGTCACCCATGA	AAAAGATGTTTAAT	CTCCTGTGCAAGATG	-----	-----	-----	-----	-----	-----	2385
8 atlanticsalmon1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	553
9 winterflounder1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	615
10 winterflounder2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	589
	2431	2445	2446	2460	2461	2475	2476	2490	2491	2505	2506	2520	
1 rainbowtrout1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	2022
2 rainbowtrout2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	1982
3 zebrafish1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	655
4 zebrafish2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	265
5 catfish	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	396
6 atlanticsalmon2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	602
7 bream	TTACTTTGAACTGT	TACGGTATAGAAGAC	ATTTGTCTACCATG	GTTTATTAAATTGG	TGAAATTAGATGAGA	GGCGTGTGCTCAGT	-----	-----	-----	-----	-----	-----	2475
8 atlanticsalmon1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	553
9 winterflounder1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	615
10 winterflounder2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	589
	2521	2535	2536	2550	2551	2565	2566	2580	2581	2595	2596	2610	
1 rainbowtrout1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	2022
2 rainbowtrout2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	1982
3 zebrafish1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	655
4 zebrafish2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	265
5 catfish	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	396
6 atlanticsalmon2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	602
7 bream	GCGAGGTGGAAAGCA	CAGTCATATCAGTG	TTGTCATTTATTGT	ACTGCTGTGCGTAG	TCACCTTGTACTTCT	GTGCATTTAGCTTT	-----	-----	-----	-----	-----	-----	2565
8 atlanticsalmon1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	553
9 winterflounder1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	615
10 winterflounder2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	589
	2611	2625	2626	2640	2641	2655	2656	2670	2671	2685	2686	2700	
1 rainbowtrout1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	2022
2 rainbowtrout2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	1982
3 zebrafish1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	655
4 zebrafish2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	265
5 catfish	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	396
6 atlanticsalmon2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	602
7 bream	TAAAATCTTTATTTC	ATAATTACATTCCA	GTGCTGTCAAGCAATG	TCTCAACCTCGGCTC	TTTAGAGATCGGTAC	TGTAATCATTGTGAC	-----	-----	-----	-----	-----	-----	2655
8 atlanticsalmon1	-----	-----	-----	-----</									

1 rainbowtrout1	2791	2805	2806	2820	2821	2835	2836	2850	2851	2865	2866	2880	2022
2 rainbowtrout2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	1982
3 zebrafish1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	655
4 zebrafish2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	265
5 catfish	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	396
6 atlanticsalmon2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	602
7 bream	TTATGAAATACAAC	AATTACAGATGATAG	ATTTTGAGTATTCA	GCAGTGTAGTTGTA	CATTCTACCTTGG	AATTAATCTGTAAA	-----	-----	-----	-----	-----	-----	2835
8 atlanticsalmon1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	553
9 winterflounder1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	615
10 winterflounder2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	589
1 rainbowtrout1	2881	2895	2896	2910	2911	2925	2926	2940	2941	2955	2956	2970	2022
2 rainbowtrout2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	1982
3 zebrafish1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	655
4 zebrafish2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	265
5 catfish	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	396
6 atlanticsalmon2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	602
7 bream	ATCTATAATCTTTG	TACATATTTGAAAT	ATTCGCTGTACTGTA	TACATTCTTAATAA	ACCAAAAGTGCCATC	CAAAAAAAAAAAAAA	-----	-----	-----	-----	-----	-----	2925
8 atlanticsalmon1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	553
9 winterflounder1	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	615
10 winterflounder2	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	-----	589
1 rainbowtrout1	2971	2985	2986	3000									
2 rainbowtrout2	-----	-----	-----	2022									
3 zebrafish1	-----	-----	-----	1982									
4 zebrafish2	-----	-----	-----	655									
5 catfish	-----	-----	-----	265									
6 atlanticsalmon2	-----	-----	-----	396									
7 bream	AAAAAAAAAAAAAA	AAAAAAA	2946										
8 atlanticsalmon1	-----	-----	-----	553									
9 winterflounder1	-----	-----	-----	615									
10 winterflounder2	-----	-----	-----	589									

Where: reverse primers; forward primers; over lap between primers.

The LPL Primers FHLPL4r and 7f were non-degenerate primers based on the primers FHLPLr1 and FHLPLf3 respectively, which can be seen on the above alignment.

Accession numbers for sequences for different species used: Rainbow trout 1: 4584059, Rainbow trout 2: 14582900, zebrafish 1: CB352163.1, zebrafish 2: CB352334.1, Catfish: 18460110, Atlantic salmon 1: 15280954, Atlantic salmon 2: 40541657, Bream: AB054062, Winter flounder 1: CF195532, Winter flounder 2: CF195573.

Appendix 3: β-actin sequence alignment using the ClustalW Multiple sequence alignment on the BCM search launcher.

The genomic sequence of Chinese minnow was included because a) it is closely related to fathead minnow and b) to identify intron/exon boundaries.

	1	15 16	30 31	45 46	60 61	75 76	90		
1 yellow_perch	-	-	-	-	-	-	-	0	
2 gilthead_bream	-	-	-	-	-	-	-	0	
3 grass_carp	-	-	-	-	-	-	-	0	
4 Chinese_minnow	-	-	-	-	-	-	-	0	
5 Jap_Silver_carp	-	-	-	-	-	-	-	0	
6 zebrafish2	-	-	-	-	-	-	-	50	
7 zebrafish1	GGGAAGCAGTGGTAT	CAACGCAGCGTGCCC	ATTACGGCTGGGAT	TGTGAGTTTCAGTG	CACGCTGAGAAGATC	TTCACT-CCCTTGT	90	90	
	91	105 106	120 121	135 136	150 151	165 166	180		
1 yellow_perch	-	-	-	AT GGATGACCAAATCGC	CGCCCTCGTGTGA	CAACGGATCCGGTAT	GTGCAAAGCTGGCTT	62	
2 gilthead_bream	-	-	-	AT GGATGATGAAATTGC	CGCACTGGTTGTGA	CAACGGATCCGGTAT	GTGCAAAGCCGGATT	62	
3 grass_carp	-	-	-	AT GGATGATGAAATTGC	CGCACTGGTTGTGA	CAACGGATCCGGTAT	GTGCAAAGCCGGATT	62	
4 Chinese_minnow	-	-	-	TCACAATAACCTACT	-AATACACAGCC-AT	GGATGAGGAAATCGC	TGCCCTGGTCGTGTGA	CAACGGCTCTGGTAT	GTGCAAAGCCGGTTT
5 Jap_Silver_carp	-	-	-	TCACAATAACCTACT	TAATACCCAGCCCCAT	GGATGAGGAAATCGC	TGCCCTGGTCGTGTGA	TAACGGCTCCGGTAT	GTGCAAAGCCGGTTT
6 zebrafish2	-	-	-	-	-	-	-	138	
7 zebrafish1	-	-	-	-	-	-	-	180	
	181	195 196	210 211	225 226	240 241	255 256	270		
1 yellow_perch	TGCAGGAGATGATGC	TCCCGTGTGTGTT	CCCCCTCATTGTTGG	ACGTCCAAGACATCA	GG-----	-	-	124	
2 gilthead_bream	-	-	-	-	-	-	-	0	
3 grass_carp	-	-	-	CGCTGGAGATGATGC	TCCCCGTGTGTCTT	CCCATCCATCGTGGG	TCGCCCCAGACATCA	GG-----	124
4 Chinese_minnow	-	-	-	CGCTGGAGATGATGC	TCCCCGAGCTGTCTT	CCCATCCATCGTGGG	TCGCCCCAGACATCA	GGTGGAGAAGCGGATG	ATAAATCGATTTAGG
5 Jap_Silver_carp	-	-	-	-	-	-	-	152	
6 zebrafish2	-	-	-	-	-	-	-	46	
7 zebrafish1	-	-	-	CC TCCCCGTGTGTCTT	CC-ATCC-TCGTGGG	TCGCCCCAGAC-TCA	GG-----	-	200
	-	-	-	TGCTGGAGATGATGC	CCCTCGTGTGTGTTT	CCCCCTCATTGTTGG	ACGACCCAGACATCA	GG-----	242
	-	-	-	TGCTGGAGATGATGC	CCCTCGTGTGTGTTT	CCCCCTCATTGTTGG	ACGACCCAGACATCA	GG-----	242
	271	285 286	300 301	315 316	330 331	345 346	360		
1 yellow_perch	-	-	-	-	-	-	-	124	
2 gilthead_bream	-	-	-	-	-	-	-	0	
3 grass_carp	-	-	-	GTTATCTGTAATGAG	AATTATTTTCGTACT	TAAGAGTGGAGTTCAT	TTCTAGTTCTAAAC	ATTTTACAAAAAATTA	ACATTGCTTTCTTTG
4 Chinese_minnow	-	-	-	-	-	-	-	124	
5 Jap_Silver_carp	-	-	-	-	-	-	-	242	
6 zebrafish2	-	-	-	-	-	-	-	46	
7 zebrafish1	-	-	-	-	-	-	-	200	
	-	-	-	-	-	-	-	242	
	361	375 376	390 391	405 406	420 421	435 436	450		
1 yellow_perch	-	-	-	GTGTGATG	GTTGGCATGGGCCAG	AAAGATAGCTATGTT	GGTGTAGAGGCACAG	AGCAAGGGGTATC	CTGACGCTGAAGTAC
2 gilthead_bream	-	-	-	-	-	-	-	207	
3 grass_carp	-	-	-	GTGTCATG	GTTGGTATGGGACAG	AAGGACAGCTACGTT	GGTGATGAGGCTCAG	AGCAAGAGGGTATC	CTGACGCTGAAGTAC
4 Chinese_minnow	-	-	-	TTACAGCGTGTGTCATG	GTCGGTATGGGACAG	AAGGACAGCTATGTT	GGTGACGAGGCTCAG	AGCAAGAGGGTATC	CTGACGCTGAAGTAC
5 Jap_Silver_carp	-	-	-	GTGTCATG	GTTGGTATGGGACAG	AAGGACAGCTACGTT	GGTGATGAGGCTCAG	AGCAGAGAGGTATC	CTGACGCTGAAGTAC
6 zebrafish2	-	-	-	GAGTGATG	GTTGGCATGGGACAG	AAAGACTCTATGTTG	GGAGATGAGGCCAG	AGCAGAGAGGTATC	CTGACGCTCAAATAC
7 zebrafish1	-	-	-	GAGTGATG	GTTGGCATGGGACAG	AAAGACTCTATGTTG	GGAGATGAGGCTCAG	AGCAAGAGGGTATC	CTGACGCTCAAATAC
	451	465 466	480 481	495 496	510 511	525 526	540		
1 yellow_perch	CCCATTGAGCATGGT	ATCGTCACCAACTGG	GACAGAAGGAGAAG	ATCTGGCATACACC	TTCTACAAACGAGCTG	AGAGTTGCGCCCGAG	297		
2 gilthead_bream	-	-	-	TCACCAACTGG	GATACAGAAGGAGA	ATCTGGCATACACC	TTCTACAAACGAGCTG	AGAGTTGCCCTGAG	71
3 grass_carp	-	-	-	CCCATTGAGCACGGT	ATTGTCACCAACTGG	GACGATATGGAGAAG	ATCTGGCATACACC	TTCTACAAACGAGCTG	CGTGTGCCCCAGAG
4 Chinese_minnow	-	-	-	CCCATTGAGCACGGT	ATTGTCACCAACTGG	GACGATATGGAGAAG	ATCTGGCATACACC	TTCTACAAATGAGCTG	422
5 Jap_Silver_carp	-	-	-	CCCATTGAGCACGGT	ATTGTCACCAACTGG	GATGATATGGAGAAG	ATCTGGCATACACC	TTCTACAAACGAGCTG	CGTGTGCCCCAGAG
6 zebrafish2	-	-	-	CCCATTGAGCACGGT	ATTGTAACTAACGTTG	GATGATATGGAGAAG	ATCTGGCATACACC	TTCTACAAATGAGCTC	373
7 zebrafish1	-	-	-	CCCATTGAGCACGGT	ATTGTGACCAACTGG	GATGATATGGAGAAG	ATCTGGCATACACC	TTCTACAAATGAGCTC	CGTGTGCCCCCTGAG
	541	555 556	570 571	585 586	600 601	615 616	630		
1 yellow_perch	GAGCACCCCGTCCTG	CTCACAGAGGC-TCC	CCTGAACCCCAAGGC	CAACAGGGAAAAGAT	GAC-----	-	-	359	
2 gilthead_bream	-	-	CCTGAACCCCTGTCCTG	CTCACAGAGGC-CCC	CCTGAACCCCAAGGC	CAACAGGGAAAAGAT	GAC-----	133	
3 grass_carp	-	-	GAGCACCCCGTCCTG	CTCACAGAGGC-CCC	CCTGAACCCCAAGGC	CAACAGGGAAAAGAT	GAC-----	359	
4 Chinese_minnow	-	-	GAGCACCCCGTCCTG	CTCACAGAGGC-CCC	CCTGAACCCCAAGGC	CAACAGGGAAAAGAT	GACACAGGTGGTTT	TTGGCTAGAAAATGG	
5 Jap_Silver_carp	-	-	GAGCACCCCGTCCTG	CTCACAGAGGC-CCC	CCTGAACCCCAAGGC	CAACAGGGAAAAGAT	GAC-----	511	
6 zebrafish2	-	-	GAGCACCCCTGTCCTG	CTCACTGAGGCTCCC	CCTGAATCCCAAAAGC	CAACAGAGAGAAGAT	GACACAGATCATGTT	CGAGACCTTCAAC--	
7 zebrafish1	-	-	GAGCACCCCTGTCCTG	-	CCTGAATCCCAAAAGC	CAACAGAGAGAAGAT	GACACAGATCATGTT	461	
	-	-	-	-	-	-	-	429	
	631	645 646	660 661	675 676	690 691	705 706	720		
1 yellow_perch	-	-	-	-	-	-	-	359	
2 gilthead_bream	-	-	-	-	-	-	-	133	
3 grass_carp	-	-	-	-	-	-	-	359	
4 Chinese_minnow	-	-	TGCATTGAGTCCTCT	TGTCTGTCTTCGTC	TTTAAGTTCTCCTT	TTCATTGTTCACTT	CCTCCAGGCTTGTG	TCCTCTGAGCTCTG	
5 Jap_Silver_carp	-	-	-	-	-	-	-	601	
6 zebrafish2	-	-	-	-	-	-	-	281	
7 zebrafish1	-	-	-	-	-	-	-	461	
	-	-	-	-	-	-	-	429	
	721	735 736	750 751	765 766	780 781	795 796	810		
1 yellow_perch	-	-	-	-	-	-	-	359	
2 gilthead_bream	-	-	-	-	-	-	-	133	
3 grass_carp	-	-	-	-	-	-	-	359	
4 Chinese_minnow	-	-	AGTTTCTCATCTTT	GCTGAAAGCAGCAGG	TTATCTATACTTTTG	CCTGCCCTGTTTGCA	GTCTCCCTGACTCTT	ATTCTTGTGCACTT	
5 Jap_Silver_carp	-	-	-	-	-	-	-	691	
6 zebrafish2	-	-	-	-	-	-	-	281	
7 zebrafish1	-	-	-	-	-	-	-	461	
	-	-	-	-	-	-	-	429	

	811	825	826	840	841	855	856	870	871	885	886	900	
1 yellow_perch	-----												359
2 gilthead_bream	-----												133
3 grass_carp	-----												359
4 Chinese_minnow	TTGTTCTTACTCT	GGATTCACAACTAAC	CCCTGCATGGATGT	TGGATTGCTGCTGT	AAATATTGAGCATC	AGTTAACTTCCTCAC							781
5 Jap_Silver_carp	-----												281
6 zebrafish2	-----												461
7 zebrafish1	-----												429
	901	915	916	930	931	945	946	960	961	975	976	990	
1 yellow_perch	-----	CCAGAT	CATGTTGAGACCTT	CAACACCCCTGCCAT	GTACGTTGCCATCCA	GGCCGTGCTGTCCT	GTATGCCTCTGGTCG						440
2 gilthead_bream	-----	CCAGAT	CATGTTGAGACCTT	CAACACCCCGCCAT	GTATGTTGCCATCCA	GGCTGTGCTGTCCT	GTATGCCTCTGGTCG						214
3 grass_carp	-----	ACAGAT	CATGTTGAGACCTT	CAACACCCCGCCAT	GTACGTTGCCATCCA	GGCTGTGCTGTCCT	GTATGCCTCTGGTCG						440
4 Chinese_minnow	TCTCTTTCCAGAT	CATGTTGAGACCTT	CAACACCCAGCCAT	GTACGTTGCCATCCA	GGCTGTGCTGTCCT	GTATGCCTCTGGTCG							871
5 Jap_Silver_carp	-----	ACAGAT	CATGTTGAGACCTT	CAACACCCCTGCCAT	GTACGTTGCCATCCA	AGCTGTGCTGTCCT	GTATGCCTCTGGTCG						362
6 zebrafish2	-----												461
7 zebrafish1	-----												429
	991	1005	1006	1020	1021	1035	1036	1050	1051	1065	1066	1080	
1 yellow_perch	TACCACTGGTATCGT	CATGGACTCCGGTGA	TGGTGTGACCCACAC	AGTCCCCATCTATGA	GGGCTATGCCCTGCC	CCACGCCATCCTGCC							530
2 gilthead_bream	TACCACTGGTATTG	CATGGATTCCGGTGA	TGGTGTGACCCACAC	AGTCCCCATCTATGA	GGGCTATGCCCTGCC	CCACGCCATCCTGCC							304
3 grass_carp	TACCACTGGTATCGT	GATGGACTCTGGTGA	TGGTGTACCCACAC	TGTGCCCATCTACGA	GGGTTACGCCCTGCC	CCACGCCATCCTCCG							530
4 Chinese_minnow	TACCACTGGTATCGT	GATGGACTCTGGTGA	TGGTGTACCCACAC	CGTGCCCATCTACGA	GGGGTACGCCCTGCC	CCATGCCATCCTCCG							961
5 Jap_Silver_carp	TACCACTGGTATTG	GATGGACTCTGGTGA	TGGTGTACCCACAC	TGTGCCCATCTACGA	GGGTTACGCCCTGCC	CCATGCCATCCTCCG							452
6 zebrafish2	-----												461
7 zebrafish1	-----												429
	1081	1095	1096	1110	1111	1125	1126	1140	1141	1155	1156	1170	
1 yellow_perch	TCTGGACTTGGCTGG	CCGTGACCTCACAGA	CTACCTCATGAAGAT	CCTGACAGAGCGTGG	GTACTCATTACACCAC	CACAGCTGAGAGGGAA							620
2 gilthead_bream	TCTGGACTTGGCCGG	CCGCGACCTCACAGA	CTACCTCATGAAGAT	CCTGACAGAGCGTGG	CTACTCCTTCACCAC	CACAGCGAGAGGGAA							394
3 grass_carp	TCTGGACTTGGCTGG	CCGTGACCTGACTGA	CTACCTCATGAAGAT	CCTGACCGAGAGAGG	CTACAGCTTCACCAC	CACAGCGAGAGGGAA							620
4 Chinese_minnow	TCTGGACTTGGCTGG	CCGTGACCTGACCGA	CTACCTCATGAAGAT	CCTGACCGAGAGAGG	TTACAGCTTCACCAC	CACAGCTGAGAGGGAA							1051
5 Jap_Silver_carp	TCTGGACTTGGCTGG	CCGTGACCTGACAGA	CTACCTCATGAAGAT	CCTGACCGAGAGAGG	CTACAGCTTCACCAC	CACAGCGAGAGGGAA							542
6 zebrafish2	-----												461
7 zebrafish1	-----												429
	1171	1185	1186	1200	1201	1215	1216	1230	1231	1245	1246	1260	
1 yellow_perch	AATCGTGCCTGACAT	CAAGGAGAAGCTGTG	CTATGTCGCCCTGGA	CTTCGAGCAGGAGAT	GGGCACTGCTGCC	CTCTTCCTCCCTGG							710
2 gilthead_bream	AATCGTGCCTGACAT	CAAGGAGAAGCTGTG	CTACGTCGCCCTGGA	CTTCGAGCAGGAGAT	GGGTACCGCTGCC	CTCCTCCCTCCCTGG							484
3 grass_carp	AATTGTCCTGACAT	CAAGGAGAAGCTCTG	CTATGTCGCCCTGGA	CTTCGAGCAGGAGAT	GGGCACTGCTGCC	CTCCTCCCTCCCTGG							710
4 Chinese_minnow	AATTGTCCTGACAT	CAAGGAGAAGCTCTG	CTATGTCGCCCTGGA	CTTCGAGCAGGAGAT	GGGCACCCTGCTTC	CTCCTCCCTCCCTGG							1141
5 Jap_Silver_carp	AATTGTCCTGACAT	CAAGGAGAAGCTCTG	CTATGTCGCCCTGGA	CTTCGAGCAGGAGAT	GGGCACCCTGCTTC	CTCCTCCCTCCCTGG							632
6 zebrafish2	-----												461
7 zebrafish1	-----												429
	1261	1275	1276	1290	1291	1305	1306	1320	1321	1335	1336	1350	
1 yellow_perch	GAAGAGCTACGAGCT	GGCCGACGGACAGGT	CATACCATCGGCAA	TGAGAGGTTCCGTTG	CCCAGAGGCCCTCTT	CCAGCCTTCCTTCCT							800
2 gilthead_bream	GAAGAGCTATGAGCT	GGCCGACGGACAGGT	CATACCATCGGCAA	TGAGAGGTTCCGTTG	CCCAGAGGCCCTCTT	CCAGCCATCCTTCCT							574
3 grass_carp	GAAGAGCTACGAGCT	GGCTGACGGACAGGT	CATACCATCGGCAA	TGAGAGGTTCCGTTG	CCCAGAGGCCCTGTT	CCAGCCATCCTTCCT							800
4 Chinese_minnow	GAAGAGCTATGAGCT	GGCTGACGGACAGGT	CATACCATCGGCAA	TGAGAGGTTCCGTTG	CCCAGAGGCCCTGTT	CCAGCCATCCTTCCT							1231
5 Jap_Silver_carp	AAAGAGCTATGAGCT	G-----											648
6 zebrafish2	-----												461
7 zebrafish1	-----												429
	1351	1365	1366	1380	1381	1395	1396	1410	1411	1425	1426	1440	
1 yellow_perch	CGGTA-----												805
2 gilthead_bream	CGGTA-----												579
3 grass_carp	GGGTA-----												805
4 Chinese_minnow	GGGTAGGTTCCCTTA	CAAACATCCCTGGT	GTGTGTATGTACACT	AGATTAATTTGAG	GGGAGTACAAGAAC	CTGGCTAATCATTT							1321
5 Jap_Silver_carp	-----												648
6 zebrafish2	-----												461
7 zebrafish1	-----												429
	1441	1455	1456	1470	1471	1485	1486	1500	1501	1515	1516	1530	
1 yellow_perch	-----	TGGAGTCCTGCG	GAATCCATGAGACCA	CCTACAAACAGCATT	TGAAGTGTATGTCTG	ACATCGTAAGGACC							877
2 gilthead_bream	-----	TGGAGTCCTGCG	GAATCCATGAGACCA	CCTACAAACAGCATT	TGAAGTGTACGTC	ACATCGTAAGGACC							651
3 grass_carp	-----	TGGAGTCCTGCG	GTATCCATGAGACCA	CCTTCAACTCCATCA	TGAAGTGTACGTC	ACATCGTAAGGACC							877
4 Chinese_minnow	TGTCTGCTCTGCAG	GTATGGATTCTGCG	GTATCCATGAGACCA	CCTTCAACTCCATCA	TGAAGTGTACGTC	ACATCGTAAGGACC							1411
5 Jap_Silver_carp	-----												648
6 zebrafish2	-----												461
7 zebrafish1	-----												429
	1531	1545	1546	1560	1561	1575	1576	1590	1591	1605	1606	1620	
1 yellow_perch	TGTACGCCAACACCG	TGCTGTCTGGAGGTA	CCACCATGTACCCCG	GCATTGCCGACAGGA	TGCAGAAGGAGATCA	CAGCCCTGGCCCCAT							967
2 gilthead_bream	TGTATGCCAACACTG	TGCTGTCTGGAGGTA	CCACCATGTACCCCG	GCATGCCGACAGGA	TGCAGAAGGAAATCA	CAGCCCTGGCCCCAT							741
3 grass_carp	TGTATGCCAACACTG	TATTGTCTGGTGGTA	CCACCATGTACCCCG	GCATTGCTGACAGGA	TGCAGAAGGAGATCA	CATCCCTGGCCCCCA							967
4 Chinese_minnow	TGTATGCCAACACTG	TATTGTCTGGTGGTA	CCACCATGTACCCCG	GCATTGCTGACAGGA	TGCAGAAGGAGATCA	CATCCCTGGCCCCCA							1501
5 Jap_Silver_carp	-----												648
6 zebrafish2	-----												461
7 zebrafish1	-----												429
	1621	1635	1636	1650	1651	1665	1666	1680	1681	1695	1696	1710	
1 yellow_perch	CCACCATGAAAATCA	AG-----											

	1801	1815	1816	1830	1831	1845	1846	1860	1861	1875	1876	1890
1 yellow_perch	CACCTTCCAGCAGAT			GTTGGATCAGCAAGCA	GGAGGTACGATGAGTC	CGGCCCCCTCCATCGT	CCACCGTAAATGCTT	CTAA		1128		
2 gilthead_bream	CACCTTCCAGCAGAT			GTTGGATT-----						846		
3 grass_carp	CACCTTCCAGCAGAT			GTTGGATTAGCAAGCA	GGAGGTACGATGAGTC	TGGACCATCCATCGT	CCACCGCGAATGCTT	CTAA		1128		
4 Chinese_minnow	CACCTTCCAGCAGAT			GTTGGATTAGCAAGCA	GGAGGTATGATGAGTC	TGGACCATCCATTGC	CCACCGCAAATGCTT	CTAA		1760		
5 Jap_Silver_carp	-----									648		
6 zebrafish2	-----									461		
7 zebrafish1	-----									429		

Where: **reverse primers**; **forward primers**; **over lap between primers**, and parts of the Chinese minnow sequence identified as introns are shown in light grey.

Accession numbers for different sequences used: Yellow Perch: AY332493; Gilthead bream: F384096; Grass Carp: AF393832; Chinese minnow: AF200957; Sliver Carp: AB020852; zebrafish 1: CB366679; zebrafish 2: CB366658.